EVOLUTIONARY AND ECOLOGICAL GENOMICS IN DEEP-SEA ORGANISMS

By

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ABSTRACT

Hydrothermal vents and coral ecosystems are conspicuous biological hot spots in the deep-sea. These ecosystems face increasing threats from human activities. Having thorough taxonomic inventories as well as understanding species' relatedness, genetic diversity, connectivity patterns, and adaptive potential is fundamental for the implementation of conservation strategies that help mitigate these threats. This thesis provides fundamental high-priority knowledge in taxonomic, evolutionary, and ecological aspects of deep-sea coral and vent species, by harnessing the power of genomic tools and overcoming long-standing methodological barriers. First, I develop bioinformatic tools that help guide the design of studies aiming to characterize eukaryotic genome diversity using restriction-site associated DNA sequencing. Using these tools I find that the predictability of restriction site frequencies in eukaryotic genomes is chiefly determined by the phylogenetic position of the target species and the recognition sequence of the selected restriction enzyme. These tools are then applied to test global-scale historical biogeographic hypotheses of vent fauna using barnacles as model. Phylogeographic inferences suggest that the western Pacific was the place of origin of the major vent barnacle lineage, followed by circumglobal colonization eastward along the southern hemisphere during the Neogene. I suggest that the geological processes and dispersal mechanisms discussed here can explain distribution patterns of many other marine taxa in addition to barnacles. Regional-scale analyses indicate that vent barnacle populations are well connected within basins and ridge systems, and that their diversity patterns do not conform to the predictions from the hypothesis that seamounts are centers of endemism. I then move on to resolve long-standing questions regarding species definitions in recalcitrant deep-sea coral taxa, by unambiguously resolving evolutionary relationships and objectively inferring species boundaries. Finally, I explore the adaptive potential of deep-sea coral species to environmental changes by examining a case of adaptation to shallow water from the deep sea. Candidate positive-selection markers shared between pairs of shallow and deep populations are identified as likely makers for genomic regions involved in adaptation. Overall, the results from this thesis constitute critical baseline data with which to assess potential effects of anthropogenic disturbances on deep-sea ecosystems.

Thesis supervisor: Dr. Timothy M. Shank Title: Associate Scientist with Tenure, Biology Department, Woods Hole Oceanographic Institution

Para mis papás, Liliana y Mauricio

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CHAPTER 1

Introduction

Threats to deep-sea coral ecosystems

Deep-sea corals are some of the most conspicuous invertebrate inhabitants of hard-bottom deep-sea benthic environments worldwide. They are not only more diverse in terms of number of species than their shallow counterparts (Cairns 2007), but they also play a fundamental role as foundation species and ecosystem engineers, creating three-dimensional habitats that are occupied by a high diversity of associate species (Buhl-Mortensen & Mortensen 2005; Costello et al. 2005; Buhl-Mortensen et al. 2010; Watling et al. 2011). Deep-sea ecosystems also support fisheries (D'Onghia et al. 2011) and have been identified as important sources of marine natural products (Leal et al. 2012). Deep-sea corals, generally speaking, have evolved in a relatively stable and energy-poor environment. They are characterized by slow growth rates. extreme longevity, low fecundity, and a late age of maturity (Roberts et al. 2009). These characteristics make deep-sea coral ecosystems fragile and with low resilience to the severe disturbances generated by many modern human activities, including physical damage caused by bottom-trawling fishing (Watling & Norse 1998; Koslow et al. 2001; Waller et al. 2007; Althaus et al. 2009; Clark & Rowden 2009; Williams et al. 2010), climate change and ocean acidification caused by emissions of greenhouse gasses (Doney et al. 2009), pollution and habitat destruction generated by waste disposal (Kvassnes et al. 2012), deep-sea mining (Van Dover 2010), and off-shore drilling for hydrocarbons (White et al. 2012). As such, the United Nations has designated deep-sea coral ecosystems as Vulnerable Marine Ecosystems (http://www.un.org/depts/los/general assembly/general assembly resolutions.htm UN General Assembly resolutions 61/105 and 64/72), requiring new management and protection strategies, such as the U.S. Magnuson-Stevens Fishery Conservation and Management Reauthorization Act of 2006. The Magnuson-Stevens act authorized Regional Fishery Management Councils to designate zones to protect deep-sea corals from damage caused by fishing gear, and established a Deep-Sea Coral Research and Technology Program to advance a critical understanding of their taxonomy and systematics, biogeography, genetic connectivity, and physiological responses to stressors in deep water coral ecosystems.

Threats to hydrothermal vent ecosystems

Benthic chemosynthetic ecosystems present a sharp contrast to other ecosystems in the deep-sea. They are characterized by high rates of *in situ* primary productivity, marked patchiness, and highly dynamic geological settings (Van Dover 2000). Deep-sea hydrothermal vent environments can have extremely

steep gradients of chemistry and temperature, and high disturbance frequencies given their occurrence on volcanic or actively spreading tectonic systems. As a consequence, hydrothermal vent environments present extreme selective pressures on evolutionary timescales, and can yield vent ecosystems with relatively low biodiversity and high endemism (Van Dover 2010). These very characteristics may make them susceptible to disturbances caused by mining of polymetallic sulfides. Although organisms at deep-sea hydrothermal vents have adapted to cope with natural disturbances, the frequency and magnitude at which these occur can vary greatly depending on the particular geophysical nature of each system (Baker & German 2004). Thus, disturbance from mining could have additive effects to natural disturbances at scales not previously experienced by these organisms, which could potentially lead to significant losses of biodiversity. There is a surging need to provide the highest-priority information needed to design optimal conservation and management strategies for areas that are being prospected for mining (Van Dover *et al.* 2012). These priorities emphasize the identification of conservation units at the genetic, species and biogeographic levels, and a better understanding of connectivity among populations.

Priorities for conservation of deep-sea ecosystems

Knowledge of conservation units is fundamental for the creation and implementation of efficient strategies that help mitigate the effects of harmful human activities on deep-sea ecosystems. Such knowledge must include well-founded taxonomic inventories that allow us to identify species and ecosystems at risk, as well as an understanding of their relatedness, genetic variance, distribution, connectivity patterns, and adaptation potential (Christensen *et al.* 1996; Dubois 2003; Roberts & Cairns 2014). Nevertheless, gaining this knowledge in deep-sea ecosystems is difficult due to the extreme challenges of working in these environments, combined with the paucity of genetic resources for deep-sea taxa.

Issues of traditional phylogenetic approaches

Traditionally, phylogenetic and population genetic studies in non-model organisms – which aims include understanding species boundaries, relationship patterns, evolutionary histories, factors that diminish or promote genetic diversity, demographic processes of populations, and interactions with environmental conditions – have based their power on a handful of homologous DNA sequence markers. Target DNA sequence markers can be easily sequenced using nearly-universal primers; however, several problems have been identified with the use of the few traditional sequence markers available for non-model organisms (e.g., mitochondrial and ribosomal genes), including low variability, biased loci sampling, poor genome representation and small statistical power, presence of pseudogenes, multiple gene copies, and non-independence caused by linkage (Brumfield 2003; Brito & Edwards 2008). Microsatellites emerged

during the last decade as a novel class of markers promising great potential to solve population-level questions. However, many problems have identified with the use of microsatellites, including poor understanding of their mutational processes, high rates of back mutations and homoplasy, presence of null alleles, low reproducibility and comparability of results, and high monetary and time expenses for individual marker development and genotyping (Brumfield 2003).

Opportunities of novel genomic approaches

The problems related to the use of traditional genetic markers have been recognized and accounted for in model organisms by comparing large amounts of genomic sequence information among individuals and identifying thousands of variable regions, such as single nucleotide polymorphisms (SNPs) across the genome, e.g. Clark *et al.* (2007) and Rokas *et al.* (2003). Single nucleotide polymorphisms (SNPs) have been recognized as the most prevalent source of variability in any given genome; they represent *ca.* 90% of the genetic variation in the human genome (Collins *et al.* 1999). As such, SNPs overcome most of the problems related with the use of traditional sequence-markers - they are present in extremely high numbers across the genome, have a wide range of mutational rates, behave largely as independent loci, and can be screened with high-throughput techniques, making them economical (Brumfield 2003). Technological and methodological developments in next-generation sequencing platforms over the last five years (e.g., Illumina, PacBio, IonTorrent, etc) have made genomic resources for SNP development and genotyping increasingly accessible and available to researchers investigating a wide spectrum of evolutionary questions in diverse organisms. Their use is now being successfully implemented in several non-model organisms, thus offering a great opportunity to overcome the difficulties inherent to the use of traditional approaches in many taxa.

SNPs have been successfully used to resolve evolutionary and biogeographic histories of diverse taxa, from bees (Whitfield *et al.* 2006), to humans (Jakobsson *et al.* 2008), nematodes (Andersen *et al.* 2012), anemones (Reitzel *et al.* 2013), and pitcher plant mosquitoes (Emerson *et al.* 2010). Recently developed methodologies allow the implementation of next-generation sequencing technologies for the rapid detection and genotyping of SNPs in organisms without reference genomes (Garvin *et al.* 2010); the so-called genotyping-by-sequencing (GBS) approaches (Baird *et al.* 2008). Restriction-site associated DNA sequencing (RAD-seq), a kind of GBS approach, enables high-throughput sequencing of homologous sites in nuclear genome after a complexity-reduction step carried out with a high-fidelity restriction enzyme. This revolutionary approach has opened a new frontier in molecular studies, with the promise of providing profound insights into the genetics, organismal biology, ecology and evolution of wild populations (Seeb *et al.* 2011).

Objective and outline

With this thesis, I aim to provide fundamental high-priority knowledge in taxonomic, genetic, evolutionary, and ecological aspects of deep-sea coral and hydrothermal vent species, by harnessing the power of novel genomic tools. This knowledge could be applied to the conservation and management of deep-sea ecosystems and their biodiversity.

To achieve this goal, I have performed bioinformatic and empirical studies implementing restriction-site associated DNA sequencing methodologies on ecologically important deep-sea coral and hydrothermal vent species. I present results from my investigations on the causes and consequences of evolutionary forces that determine biodiversity patterns in the deep-sea. In Chapter 2, I test the hypothesis that genome composition can be used to predict the number of restriction sites for a given combination of restriction enzyme and genome across the eukaryotic tree of life. I develop a methodology to predict the frequency of restriction sites that helps guide the design of GBS studies in eukaryotic organisms. This methodology is then used throughout the rest of the thesis. Chapters 3-6 can be categorized into chapters that examine macroevolutionary processes within species (chapters 4 and 6). Alternatively, they can also be categorized according to the target taxon system: deep-sea hydrothermal vent barnacles (chapters 3 and 4), or deep-sea corals (chapters 5 and 6) (Fig. 1).



Figure 1. Conceptual categorization of chapters according to the time scale of the processes and the target taxon systems examined.

In Chapter 3, I compare traditional DNA sequence makers with novel genomic data from restriction site associated DNA sequencing (RAD-seq) to characterize the global genetic diversity of barnacles from deep-sea hydrothermal vents, which due to their widespread but discontinuous distribution represent an excellent model for testing global biogeographic hypotheses. I infer their time and place of origin, mode of dispersal, and diversification throughout the world's vents. I then complement this research in Chapter 4, by using vent barnacles as a model to test smaller-scale biodiversity hypothesis. Specifically, I test the hypothesis that seamounts act as islands promoting divergence and speciation in deep-sea vent fauna. For this I compare the genetic diversity contained in single nucleotide polymorphisms (SNPs) obtained through RAD-seq to examine population-structuring patterns of populations in different barnacle species from seamount and spreading ridges.

In Chapter 5, I demonstrate the empirical utility of RAD-seq to solve evolutionary questions in deeper time by unambiguously resolving phylogenetic relationships among recalcitrant octocoral taxa with divergences greater than 80 million years and performing unambiguous species delimitations. In Chapter 6, I make use of the knowledge gained in defining species boundaries in deep-sea coral species by addressing questions regarding interactions of coral populations from the same species with their environment, in shorter time scales. I focus on a deep-sea coral species that can also be found in shallow high-latitude fjords with distinct environmental conditions from those found in the deep-sea. Here I aim to identify and characterize genomic regions that have may have enabled the successful adaptation to shallow-water in this deep-sea species. Finally, I summarize findings and draw general conclusions in Chapter 7.

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CHAPTER 2

Genome-wide predictability of restriction sites across the eukaryotic tree of life

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ABSTRACT

High-throughput sequencing of reduced representation libraries obtained through digestion with restriction enzymes – generally known as restriction site associated DNA sequencing (RAD-seq) – has become a common strategy to generate genome-wide genotypic and sequence data from eukaryotes. The choice of restriction enzyme is critical for the design of any RAD-seq study as it determines the number of genetic markers that can be obtained for a given taxon enabling a broad spectrum of applications, including marker discovery, population genomics, genomic mapping and phylogenetics. Here, we test the hypothesis that genome composition, in terms of GC content, and mono-, di- and trinucleotide composition, can be used to predict the number of restriction sites for a given combination of restriction enzyme and genome across the eukaryotic tree of life. Our analyses reveal that in most cases the trinucleotide genome composition model is the best predictor of the expected number of restriction sites in a eukaryotic genome, and the GC content and mononucleotide models the worst. We conclude that the predictability of restriction site frequencies in eukaryotic genomes needs to be treated on a case-specific

basis, whereby the phylogenetic position of the taxon of interest and the specific recognition sequence of the selected restriction enzyme are the chief foci among the most determinant factors. The software here developed, PredRAD (https://github.com/phrh/PredRAD), and the resulting databases constitute a valuable reference resource that will help guide the choice of restriction enzyme for any study using RAD-seq or related methods.

INTRODUCTION

The use of restriction enzymes to obtain reduced representation libraries from nuclear genomes, combined with the power of next-generation sequencing technologies, is rapidly becoming one of the most used strategies to generate genome-wide genotypic and sequence data in both model and non-model organisms (Baird *et al.* 2008; Andolfatto *et al.* 2011; Elshire *et al.* 2011; Peterson *et al.* 2012). The hundreds, thousands or tens of thousands of single nucleotide polymorphisms (SNPs) embedded in the resulting restriction site associated DNA (RAD) sequence tags (Miller *et al.* 2007; Baird *et al.* 2008) have a myriad of uses in biology ranging from genetic mapping (Wang *et al.* 2013; Weber *et al.* 2013) to population genomics (Hohenlohe *et al.* 2010; Andersen *et al.* 2012; White *et al.* 2013), phylogeography (Emerson *et al.* 2010; Reitzel *et al.* 2013), phylogenetics (Wagner *et al.* 2012; Eaton & Ree 2013), and SNP marker discovery (Scaglione *et al.* 2012; Toonen *et al.* 2013).

The choice of appropriate restriction enzyme(s) is critical for the effective design and application of RAD sequencing and a rapidly growing number of related methods such as genotyping-by-sequencing (Elshire *et al.* 2011), multiplexed shotgun genotyping (Andolfatto *et al.* 2011), double digest RAD-seq (Peterson *et al.* 2012), and ezRAD (Toonen *et al.* 2013). This choice determines the number of RAD markers that can be obtained, the amount of sequencing needed for a desired coverage level, the number of samples that can be multiplexed, the monetary cost, and ultimately the success of a project. It has been widely suggested that the number of restriction sites in a genome, for a given enzyme, can be roughly predicted using simple probability, if one has an estimate of the genome size and guanine-cytosine (GC) composition (Baird *et al.* 2008; Davey *et al.* 2011). Both of these parameters can be approximated in non-model organisms through sequencing-independent techniques such as flow cytometry (Vinogradov 1994, 1998; Šmarda *et al.* 2011). However, preliminary evidence suggests that there can be significant departures from expectations for particular combinations of taxa and restriction enzymes (Davey & Blaxter 2011; Davey *et al.* 2011).

Type II restriction enzymes, which are endonucleases chiefly produced by prokaryotic microorganisms, cleave double stranded DNA (dsDNA) at specific unmethylated recognition sequences that are 4 to 8 base pairs long and typically palindromic. These enzymes are thought to play an important role as defense systems against foreign phage dsDNA during infection or as selfish parasitic elements, and therefore have been the center of an evolutionary 'arms race' (Rambach & Tiollais 1974; Karlin *et al.* 1992; Rocha *et al.* 2001). Type II restriction enzymes are not known in eukaryotes and are not used as virulence factors by bacteria to infect eukaryotic hosts. Therefore there are no *a priori* reasons to believe that recognition sites in eukaryotic genomes are subject to selective pressures, but rather they should be evolutionarily neutral. Eukaryotic genomes have heterogeneous compositions with characteristic signatures at the level of diand trinucleotides that are largely independent of coding status or function (Karlin & Mrázek 1997; Karlin *et al.* 1998; Gentles & Karlin 2001). Thus, it is possible that genome composition at these levels has a large influence on the abundance of short sequence patterns such as recognition sequences of restriction enzymes.

Here, we test the hypothesis that genome composition can be used to predict the number of restriction sites for a given combination of restriction enzyme and taxon. For this we: i) performed systematic in silico genome-wide surveys of restriction sites for diverse type II restriction enzymes in 434 eukaryotic whole and draft genomes to determine their frequencies across taxa; ii) examined the composition of genomes at the level of di- and trinucleotides to determine patterns of compositional biases among taxa; iii) developed stochastic models based on GC content, and mono-, di- and trinucleotide compositions to predict the frequencies of restriction sites across taxa and diverse kinds of type II restriction enzymes; iv) evaluated the accuracy of the predictive models by comparing the in silico observed frequencies of restriction sites to the expected frequencies predicted by the models. The number of restriction sites in a genome is not the only factor that determines the number of RAD loci that can be recovered experimentally. The architecture of each genome, and in particular the number of repetitive elements and gene duplications, can significantly decrease the number of unambiguous loci obtained via alignment to a reference genome or *de novo* assembly. To quantify this contribution we assessed the proportion of restriction-site associated DNA tags that can potentially be recovered unambiguously after empirical sequencing. The software here developed, PredRAD (https://github.com/phrh/PredRAD), and the resulting databases constitute a reference resource that will help guide the choice of restriction enzyme for any study using RAD-related methods.

RESULTS

Observed frequencies of restriction sites

To explore restriction site frequencies across the Eukaryotic tree of life we surveyed recognition sequences for 18 commonly used palindromic type II restriction enzymes in 434 whole and draft genomes. Observed frequencies of restriction sites were highly variable among broad taxonomic groups for the set of restriction enzymes here examined (Table 1) – except for FatI – with clear clustering patterns determined by phylogeny (Fig. 1). For example, with NgoMIV we observed 45.8 restriction sites per megabase (RS/Mb) \pm 24.6 (mean \pm SD) in core eudicot plants, compared to 277.4 \pm 131.3 RS/Mb in commelinid plants (monocots). Among closely-related species the frequency patterns were similar and variability generally small. Observed frequencies of RS/Mb were inversely proportional to the length of the recognition sequence, with orders of magnitude differences among the 4-, 6-, and 8- cutters when compared within the same species. For example, in the starlet anemone *Nematostella vectensis* there were 3917.6, 167.6, and 6.9 RS/Mb for the 4-cutter FatI, 6-cutter PstI and 8-cutter SbfI, respectively. In contrast, nucleotide composition of the recognition sequence did not show a clear correlation with the observed frequency of restriction sites. For example, 83.6 RS/Mb \pm 25.1 were observed in Neopterigii vertebrates for KpnI (GGTACC) and 622.6 RS/Mb \pm 119.1 were observed for PstI (CTGCAG), both recognition sequences with a GC content of 66.7%.

Dinucleotide compositional biases

Dinucleotide odds ratios ($\bar{\rho}_{XY}^{*}$) (Burge *et al.* 1992), a measurement of relative dinucleotide abundances given observed component frequencies used to explore genomic compositional biases, revealed significant compositional biases for all possible dinucleotides (Fig. 2). Both dinucleotides and trinucleotides are considered significantly underrepresented if the odds ratio is ≤ 0.78 , significantly overrepresented if ≥ 1.23 , and equal to expectation if =1 (Karlin *et al.* 1998). The dinucleotide compositional biases were highly variable among broad taxonomic groups (e.g., core eudicot plants) but generally similar within. Two dinucleotide complementary pairs, CG/GC and AT/TA, had highly dissimilar relative frequencies between the members of each pair. The largest biases were for CG, being significantly underrepresented in groups like core eudicot plants ($\bar{\rho}_{XY}^{*}=0.68 \pm 0.11$), gnathostomate vertebrates ($\bar{\rho}_{XY}^{*}=0.32 \pm 0.12$), the Pucciniales rust fungi ($\bar{\rho}_{XY}^{*}=0.61 \pm 0.19$) and the Saccharomycetales yeast ($\bar{\rho}_{XY}^{*}=0.78 \pm 0.17$). CG was significantly overrepresented in groups like the Apocrita insects $(\bar{\rho}_{XY}^*=1.59 \pm 0.18)$. The complementary dinucleotide GC was not particularly underrepresented in any broad taxonomic group, but tended towards overrepresentation in ecdysozoan invertebrates ($\bar{\rho}_{XY}^*=1.24 \pm 0.12$), being significant in several arthropod and nematode species. Other taxa that showed significant overrepresentation of GC dinucleotides included the Trebouxiophyceae ($\bar{\rho}_{XY}^*=1.39 \pm 0.04$) and microsporidia fungi ($\bar{\rho}_{XY}^*=1.28 \pm 0.17$). Relative abundances of the dinucleotide AT were within expectations for all eukaryotes, except for the fungus *Sporobolomyces roseus* ($\rho_{XY}^*=0.78$). Contrastingly, the TA dinucleotide tended towards underrepresentation throughout the eukaryotes ($\bar{\rho}_{XY}^*=0.8 \pm 0.13$), except in a few hypocreomycetid fungal species, for which it was significantly underrepresented. The TA dinucleotide was significantly underrepresented in trypanosomatids ($\bar{\rho}_{XY}^*=0.79 \pm 0.03$), choanoflagellids ($\bar{\rho}_{XY}^*=0.43 \pm 0.09$), chlorophytes ($\bar{\rho}_{XY}^*=0.62 \pm 0.15$), stramenopiles ($\bar{\rho}_{XY}^*=0.76 \pm 0.03$) and the Basidiomycota ($\bar{\rho}_{XY}^*=0.74 \pm 0.09$), among others.

The remaining dinucleotides had identical relative frequencies between the members of each complementary pair. The dinucleotide pair GG/CC was marginally underrepresented in most eukaryotes $(\bar{\rho}_{XY}^*=0.88 \pm 0.15)$. In the Sarcopterygii vertebrates $(\bar{\rho}_{XY}^*=1.02 \pm 0.06)$ and embryophyte plants $(\bar{\rho}_{XY}^*=1.03 \pm 0.06)$ ± 0.07) GG/CC relative frequencies closely conformed to expectation, whereby GG/CC was significantly overrepresented in handful of isolated ecdysozoan, microsporidia and alveolate species, and significantly underrepresented in chlorophytes ($\bar{\rho}_{XY}^*=0.72$, SD=0.11), oomycetes ($\bar{\rho}_{XY}^*=0.71 \pm 0.05$), and in several species of the Basidiomycota and the Dothideomycetes. Only the choanoflagellate Salpingoeca and the green alga Asterochloris presented a marginally significant bias for the dinucleotide pair AA/TT ($\rho_{XY}^*=0.77$ and 0.75 respectively). Similarly, Salpingoeca was the only taxon to show a significant bias for AC/GT ($\rho_{XY}^*=1.42$). Dinucleotide pair CA/TG was among the pairs with largest biases. Significant overrepresentation of CA/TG was found in several groups with large CG underrepresentation such as gnathostomates ($\bar{\rho}_{XY}^*=1.31 \pm 0.05$), gastropods ($\bar{\rho}_{XY}^*=1.29 \pm 0.05$), the Pucciniales ($\bar{\rho}_{XY}^*=1.27 \pm 0.02$), the Trebouxiophyceae ($\bar{\rho}_{XY}^* = 1.62 \pm 0.14$), as well as several species of core eudicots and the Saccharomycetales. Other groups with significant CA/TG overrepresentation include onchocercid nematodes ($\bar{\rho}_{XY}^*=1.26 \pm 0.01$), the Ustilaginomycotina fungi ($\bar{\rho}_{XY}^*=1.28 \pm 0.05$), trypanosomatids $(\bar{\rho}_{XY}^*=1.25 \pm 0.04)$, and amoebozoans $(\bar{\rho}_{XY}^*=1.33 \pm 0.06)$. Overrepresentation biases for the AG/CT

dinucleotide pair were only present in amniotes ($\bar{\rho}_{XY}^*=1.26 \pm 0.02$), the Sporidiobolales fungi ($\bar{\rho}_{XY}^*=1.24 \pm 0.01$), and oxytrichid alveolates ($\bar{\rho}_{XY}^*=1.24 \pm 0.04$), and other isolated species. Most of these taxa also had large CG underrepresentation. Lastly, most eukaryotes had GA/TC relative frequencies that conformed to expectations, except for few scattered species and small groups such as the Microbotryomycetes fungi ($\bar{\rho}_{XY}^*=1.45 \pm 0.13$), the Mamiellales green algae ($\bar{\rho}_{XY}^*=1.40 \pm 0.08$), and the Eimeriorina alveolates ($\bar{\rho}_{XY}^*=1.26 \pm 0.02$).

Trinucleotide compositional biases

Trinucleotide odds ratios (γ_{XYZ}^*) (Burge *et al.* 1992) are another important measurement used to explore genomic compositional biases. Among the examined taxa, these ratios revealed compositional biases for most possible trinucleotides (Fig. 3). However, most of these biases were only significant in scattered individual species (Fig. 4). Among the trinucleotide pairs with significant underrepresentation, CTA/TAG and CGA/TCG showed the most definite broad taxonomic patterns. CTA/TAG was significantly underrepresented in most taxa, except for groups like commelinid plants (monocots) ($\gamma_{XYZ}^*=0.87 \pm 0.03$), most core eudicots ($\gamma_{XYZ}^*=0.81 \pm 0.02$), eleutherozoans ($\gamma_{XYZ}^*=0.82 \pm 0.01$), molluscs ($\gamma_{XYZ}^*=0.83 \pm 0.01$), and gnathostomates ($\gamma_{XYZ}^*=0.82 \pm 0.02$) – exclusive of the chimaera *Callorhinchus milii*. Contrastingly, the trinucleotide CGA/TCG was only significantly underrepresented in most tetrapod vertebrates ($\gamma_{XYZ}^*=0.82 \pm 0.02$) – exclusive of muroid rodents, bovid ruminants and the Afrotheria – a group containing aarvdvarks, hyraxes, and elephants.

The largest and more widespread overrepresentation biases were for the trinucleotide pair AAA/TTT, being significant in most eukaryotes, except for the majority of the Dikarya fungi ($\gamma_{XYZ}^*=1.18 \pm 0.07$). The trinucleotide pairs TAA/TTA and AAT/ATT were significantly overrepresented in many metazoan taxa, particularly in the Neopterygii vertebrates ($\gamma_{XYZ}^*=1.3 \pm 0.05$, and $\gamma_{XYZ}^*=1.26 \pm 0.05$ respectively). AAG/CTT was significantly overrepresented in the Bacillariophyta diatoms ($\gamma_{XYZ}^*=1.24 \pm 0.03$), oomycetes ($\gamma_{XYZ}^*=1.28 \pm 0.02$), and the Saccharomycetales ($\gamma_{XYZ}^*=1.26 \pm 0.04$). Lastly, CCA/TTG was significantly overrepresented in several tetrapod groups, including the Laurasiatheria – exclusive of the Chiroptera – ($\gamma_{XYZ}^*=1.25 \pm 0.02$) and Hominoidea ($\gamma_{XYZ}^*=1.23 \pm 0.004$).

Core Sequence	Restriction Enzyme	Recognition Sequence	Recognition Sequence Length	GC Content of Recongition Sequence
GGCC			U	-
	NotI	GCGGCCGC	8	100.0
CCGG				
	SgrAI	CRCCGGYG	8	87.5
	BsrFI	RCCGGY	6	83.3
	NgoMIV	GCCGGC	6	100.0
	AgeI	ACCGGT	6	66.7
	MspI	CCGG	4	100.0
TGCA				
	SbfI	CCTGCAGG	8	75.0
	PstI	CTGCAG	6	66.7
	NsiI	ATGCAT	6	33.3
AATT				
	ApoI	RAATTY	6	16.7
	EcoRI	GAATTC	6	33.3
	MluCI	AATT	4	0.0
TTAA				
	MseI	TTAA	4	0.0
CATG				
	NspI	RCATGY	6	50.0
	NcoI	CCATGG	6	66.7
	PciI	ACATGT	6	33.3
	FatI	CATG	4	50.0
GTAC				
	KpnI	GGTACC	6	66.7

 Table 1. Restriction enzymes included in this study.



Figure 1. Observed restriction site frequencies. Left: phylogenetic tree of all eukaryotic taxa analyzed in this study. The tree is based on the NCBI taxonomy tree retrieved on May 16, 2013 using the iTOL tool <u>http://itol.embl.de</u> (Letunic & Bork 2011). Branch colors and labels indicate broad taxonomic groups. Organism silhouettes and cartoons were created by the authors or obtained from <u>http://phylopic.org/</u>. Right: heatmap of the observed frequency of restriction sites. Each row corresponds to a species from the tree on the left, and each column corresponds to a different restriction enzyme. Gray line in the color-scale box shows the distribution histogram of all values.



Figure 2. Dinucleotide compositional biases and significances. Left: phylogenetic tree as in Fig. 1. Center: heatmap of the ρ_{XY}^* odds ratio values. Right: heatmap of the ρ_{XY}^* odds ratio significant values $\rho_{XY}^* < 0.78$ and $\rho_{XY}^* > 1.23$. Each row corresponds to a species from the tree on the left, and each column corresponds to a different dinucleotide. Green indicates underrepresentation and red indicates overrepresentation. Cyan line in the color-scale box shows the distribution histogram of all values.





Figure 3. Trinucleotide compositional biases. Left: phylogenetic tree as in Fig. 1. Right: heatmap of the γ_{XYZ}^* odds ratio values. Each row corresponds to a species from the tree on the left, and each column corresponds to a different trinucleotide. Green indicates underrepresentation and red indicates overrepresentation. Cyan line in the color-scale box shows the distribution histogram of all values.







Figure 4. Trinucleotide compositional biases significances. Left: phylogenetic tree as in Fig. 1. Right: heatmap of the γ_{XYZ}^* odds ratio significant values $\rho_{XY}^* < 0.78$ and $\rho_{XY}^* > 1.23$. Each row corresponds to a species from the tree on the left, and each column corresponds to a different trinucleotide. Green indicates underrepresentation and red indicates overrepresentation. Cyan line in the color-scale box shows the distribution histogram of all values.

Expected frequencies of restriction sites

To test the hypothesis that compositional heterogeneity in eukaryotic genomes can determine the frequency of restriction sites of each genome we developed probability models based on the GC content of each genome, as well as the mononucleotide, dinucleotide and trinucleotide compositions to predict the expected frequency of recognition sequences for each restriction enzyme. We evaluated the fit of each model using a similarity index (SI), defined as the quotient of the number of observed and expected restriction sites, minus one. A positive SI indicates that the number of observed restriction sites is greater than the expected, whereas a negative SI indicates a smaller number of observed sites than expected. If SI is equal to 0, then the number of observed sites is equal to the expectation. For example, a SI = 1 indicates that the number of observed restriction sites for a particular enzyme in a given genome is twice the number of expected sites predicted by a particular model. Trinucleotide composition models were in general a better predictor, in terms of their accuracy and precision, of the expected number of restriction sites than any of the other models (Fig. 5, Fig. 6). The mononucleotide and GC content models produced relatively poor predictions that were indistinguishable from one another (Fig. 5, Fig. 6). In a few cases the other models outperformed the trinucleotide model, e.g., EcoRI (Fig. 5, Fig. 6, Fig. 7). The fit of the predictions was highly variable among broad taxonomic groups but generally similar within, e.g., in Neopterigii vertebrates an average SI of 0.14 \pm 0.19 for AgeI with the dinucleotide model, compared to - 0.31 ± 0.19 in Sarcopterigii.

Recovery of RAD-tags after in silico sequencing

In most cases, the recovery of RAD-tags after *in silico* sequencing was notably high, with a median percentage of suppressed alignments to the reference genome assembly of only 3% (Fig. 8). There was no evident recovery bias by restriction enzyme, but rather bias was pronounced in a few individual species, likely indicating an enrichment of repetitive regions or duplications.



Figure 5. Overall fit of genome composition models per restriction enzyme. Vertical axes in the box and whisker plots indicate the values of the similarity index (*SI*) for each species per enzyme (see Methods section). Horizontal axes in the box and whisker plots indicate the genome composition model: GC content (gc), mononucleotide (mono), dinucleotide (di), and trinucleotide (tri). Horizontal edges of range boxes indicate the first and third quartiles of the *SI* values under each composition model. The thick horizontal black line represents the median. Whiskers indicate the value of 1.5 times the inter-quartile range from the first and third quartiles. Outliers are defined as SI values outside the whiskers range and are represented by dots. Outlier value of *Entamoeba histoyitica* for *NotI* was excluded. Red dotted lines indicate *SI*=0.


Figure 6. Similarity indexes for dinucleotide and trinucleotide genome composition models. Left: phylogenetic tree as in Fig. 1. Center: heatmap of the similarity indexes for the dinucleotide model Right: heatmap of the similarity indexes for the trinucleotide model. Each row corresponds to a species from the tree on the left, and each column corresponds to a different restriction enzyme. Cyan indicates SI < 0 and yellow indicates SI > 0. Red line in the color-scale box shows the distribution histogram of all values.



Figure 7. Similarity indexes for GC content and mononucleotide genome composition models. Left: phylogenetic tree as in Fig. 1. Center: heatmap of the similarity indexes for the GC content model Right: heatmap of the similarity indexes for the mononucleotide model. Each row corresponds to a species from the tree on the left, and each column corresponds to a different restriction enzyme. Cyan indicates SI < 0 and yellow indicates SI > 0. Red line in the color-scale box shows the distribution histogram of all values.



Figure 8. Recovery of RAD-tags after *in silico* genome digestion and sequencing. Left: phylogenetic tree as in Fig. 1. Right: heatmap of the percentage of RAD-tags that produced more than one unique alignment to their reference genome. Each row corresponds to a species from the tree on the left, and each column corresponds to a different restriction enzyme. Green line in the color-scale box shows the distribution histogram of all values.

DISCUSSION

Genome-wide surveys of restriction sites

Observed cut frequencies for a given restriction enzyme are strikingly variable among broad eukaryotic taxonomic groups, but similar among closely-related species. This is consistent with the hypothesis that the abundance of restriction sites is largely determined by phylogenetic relatedness. This pattern is most evident in groups that have a larger taxonomic representation, such as mammals. As more genome assemblies become available the pattern resolution will become clearer in many other underrepresented taxonomic groups, and through the use of comparative methods in a robust phylogenetic framework, it will be possible to establish taxon-specific divergence thresholds diagnostic of significant evolutionary changes in genome architecture.

As expected, observed frequencies of restriction sites with shorter recognition sequences are generally higher than the observed frequencies with longer recognition sequences. However this pattern in not universal. There are several instances in which the frequency of restriction sites for a high-denomination cutter is higher than for a low-denomination cutter. For example, in primates the frequency of the 8-cutter Sbf1 (24.6 \pm 1.7 RS/Mb) is significantly higher than the frequency of the 6-cutter AgeI (18.4 \pm 1.4 RS/Mb). These deviations from expectation are indicative of enzyme-specific frequency biases for particular taxa, and, as illustrated in the results section, are not correlated with the base composition of recognition sequences.

Genomic compositional biases

Our analyses indicate that there are significant compositional biases for most dinucleotides and trinucleotides across the eukaryotes. Many of these biases are significant only within individual species scattered throughout the eukaryotic tree of life. However, there are several particular dinuclotides and trinucleotides that show significant biases across the eukaryotic tree of life. Our observation that these biases are highly variable among broad taxonomic groups but generally similar within is congruent with findings from previous studies (e.g., Gentles & Karlin 2001). The most obvious biases across taxa are observed in the gnatostomate vertebrates; however, this is most likely due to rampant undersampling in

most other groups of eukaryotes (vertebrate genome assemblies represent 21% of all the taxa in this study).

The dinucleotides CG, GC, TA, and CA/TG show the most conspicuous bias patterns across the eukaryotic tree of life. Biases in most of these dinucleotides are likely linked to important biological processes. Notably the underrepresented dinucleotide CG is a widely known target for methylation related to transcriptional regulation (Bird 1980) and retrotransposon inactivation (Yoder *et al.* 1997) in vertebrates and eudicots. The corresponding overrepresentation of AG/CT fits the classic model of "methylation-deamination-mutation" by which a methylated cytosine in the CG pair tends to deaminate when unpaired and mutate into a thymidine with a corresponding CA complement. Interestingly CG and GC dinucleotides are significantly overrepresented in several groups of apocritic insects, as well as in some fungi and single-cell eukaryotes. CG is not a primary target for methylation in *Drosophila* (Lyko *et al.* 2000), instead CT, and in lesser degree CA and CC, are methylated in higher proportion. None of these dinucleotide pairs is significantly underrepresented in apocritic insects. The widespread TA underrepresentation has been traditionally attributed to stop codon biases, thermodynamic instability and susceptibility of UA to cleavage by RNAses in RNA transcripts (Beutler *et al.* 1989).

The trinucleotides CTA/TAG, AAA/TTT, TAA/TTA, CCA/TGG show the most conspicuous bias patterns across the eukaryotic tree of life. The biases in CTA/TAG have been widely attributed to the stop codon nature of UAG. However, the trinucleotides corresponding to the other stop codons (Burge *et al.* 1992), UAA and UGA, are overrepresented or not biased across eukaryotes. The reasons behind other cases of trinucleotide biases are less understood.

Predictability of restriction site frequencies

Our analyses indicate that in most cases the trinucleotide genome composition model is the best predictor, and the GC content and mononucleotide models are the worst predictors of the expected number of restriction sites in a eukaryotic genome. It is possible that the greater number of parameters in the trinucleotide model (64, compared to 16, 4 and 2 of the dinucleotide, mononucleotide and GC content model, respectively) is the cause of the better fit. However, this trend is not universal. As illustrated in the results section, the other models in a few cases outperformed the trinucleotide composition model. Neither the GC content nor length of the recognition sequence can confidently explain the observed discrepancies. It is not surprising that fit of the predictions made by the models is highly variable across taxonomic groups, given the high variability observed in the frequencies of restriction sites and genetic compositions across the eukaryotic tree of life. We conclude that the predictability of restriction site frequencies in eukaryotic genomes needs to be treated on a case-specific basis, whereby the phylogenetic position of the taxon of interest and the specific recognition sequence of the selected restriction enzyme are the chief foci among the most determinant factors.

Implications for RAD-seq and related methodologies

For the design of a study using RAD-seq, or a related methodology, there are two fundamental questions that researchers face: i) what is the best restriction enzyme to use to obtain a desired number of RAD tags in the organism of interest? and ii) how many markers can be obtained with a particular enzyme in the organism of interest? The results from this study coupled with the developed software pipeline PredRAD, will allow any researcher to obtain an approximate answer to these questions.

In a best-case scenario for the practical design of a study using RAD-seq, or a related methodology, the species of interest is already included in the database presented here. In this case, the best proxy for the estimated number of RAD tags that could be obtained empirically would be twice the number of *in silico* observed restriction sites for each restriction enzyme (each restriction site is expected to produce two RAD tags, one in each direction from the restriction site) minus the number of RAD tags for SbfI in starlet anemone *Nematostella vectensis* is 3,370, being highly similar to the range of RAD tags (2,300-2,800) obtained empirically by Reitzel *et al.* (2013). For library preparation protocols in which a fragment size selection step is done without a prior shearing step, e.g., ddRAD (Peterson *et al.* 2012) and ezRAD (Toonen *et al.* 2013), the size.select function of the software package SimRAD (Lepais & Weir 2014) constitutes a valuable complementary study-design tool. If a new genome assembly becomes available for the target species and/or the researcher wishes to evaluate an additional restriction enzyme, PredRAD can be re-executed with these data to quantify the number of restriction sites and the recovery potential, as well as to estimate the probability of the new recognition sequence based on genome composition models.

In the scenario that the genome sequence of the species of interest is not available, the best alternative is to look at the closest relative with a genome assembly. A range of approximate values for the number of RAD tags can be obtained from i) the number of *in silico* observed restriction sites in the closely related species; ii) the frequency of restriction sites in the closely related species, and the genome size of the species of interest; and iii) the probability of the recognition sequence for the enzyme(s) based on the best-fit genome composition model (*SI* closest to 0) from the closely related species and the genome size of the target species. The genome size of the species of interest can be estimated through sequencing-independent techniques such as flow cytometry (Vinogradov 1994, 1998; Šmarda *et al.* 2011). For

example, the predicted range in the number of RAD tags that could be obtained using SbfI in thoracican barnacles, a group for which no genome sequence assemblies are available, is 10,000-30,000, based on the observed frequency of the SbfI recognition sequence and its probability using a trinucleotide composition model in the genome of the crustacean *Daphnia pulex* (ranges of genome size for barnacles were obtained from the Animal Genome Size Database, <u>http://ww.genomesize.com</u>). Herrera *et al.* (Chapter 3) empirically obtained *ca.* 18,000 RAD tags for several species in this group. The possibility that the frequency of restriction sites and genome composition can be accurately estimated from alternative datasets such as a transcriptome is worth evaluating during restriction-enzyme selection for taxa without sequenced genomes.

Additional factors that can influence the number of RAD tag markers that can be obtained experimentally, and need to be considered during study design and data analysis steps, include: genome differences among individuals, level of heterozygosity, the amount of methylation in the genome, the number of repetitive regions and gene duplicates present in the target genome, the sensitivity of a particular restriction enzyme to methylation, the efficiency of the enzymatic digestion, the quality of library preparation and sequencing, the amount of sequencing, sequencing and library preparation biases, and the parameters used to clean, cluster and analyze the data, among others (see Davey *et al.* (2013), (Catchen *et al.* 2013), DaCosta and Sorenson (2014), and Mastretta-Yanes *et al.* (2014) for further discussion).

CONCLUSIONS

In this study we tested the hypothesis that genome composition can be used to predict the number of restriction sites for a given combination of restriction enzyme and genome. Our analyses reveal that in most cases the trinucleotide genome composition model is the best predictor, and the GC content and mononucleotide models are the worst predictors of the expected number of restriction sites in a eukaryotic genome. We conclude that the predictability of restriction site frequencies in eukaryotic genomes needs to be treated on a case-specific basis, whereby the phylogenetic position of the taxon of interest and the specific recognition sequence of the selected restriction enzyme are the chief foci among the most determinant factors. The results from this study and the software developed from it will help guide the design of any study using RAD sequencing and related methods. As more genome assemblies become available in underrepresented taxonomic groups the patterns of compositional biases and restriction site frequencies across the eukaryotic tree of life will become clearer and will improve our understanding of genome evolution.

METHODS

Observed frequencies of restriction sites

Assemblies from eukaryotic whole genome shotgun (WGS) sequencing projects available as of December 2012 were retrieved primarily from the U.S. National Center for Biotechnology Information (NCBI) WGS database (Additional file 1). Only one species per genus was included. Of the 434 genome assemblies included in this study, 42% corresponded to fungi, 21% to vertebrates, 16% invertebrates, and 9% plants. Only unambiguous nucleotide calls were taken into account. Genome sequence sizes were measured as the number of unambiguous nucleotides in the assembly. A set of 18 commonly used palindromic type II restriction enzymes with variable nucleotide compositions was screened in each of the genome assemblies (Table 1). The number of restriction sites present in each genome was obtained by counting the number of unambiguous matches for each recognition sequence pattern. Under optimal experimental conditions each restriction site should produce two RAD tags in each direction from the restriction site. Therefore, we define the number of observed RAD tags in each genome assembly as twice the number of recognition sequence pattern matches.

Expected frequencies of restriction sites

To test the hypothesis that compositional heterogeneity in eukaryotic genomes can determine the frequency of restriction sites of each genome, we characterized the GC content, as well as the mononucleotide, dinucleotide and trinucleotide compositions of each genome and developed probability models to predict the expected frequency of recognition sequences for each restriction enzyme. GC content was calculated as the proportion of unambiguous nucleotides in the assembly that are either guanine or cytosine, assuming that the frequency of guanine is equal to the frequency of cytosine. Mononucleotide compositions were determined as the frequency of each one of the four nucleotides. Dinucleotide combinations, respectively. The odds ratios proposed by Burge *et al.* (1992) were used to estimate compositional biases of dinucleotides (1) and trinucleotides (2) across genomes.

(1)

$$\rho_{XY}^* = \frac{f_{XY}^*}{f_X^* f_Y^*}$$

(2)

$$\gamma_{XYZ}^{*} = \frac{f_{XYZ}^{*} f_{X}^{*} f_{Y}^{*} f_{Z}^{*}}{f_{XY}^{*} f_{YZ}^{*} f_{XNZ}^{*}}$$

Where f_X^* is the relative frequency of the mononucleotide X, f_{XY}^* is the relative frequency of the dinucleotide XY, and f_{XYZ}^* is the relative frequency of the trinucleotide XYZ. All frequencies take into account the antiparallel structure of double stranded DNA. N represents any mononucleotide.

Mononucleotide and GC content sequence models were used to estimate the probability of a particular recognition sequence (3) assuming that each nucleotide is independent of the others and of its position on the recognition sequence. The GC content model assumes that the relative frequencies of guanine and cytosine in the genome sequence are equal. This model has only two parameters, the GC and AT frequencies. In the mononucleotide model there are four parameters, one for each of the four possible nucleotides.

(3)

$$p(s) = \prod_{i=1,\dots,n(s)} p(s_i)$$

Here, $p(s_i)$ is the probability of nucleotide s_i at the position *i* of the recognition sequence. In the GC content model $p(s_i)$ can take the values of $f_{G,C}$ or $f_{A,T}$. In the mononucleotide model $p(s_i)$ can take the values of f_A , f_G , f_C , or f_T . Where f_X is the frequency of a given mononucleotide.

Dinucleotide and trinucleotide sequence models were defined as first and second degree Markov chain transition probability models with 16 or 64 parameters, respectively (Karlin *et al.* 1992; Singh 2009). These models take into account the position of each nucleotide in the recognition sequence. Nucleotides along the recognition sequence are not independent from nucleotides in neighboring positions. The probability of a particular recognition sequence for these Markov chain models was calculated as:

(4)

$$p(s) = p(s_1) \prod_{i=2,...,n(s)} p_c(s_i | s_{i-1}, ..., s_{i-n})$$

Where $p(s_1)$ is the probability at the first position on the recognition sequence and p_c is the conditional probability of a subsequent nucleotide on the recognition sequence depending on the previous n nucleotides. In the dinucleotide sequence model n = 1 and in the trinucleotide sequence models n = 2.

Expectations versus observations

To assess the effectiveness of the predictive recognition sequence models, we compared the number of observed restriction sites in the genome assemblies with the expected number. The expected number of restriction sites in a given genome was calculated as the product of the probability of a recognition sequence multiplied by the genome sequence size. To quantify the departures from expectation, we define a similarity index (SI) as SI = (O - E)/E, where O and E are the observed and expected number of restriction sites, respectively. If SI = 0, then E = O. If SI < 0, then E > O, and vice versa.

Recovery of restriction-site associated DNA tags

To assess the proportion of restriction-site associated DNA tags that can potentially be recovered unambiguously after empirical sequencing, we performed *in silico* sequencing experiments for all genome assembly-restriction enzyme combinations. For each restriction site located in the genome assemblies, 100 base pairs up- and down-stream of the restriction site were extracted. This sequence read length is typical of sequencing experiments performed with current Hi-Seq platforms (Illumina Inc.). The resulting RAD tags were aligned back to their original genome assemblies using BOWTIE v0.12.7 (Langmead *et al.* 2009). Only reads that produced a unique best alignment were retained.

The analytical software pipeline here described and output database files are publicly available at https://github.com/phrh/PredRAD.

AUTHOR'S CONTRIBUTIONS

SH conceived and designed research. SH and PHR developed the software. SH analyzed the data. SH, TMS and HW contributed computing equipment. SH wrote the paper with comments from PHR and TMS.

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SUPPORTING INFORMATION

Table S1. Genome assemblies included in this study. Note that web addresses to individual assembly files, and the assembly files themselves, were current as of December 2012 and may have changed.

TTN	^	•
113	- Nn	PUPE
112	$-\omega \mu$	

Genome Assembly

Acanthamoeba castellanii http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEYA01.fasta.gz Acacas Aciculosporium take Acitak http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFQZ01.fasta.gz Aciric Acidomyces richmondensis http://genome.jgi-psf.org/Aciri1_iso/download/Aciri1_iso_AssemblyScaffolds.fasta.gz http://genome.jgi-psf.org/Acral2/download/Acral2_AssemblyScaffolds.fasta.gz Acremonium alcalophilum Acralc Acromyrmex echinatior Acrech http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEVX01.fasta.gz Acrdig Acropora digitifera http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=BACK01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABLF02.fasta.gz Acypis Acyrthosiphon pisum http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAGE02.fasta.gz Aedes aegypti Aedaeg Agaricus bisporus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEOK01.fasta.gz Agabis Ailuropoda melanoleuca Ailmel http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACTA01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAJI01.fasta.gz Ajellomyces capsulatus Ajecap Alatina moseri http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AHZO01.fasta.gz Alamos Albugo candida http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CAJG01.fasta.gz Albcan http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AKHW01.fasta.gz Allmis Alligator mississippiensis http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACDU01.fasta.gz Allomyces macrogynus Allmac Altarb Alternaria arborescens http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AIIC01.fasta.gz http://genome.jgi-psf.org/Amamu1/download/Amamu1_AssemblyScaffolds.fasta.gz Amamus Amanita muscaria ttp://ttp.ncbi.nlm.nih.gov/genbank/genomes/Eukaryotes/invertebrates/Amphimedon_queenslandica/v1.0/I Ampque Amphimedon queenslandica Anncaliia algerae http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CAIR01.fasta.gz Annalg Anolis carolinensis Anocar http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAWZ02.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAAB01.fasta.gz Anogam Anopheles gambiae Apimon Apiospora montagnei http://genome.jgi-psf.org/Apimo1/download/Apimo1_AssemblyScaffolds.fasta.gz Apis mellifera Apimel http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AADG06.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AASC02.fasta.gz Aplysia californica Aplcal Arabidopsis thaliana http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFNA01.fasta.gz Aratha Arthrobotrys oligospora Artoli http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADOT01.fasta.gz Arthroderma benhamiae http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABSU01.fasta.gz Artben Ascaris suum http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AMPH01.fasta.gz Ascsuu http://genome.jgi-psf.org/Ascru1/download/Ascru1_AssemblyScaffolds.fasta.gz Ascrub Ascoidea rubescens http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AARE01.fasta.gz Ascosphaera apis Ascapi Aspergillus fumigatus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAHF01.fasta.gz Aspfum http://genome.jgi-psf.org/Astpho1/download/Astpho1_genomic_scaffolds.fasta.gz Asterochloris sp Ast_sp Attcep Atta cephalotes http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADTU01.fasta.gz Aurantiochytrium limacinum http://genome.jgi-psf.org/Aurli1/download/Aurli1_AssemblyScaffolds.fasta.gz Aurlim Aurpul Aureobasidium pullulans http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AMCU01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACJI01.fasta.gz Aureococcus anophagefferens Aurano Auricularia delicata http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFVO01.fasta.gz Aurdel Babesia bovis Babbov http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAXT01.fasta.gz http://genome.jgi-psf.org/Babin1/download/Babin1_AssemblyScaffolds.fasta.gz Babjeviella inositovora Babino

Batden Batrachochytrium dendrobatidis http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADAR01.fasta.gz Baucom Baudoinia compniacensis http://genome.jgi-psf.org/Bauco1/download/Bauco1_AssemblyScaffolds.fasta.gz **Beabas** Beauveria bassiana http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADAH01.fasta.gz Betnan Betula nana http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CAOK01.fasta.gz **Bigelowiella** natans Bignat http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADNK01.fasta.gz Bjeadu Bjerkandera adusta http://genome.jgi-psf.org/Bjead1_1/download/Bjead1_1_AssemblyScaffolds.fasta.gz Blahom Blastocystis hominis http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CABX01.fasta.gz Blugra Blumeria graminis http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABSB02.fasta.gz Bomter Bombus terrestris http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AELG01.fasta.gz Bommor Bombyx mori http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=BABH01.fasta.gz Bostau Bos taurus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAFC03.fasta.gz Botbot Botryobasidium botryosum http://genome.jgi-psf.org/Botbo1/download/Botbo1_AssemblyScaffolds.fasta.gz Botfuc Botryotinia fuckeliana http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ALOC01.fasta.gz Bradis Brachypodium distachyon http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADDN01.fasta.gz Braflo Branchiostoma floridae http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABEP02.fasta.gz Brarap Brassica rapa http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AENI01.fasta.gz Brumal Brugia malayi http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAQA01.fasta.gz Burxyl Bursaphelenchus xylophilus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CADV01.fasta.gz Caeele Caenorhabditis elegans ftp://ftp.wormbase.org/pub/wormbase/genomes/c_elegans/sequences/dna/c_elegans.WS201.dna.fa.gz Cajcaj Cajanus cajan http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFSP01.fasta.gz Caljac Callithrix jacchus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACFV01.fasta.gz Calmil Callorhinchus milii http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAVX01.fasta.gz Camfer Camelus ferus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGVR01.fasta.gz Camflo Camponotus floridanus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEAB01.fasta.gz Canalb Candida albicans http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAFO01.fasta.gz Canlup Canis lupus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAEX03.fasta.gz Cansat Cannabis sativa http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGQN01.fasta.gz Captel Capitella teleta ftp://ftp.jgi-psf.org/pub/JGI_data/Capitella/v1.0/Capitella_spl.fasta.gz Capowc Capsaspora owczarzaki http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACFS01.fasta.gz Carpap Carica papaya http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABIM01.fasta.gz Catang Catenaria anguillulae http://genome.jgi-psf.org/Catan1/download/Catan1_AssemblyScaffolds.fasta.gz Cavpor Cavia porcellus http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAKN02.fasta.gz Ceratotherium simum Cersim http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AKZM01.fasta.gz Cerzea Cercospora zeaemaydis http://genome.jgi-psf.org/Cerzm1/download/Cerzm1_AssemblyScaffolds.fasta.gz Cersub Ceriporiopsis subvermispora http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEOV01.fasta.gz Chathe Chaetomium thermophilum http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADUW01.fasta.gz Chilan Chinchilla lanigera http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGCD01.fasta.gz Chlrei Chlamvdomonas reinhardtii http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABCN01.fasta.gz Chlvar Chlorella variabilis http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADIC01.fasta.gz Chohof Choloepus hoffmanni http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABVD01.fasta.gz Chrpic Chrysemys picta bellii http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AHGY01.fasta.gz Chrysochloris asiatica Chrasi http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AMDV01.fasta.gz Cioint Ciona intestinalis http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AABS01.fasta.gz

Citlan	Citrullus lanatus	http://www.ncbi.nim.nih.gov/Traces/wgs/?download=AGCB01.fasta.gz
Citsin	Citrus sinensis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AJPS01.fasta.gz
Clagra	Cladonia grayi	http://genome.jgi-psf.org/Clagr2/download/Clagr2_genomic_scaffolds.fasta.gz
Clasph	Cladosporium sphaerospermum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AIIA01.fasta.gz
Clafus	Claviceps fusiformis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFRA01.fasta.gz
Clalus	Clavispora lusitaniae	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAFT01.fasta.gz
Closin	Clonorchis sinensis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=BADR02.fasta.gz
Cocpos	Coccidioides posadasii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACFW01.fasta.gz
Cocsub	Coccomyxa subellipsoidea	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGSI01.fasta.gz
Coclun	Cochliobolus lunatus	http://genome.jgi-psf.org/Coclu2/download/Coclu2_AssemblyScaffolds.fasta.g
Coerev	Coemansia reversa	http://genome.jgi-psf.org/Coere1/download/Coere1_AssemblyScaffolds.fasta.g
Colglo	Colletotrichum gloeosporioides	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ANPB01.fasta.gz
Concri	Condylura cristata	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AJFV01.fasta.gz
Concor	Conidiobolus coronatus	http://genome.jgi-psf.org/Conco1/download/Conco1_AssemblyScaffolds.fasta.
Conput	Coniophora puteana	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEIT01.fasta.gz
Conapo	Coniosporium apollinis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AJKL01.fasta.gz
Copcin	Coprinopsis cinerea	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AACS02.fasta.gz
Cormil	Cordyceps militaris	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEVU01.fasta.gz
Cragig	Crassostrea gigas	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFTI01.fasta.gz
Crigri	Cricetulus griseus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFTD01.fasta.gz
Croque	Cronartium quercuum	http://genome.jgi-psf.org/Croqu1/download/Croqu1_AssemblyScaffolds.fasta.g
Crypa2	Cryphonectria parasitica	http://genome.jgi-psf.org/Crypa2/download/Cryphonectria_parasiticav2.nuclea
Crygat	Cryptococcus gattii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAFP01.fasta.gz
Crypal	Cryptosporidium parvum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAEE01.fasta.gz
Cucsat	Cucumis sativus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACHR01.fasta.gz
Culqui	Culex quinquefasciatus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAAB01.fasta.gz
Cyamer	Cyanidioschyzon merolae	http://www.ebi.ac.uk/ena/data/view/AP006483.2,AP006492.2,AP006493.2,AP
Cybjad	Cyberlindnera jadinii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=BAEL01.fasta.gz
Dac_sp	Dacryopinax sp	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEUS01.fasta.gz
Dalesc	Daldinia eschscholzii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AIID01.fasta.gz
Danple	Danaus plexippus	ftp://ftp.ensemblgenomes.org/pub/metazoa/release-16/fasta/danaus_plexippu
Danrer	Danio rerio	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CABZ01.fasta.gz
Dappul	Daphnia pulex	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACJG01.fasta.gz
Dasnov	Dasypus novemcinctus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAGV03.fasta.gz
Daumad	Daubentonia madagascariensis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGTM01.fasta.gz
Debhan	Debaryomyces hansenii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AHBE01.fasta.gz
Dekbru	Dekkera bruxellensis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AHIQ01.fasta.gz
Diclab	Dicentrarchus labrax	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CABK01.fasta.gz
Dicsqu	Dichomitus squalens	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEID01.fasta.gz
Dicdis	Dictyostelium discoideum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAFI02.fasta.gz
Didexi	Didymella exigua	http://genome.jgi-psf.org/Didex1/download/Didex1_AssemblyScaffolds.fasta.g
Dipord	Dipodomys ordii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABRO01.fasta.gz
Dromel	Drosophila melanogaster	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AABU01.fasta.gz

download=AFRA01.fasta.gz ?download=AAFT01.fasta.gz ?download=BADR02.fasta.gz ?download=ACFW01.fasta.gz ?download=AGSI01.fasta.gz ad/Coclu2_AssemblyScaffolds.fasta.gz ad/Coere1_AssemblyScaffolds.fasta.gz ?download=ANPB01.fasta.gz ?download=AJFV01.fasta.gz ad/Conco1_AssemblyScaffolds.fasta.gz ?download=AEIT01.fasta.gz ?download=AJKL01.fasta.gz ?download=AACS02.fasta.gz ?download=AEVU01.fasta.gz ?download=AFTI01.fasta.gz ?download=AFTD01.fasta.gz ad/Croqu1_AssemblyScaffolds.fasta.gz ad/Cryphonectria_parasiticav2.nuclearAssembly.unmasked.gz ?download=AAFP01.fasta.gz ?download=AAEE01.fasta.gz ?download=ACHR01.fasta.gz ?download=AAAB01.fasta.gz 06483.2,AP006492.2,AP006493.2,AP006494.2,AP006495.2,AP(?download=BAEL01.fasta.gz ?download=AEUS01.fasta.gz ?download=AIID01.fasta.gz .coa/release-16/fasta/danaus_plexippus/dna/Danaus_plexippus.D ?download=CABZ01.fasta.gz ?download=ACJG01.fasta.gz /?download=AAGV03.fasta.gz /?download=AGTM01.fasta.gz /?download=AHBE01.fasta.gz /?download=AHIQ01.fasta.gz /?download=CABK01.fasta.gz /?download=AEID01.fasta.gz /?download=AAFI02.fasta.gz ad/Didex1_AssemblyScaffolds.fasta.gz /?download=ABRO01.fasta.gz /?download=AABU01.fasta.gz

Echtel Echinops telfairi Ectsil Ectocarpus siliculosus Edhaed Edhazardia aedis Eleedw Elephantulus edwardii Emihux Emiliania huxleyi Encephalitozoon cuniculi Enccun Enthis Entamoeba histolytica Entbie Enterocytozoon bieneusi Epityp Epichloe typhina Eptfus Eptesicus fuscus Equcab Equus caballus Erieur Erinaceus europaeus Erypis Erysiphe pisi Euccam Eucalyptus camaldulensis Eurher Eurotium herbariorum Eutpar Eutrema parvulum Exoder Exophiala dermatitidis Felcat Felis catus Fibrad Fibroporia radiculosa Ficalb Ficedula albicollis Fommed Fomitiporia mediterranea Fompin Fomitopsis pinicola Fraves Fragaria vesca Fracyl Fragilariopsis cylindrus Fusoxy Fusarium oxysporum Gadmor Gadus morhua Gaegra Gaeumannomyces graminis Galmar Galerina marginata Galgal Gallus gallus Ganluc Ganoderma lucidum Gasacu Gasterosteus aculeatus Geodes Geomyces destructans Geofor Geospiza fortis Gialam Giardia lamblia Gibmon Gibberella moniliformis Glaloz Glarea lozovensis Glotra Gloeophyllum trabeum Glogra Glomerella graminicola Glymax Glycine max Gonpro Gonapodya prolifera Gorgor Gorilla gorilla Gosrai Gossypium raimondii Grocla Grosmannia clavigera

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Guithe	Guillardia theta
Gymlux	Gymnopus luxurians
Hamtva	Hamiltosporidium tvaerminnensis
Hamham	Hammondia hammondi
Hanval	Hanseniaspora valbyensis
Hapbur	Haplochromis burtoni
Harsal	Harpegnathos saltator
Hebcyl	Hebeloma cylindrosporum
Helmel	Heliconius melpomene
Helrob	Helobdella robusta
Hetirr	Heterobasidion irregulare
Hetgla	Heterocephalus glaber
Hetgly	Heterodera glycines
Hetbac	Heterorhabditis bacteriophora
Homsap	Homo sapiens
Hompol	Homoloaphlyctis polyrhiza
Horvul	Hordeum vulgare
Hyaara	Hyaloperonospora arabidopsidis
Hydpin	Hydnomerulius pinastri
Hydmag	Hydra magnipapillata
Hypcat	Hyphochytrium catenoides
Hypsub	Hypholoma sublateritium
Hypbur	Hyphopichia burtonii
Hyp_sp	Hypoxylon sp
Ichmul	Ichthyophthirius multifiliis
Ixosca	Ixodes scapularis
Jaaarg	Jaapia argillacea
Jacjac	Jaculus jaculus
Jatcur	Jatropha curcas
Klumar	Kluyveromyces marxianus
Kompas	Komagataella pastoris
Labfue	Labeotropheus fuelleborni
Lacbic	Laccaria bicolor
Lacklu	Lachancea kluyveri
Lacsat	Lactuca sativa
Latcha	Latimeria chalumnae
Leeper	Leersia perrieri
Leibra	Leishmania braziliensis
Lepsal	Lepeophtheirus salmonis
Lepocu	Lepisosteus oculatus
Leueri	Leucoraja erinacea
Linhum	Linepithema humile
Linusi	Linum usitatissimum

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Lipsta Lipomyces starkeyi Loaloa Loa loa Lodelo Lodderomyces elongisporus Lotgig Lottia gigantea Lotjap Lotus japonicus Loxafr Loxodonta africana Lutlon Lutzomyia longipalpis Lytvar Lytechinus variegatus Macfas Macaca fascicularis Macpha Macrophomina phaseolina Maceug Macropus eugenii Magnaporthe oryzae Magory Malglo Malassezia globosa Malus domestica Maldom Mansex Manduca sexta Manesc Manihot esculenta Marbru Marssonina brunnea Maydes Mayetiola destructor Mayzeb Maylandia zebra Mchcon Mchenga conophoros Medtru Medicago truncatula Megrot Megachile rotundata Mellar Melampsora larici Mel sp Melanconium sp Melanochromis auratus Melaur Melgal Meleagris gallopavo Melinc Meloidogyne incognita Melund Melopsittacus undulatus Menmol Mengenilla moldrzyki Metarhizium acridum Metacr Metocc Metaseiulus occidentalis Metfru Metschnikowia fructicola Meygui Meyerozyma guilliermondii Microbotryum violaceum Micvio Micpus Micromonas pusilla Micoch Microtus ochrogaster Mimgut Mimulus guttatus Mixosm Mixia osmundae Mnelei Mnemiopsis leidyi Monper Moniliophthora perniciosa Mondom Monodelphis domestica Monbre Monosiga brevicollis Morelo Mortierella elongata

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Mucor circinelloides Muccir Musmus Mus musculus Musacu Musa acuminata Mustela putorius Musput Myceliophthora thermophila Mycthe Mycosphaerella populicola Mycpop Myodav Myotis davidii Nadful Nadsonia fulvescens Naegleria gruberi Naegru Nannochloropsis gaditana Nangad Nasonia vitripennis Nasvit Naucas Naumovozyma castellii Nechae Nectria haematococca Nematocida parisii Nempar Nemvec Nematostella vectensis Neolamprologus brichardi Neobri Neosartorya fischeri Neofis Neotyphodium gansuense Neogan Neucra Neurospora crassa Nomleu Nomascus leucogenys Nosema ceranae Noscer Ochotona princeps Ochpri Octdeg Octodon degus Ogataea parapolymorpha Ogapar Oidmai Oidiodendron maius Oikdio Oikopleura dioica Ompole **Omphalotus** olearius Onchocerca volvulus Oncvol Ophiognomonia clavigignenti Ophcla Orenil Oreochromis niloticus Ornithorhynchus anatinus Ornana Orycteropus afer Oryafe Oryctolagus cuniculus Orycun Orysat Oryza sativa Orylat Oryzias latipes Ostreococcus lucimarinus Ostluc Otolemur garnettii Otogar **Ovis** aries Oviari Oxytricha trifallax Oxytri Pactan Pachysolen tannophilus Pan troglodytes Pantro Papio anubis Papanu Paracoccidioides brasiliensis Parbra

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Partet Paramecium tetraurelia Patmin Patiria miniata Paxrub Paxillus rubicundulus Pedhum Pediculus humanus Pelsin Pelodiscus sinensis Pendig Penicillium digitatum Peripo Periglandula ipomoeae Permar Perkinsus marinus Petmar Petromyzon marinus Phatri Phaeodactylum tricornutum Phanod Phaeosphaeria nodorum Phacar Phanerochaete carnosa Phlbre Phlebia brevispora Phlgig Phlebiopsis gigantea Phlpap Phlebotomus papatasi Phodac Phoenix dactylifera Phybla Phycomyces blakesleeanus Phypat Physcomitrella patens Phyinf Phytophthora infestans Pickud Pichia kudriavzevii Piloderma croceum Pilcro Pinfuc Pinctada fucata Pirind Piriformospora indica Pir sp Piromyces sp Pismic Pisolithus microcarpus Plaviv Plasmodium vivax Pleost Pleurotus ostreatus Plicri Plicaturopsis crispa Pogbar Pogonomyrmex barbatus Polpal Polysphondylium pallidum Ponabe Pongo abelii Poptri Populus trichocarpa Postia placenta Pospla Pripac Pristionchus pacificus Procap Procavia capensis Pruper Prunus persica Psemul Pseudo-nitzschia multiseries Psecub Pseudoperonospora cubensis Ptevam Pteropus vampyrus Pucgra Puccinia graminis Punstr Punctularia strigosozonata Pundamilia nyererei Punnye Pyrter Pyrenophora teres

http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CAAL01.fasta.gz ftp://ftp.ncbi.nlm.nih.gov/genbank/genomes/Eukaryotes/invertebrates/Patiria_miniata/Pmin_1.0/Primary_, http://genome.jgi-psf.org/Paxru1/download/Paxru1_AssemblyScaffolds.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAZO01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGCU01.fsa.1.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AKCT01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFRD01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAXJ01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEFG01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABQD01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAGI01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEHB01.fasta.gz http://genome.jgi-psf.org/Phlbr1/download/Phlbr1_AssemblyScaffolds.fasta.gz http://genome.jgi-psf.org/Phlgi1/download/Phlgi1_AssemblyScaffolds.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AJVK01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACYX02.fasta.gz http://genome.jgi-psf.org/Phybl2/download/Phycomyces_blakesleeanus_v2_scaffolds.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABEU01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AATU01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ALNQ01.fasta.gz http://genome.jgi-psf.org/Pilcr1/download/Pilcr1_AssemblyScatfolds.fasta.gz http://marinegenomics.oist.jp/genomes/download/pfu_genome1.0.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CAFZ01.fasta.gz http://genome.jgi-psf.org/PirE2_1/download/PirE2_1_AssemblyScaffolds.fasta.gz http://genome.jgi-psf.org/Pismi1/download/Pismi1_AssemblyScaffolds.fasta.gz http://www.ncbi.nim.nih.gov/Traces/wgs/?download=AAKM01.fsa http://genome.jgi-psf.org/PleosPC15_2/download/PleosPC15_2_Assembly_scaffolds.fasta.gz http://genome.jgi-psf.org/Plicr1/download/Plicr1_AssemblyScatfolds.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADIH01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADBJ01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABGA01.fsa.1.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AARH01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABWF01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABKE01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABRQ01.fsa.1.gz http://www.ncbi.nim.nih.gov/Traces/wgs/?download=AEJG01.fasta.gz http://genome.jgi-psf.org/Psemu1/download/Psemu1_AssemblyScaffolds.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AHJF01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABRP01.fsa.1.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAWC01.fasta.gz http://www.ncbi.nim.nih.gov/Traces/wgs/?download=AEGM01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFNX01.fasta.gz http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEEY01.fasta.gz

Pytult	Pythium ultimum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADOS01.fasta.gz
Pytmol	Python molurus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEQU01.fsa.1.gz
Ratnor	Rattus norvegicus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AABR06.fsa.1.gz
Rhaeso	Rhamphochromis esox	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABPN01.fasta.gz
Rhopro	Rhodnius prolixus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACPB02.fasta.gz
Rhotor	Rhodosporidium toruloides	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AJMJ01.fasta.gz
Rhoglu	Rhodotorula glutinis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEVR01.fasta.gz
Riccom	Ricinus communis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AASG02.fasta.gz
Saccer	Saccharomyces cerevisiae	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACVY01.fasta.gz
Sackow	Saccoglossus kowalevskii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACQM01.fasta.gz
Saibol	Saimiri boliviensis	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGCE01.fsa.1.gz
Saicom	Saitoella complicata	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=BACD01.fasta.gz
Salsal	Salmo salar	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGKD01.fasta.gz
Sal_sp	Salpingoeca sp	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACSY01.fasta.gz
Sappar	Saprolegnia parasitica	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADCG01.fasta.gz
Sarhar	Sarcophilus harrisii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEFK01.fsa.1.gz
Schman	Schistosoma mansoni	ftp://ftp.sanger.ac.uk/pub/pathogens/Schistosoma/mansoni/genome/Assembly-v5/sma_v5.0.chr.fa.gz
Schcom	Schizophyllum commune	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADMJ01.fasta.gz
Schjap	Schizosaccharomyces japonicus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AATM01.fasta.gz
Schmed	Schmidtea mediterranea	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAWT01.fasta.gz
Sclcit	Scleroderma citrinum	http://genome.jgi-psf.org/Sclci1/download/Sclci1_AssemblyScaffolds.fasta.gz
Sclscl	Sclerotinia sclerotiorum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AAGT01.fasta.gz
Sebver	Sebacina vermifera	http://genome.jgi-psf.org/Sebve1/download/Sebve1_AssemblyScaffolds.fasta.gz
Selmoe	Selaginella moellendorffii	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADFJ01.fasta.gz
Serlac	Serpula lacrymans	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEQB01.fasta.gz
Setita	Setaria italica	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGNK01.fasta.gz
Settur	Setosphaeria turcica	http://genome.jgi-psf.org/Settu1/download/Settu1_AssemblyScaffolds.fasta.gz
Sollyc	Solanum lycopersicum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEKE02.fasta.gz
Solinv	Solenopsis invicta	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEAQ01.fasta.gz
Sormac	Sordaria macrospora	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CABT02.fasta.gz
Sorara	Sorex araneus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AALT02.fsa.1.gz
Sorbic	Sorghum bicolor	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ABXC01.fasta.gz
Spapas	Spathaspora passalidarum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEIK01.fasta.gz
Spetri	Spermophilus tridecemlineatus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AGTP01.fsa.1.gz
Sphste	Sphaerobolus stellatus	http://genome.jgi-psf.org/Sphst1/download/Sphst1_AssemblyScaffolds.fasta.gz
Spipun	Spizellomyces punctatus	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ACOE01.fasta.gz
Sporos	Sporobolomyces roseus	ftp://ftp.jgi-psf.org/pub/JGI_data/Sporobolomyces_roseus/assembly/v1.0/Sporobolomyces_roseus.allmas
Stehir	Stereum hirsutum	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AEGX01.fasta.gz
Strmar	Strigamia maritima	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=AFFK01.fasta.gz
Strpur	Strongylocentrotus purpuratus	ftp://ttp.ncbi.nlm.nih.gov/genbank/genomes/Eukaryotes/invertebrates/Strongylocentrotus_purpuratus/Spu
Strrat	Strongyloides ratti	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=CACX01.fasta.gz
Stylem	Stylonychia lemnae	http://www.ncbi.nlm.nih.gov/Traces/wgs/?download=ADNZ01.fasta.gz
Suilut	Suillus luteus	http://genome.jgi-psf.org/Suilu1/download/Suilu1_AssemblyScaffolds.fasta.gz

Susscr Sus scrofa Taegut Taeniopygia guttata Takrub Takifugu rubripes Talmar Talaromyces marneffei Tapdef Taphrina deformans Tarsyr Tarsius syrichta Tetthe Tetrahymena thermophila Teturt Tetranychus urticae Tetnig Tetraodon nigroviridis Thapse Thalassiosira pseudonana Thepar Theileria parva Thecac Theobroma cacao Thelan Thermomyces lanuginosus Thiter Thielavia terrestris Toxgon Toxoplasma gondii Traver Trametes versicolor Tremes Tremella mesenterica Tricas Tribolium castaneum Triman Trichechus manatus Trispi Trichinella spiralis Triree Trichoderma reesei Trivag Trichomonas vaginalis Trirub Trichophyton rubrum Triadh Trichoplax adhaerens Triasa Trichosporon asahii Triaes Triticum aestivum Trycru Trypanosoma cruzi Tubmel Tuber melanosporum Tulcal Tulasnella calospora Tupaia belangeri Tupbel Turtru Tursiops truncatus Uncree Uncinocarpus reesii Ustilago hordei Usthor Vanpol Vanderwaltozyma polyspora Vardes Varroa destructor Vavcul Vavraia culicis Verdah Verticillium dahliae Vicpac Vicugna pacos Vitvin Vitis vinifera Vitcor Vittaforma corneae Volcar Volvox carteri Walseb Wallemia sebi Wicano Wickerhamomyces anomalus

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Wolcoc	Wolfiporia cocos
Wucban	Wuchereria bancrofti
Xanpar	Xanthoria parietina
Xentro	Xenopus tropicalis
Xipmac	Xiphophorus maculatus
Zascel	Zasmidium cellare
Zeamay	Zea mays
Zymard	Zymoseptoria ardabiliae

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CHAPTER 3

Evolutionary and biogeographical patterns of barnacles from deep-sea hydrothermal vents

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ABSTRACT

The characterization of evolutionary and biogeographical patterns is of fundamental importance to identify factors driving biodiversity. Due to their widespread but discontinuous distribution, deep-sea hydrothermal vent barnacles represent an excellent model for testing biogeographic hypotheses regarding the origin, dispersal, and diversity of modern vent fauna. Here we characterize the global genetic diversity of vent barnacles to infer their time of radiation, place of origin, mode of dispersal, and diversification. Our approach was to target a suite of multiple loci in samples representing seven out of the eight described genera. We also performed restriction-site associated DNA sequencing on individuals from each species. Phylogenetic inferences and topology hypothesis tests indicate that vent barnacles have colonized deep-sea hydrothermal vents at least twice in history. Consistent with preliminary estimates, we find a likely radiation of barnacles in vent ecosystems during the Cenozoic. Our analyses suggest that the western Pacific was the place of origin of the major vent barnacle lineage, followed by circumglobal colonization eastward through the southern hemisphere during the Neogene. The inferred time of radiation rejects classic hypotheses of antiquity of vent taxa. The timing and the mode of origin, radiation and dispersal are consistent with recent inferences made for other deep-sea taxa, including non-vent species, and are correlated with the occurrence of major geological events and mass extinctions. Thus, we suggest that the geological processes and dispersal mechanisms discussed here can explain current distribution patterns of many other marine taxa and have played an important role shaping deep-sea

faunal diversity. These results also constitute critical baseline data with which to assess potential effects of anthropogenic disturbances on deep-sea ecosystems.

INTRODUCTION

The characterization of evolutionary and biogeographical patterns is of fundamental importance for identifying the factors that shape the ranges of deep-sea taxa, and that ultimately drive biodiversity patterns in the ocean (McClain & Mincks 2010). This is particularly relevant in the light of the increasing interest in commercial resource extraction in the deep-sea (Thurber et al. 2014). Mining of seafloor massive sulphide deposits at deep-sea hydrothermal vent fields has become one of the main industrial targets for exploitation (Boschen et al. 2013). Understanding the biodiversity contained in these areas and its connection with the fauna found elsewhere is critical for assessing the potential impacts of exploiting these mineral resources (Van Dover 2010; Van Dover et al. 2012). Although organisms living at deep-sea hydrothermal vents have adapted to cope with natural disturbances inherent to these ephemeral habitats, the intensity and frequency at which these occur can vary greatly depending on the particular geophysical nature of each system (Baker & German 2004). Thus, disturbance from mining could have additive or synergistic effects to natural disturbances at unprecedented scales, which could potentially lead to significant losses of biodiversity (Van Dover 2010). Due to their widespread distribution (Fig. 1), vent barnacles represent an excellent model for testing hypotheses regarding the historical biogeographic patterns of origin, dispersal, and current diversity of modern deep-sea chemosynthetic fauna; therefore, barnacles hold the promise of providing critical baseline data with which to assess potential effects of anthropogenic disturbances on deep-sea ecosystems.

Barnacles (Cirripedia Burmeister, 1834) are some of the most conspicuous organisms in deep-sea hydrothermal vent ecosystems worldwide. These sessile crustaceans can be found in active vent fields in most of the major spreading ridge systems and island arcs worldwide (Fig. 1), including the Central Indian Ridge (Van Dover *et al.* 2001; Nakamura *et al.* 2012), Southwest Indian Ridge (Tao *et al.* 2011), East Scotia Ridge (Rogers *et al.* 2012), northern and southern East Pacific Rise (Newman 1979; Jones 1993), Pacific-Antarctic Ridge (Southward 2005), Izu-Ogasawara Arc (Ohno *et al.* 1996), Okinawa Trough (Ohta 1990), Mariana Trough (Hessler & Lonsdale 1991), Sangihe Talaud (Herrera *et al.* 2010; Shank *et al.* 2010), Manus Basin (Tufar 1990), Edison Seamount (Tunnicliffe & Southward 2004), North Fiji Basin (Desbruyeres *et al.* 1994), Lau Basin (Southward & Newman 1998), Kermadec Arc (Buckeridge 2000), and are likely to be present in other unexplored areas. Hydrothermal vent barnacles inhabit areas of low-temperature diffuse fluid flow. Populations can reach high densities with high

biomass at over 1500 individuals per square meters (Tunnicliffe & Southward 2004; Marsh *et al.* 2012), playing key roles in vent communities as micro-habitat engineers and funnelling the flow of energy through ecosystems from primary producers to higher trophic levels (Southward & Newman 1998; Van Dover 2002; Tunnicliffe & Southward 2004; Cubelio *et al.* 2007; Rogers *et al.* 2012; Reid *et al.* 2013).



Figure 1. Global distribution map of hydrothermal vent barnacles. Ovals indicate regions where hydrothermal vent barnacles have been described (yellow: regions sampled in this study; blue: regions not sampled in this study). Red lines indicate active tectonic margins (solid lines: spreading centers; dotted lines: subduction zones).

Hydrothermal vent barnacles are presently grouped into four families belonging to the orders Pedunculata Lamarck, 1818 (suborder Scalpellomorpha, family Eolepadidae; commonly known as stalked or gooseneck barnacles) and Sessilia Lamarck, 1818 (suborder Verucomorpha, family Neoverrucidae; suborder Brachylepadomorpha, family Neobrachylepadidae; and suborder Balanomorpha, family Chionelasmatidae; commonly known as acorn barnacles) (Newman *et al.* 2006). There are approximately 13 described vent barnacle species, with several new species awaiting description (Newman *et al.* 2006). A molecular phylogenetic study of the Cirripedia, employing nuclear ribosomal genes and the histone *H3* gene, indicates that these morphologically-based taxonomic groupings (orders) are polyphyletic and thus incongruent with evolutionary history (Pérez-Losada *et al.* 2008). These results, together with those from (Linse *et al.* 2013), also suggest that vent barnacles form a monophyletic clade that likely originated in the Cretaceous; however, the possibility of a single origin remains an open question due to the paucity of

taxonomic sampling in that study. Furthermore, the relationships among morphospecies of vent barnacles also remain unresolved due to the low variability of markers examined to date.

Many putative species of vent barnacles appear to be restricted to particular ridge systems and neighboring arc and back-arc basins, and significant population structure has also been found at these scales (Watanabe *et al.* 2005). Together these observations suggest a role of habitat discontinuity as an important mechanism of speciation. By far, the region of highest diversity of putative chemosynthetic barnacle species (measured as species richness) is the western Pacific, which is considered the center of their distribution and possible place of origin (Newman *et al.* 2006). The western Pacific is also considered a biodiversity hotspot and potential place of origin of many modern groups of terrestrial and marine organisms, including deep-sea taxa (Cairns 2007; Carpenter *et al.* 2011; Herrera *et al.* 2012). In a similar way, a recent biogeographic analyses using network theory hypothesizes a possible ancestral position of the western Pacific for modern fauna associated with hydrothermal vents, having exclusive edge connections (indicating faunal similarity possible exchange paths) with the Northeast Pacific, the East Pacific Rise and the Indian Ocean (Moalic *et al.* 2011).

In this study, we aim to characterize the global genetic diversity, evolutionary and biogeographic history of barnacles from deep-sea hydrothermal vents. Our approach was to build on previous phylogenetic studies by significantly expanding the taxonomic sampling and number of genetic markers. We targeted one mitochondrial gene region, the cytochrome c oxidase subunit I (*coxI*), and two nuclear gene regions, the large ribosomal sub-unit 28S, and the histone *H3* gene, obtaining complete sequences for 94 individuals, representing seven out of the eight described genera, from 18 vent fields worldwide. We also performed restriction-site associated DNA sequencing (RAD-seq) on individuals from each identified species. Here we: (1) test the hypothesis of monophyly (i.e., a single evolutionary origin) of barnacles from deep-sea hydrothermal vents; (2) infer the place and time of origin and radiation of vent barnacles in geologic time; (3) infer historical patterns of dispersal and colonization of vent barnacle taxa worldwide; and (4) identify species boundaries and compare them to current morphospecies hypotheses.

METHODS

Morphological identifications were performed on 94 barnacle specimens (Table S1) from deep-sea hydrothermal vents using stereo-microscopy and species descriptions as references. Individuals were collected from the Central Indian Ridge, East Pacific Rise, southern East Pacific Rise, Southwest Indian

Ridge, East Scotia Ridge, Mariana Trough, the Kermadec Arc, Lau Basin, Tonga Arc, Manus Basin, Izu-Ogasawara (Bonin) Arc, and the Okinawa Trough.

Partial DNA sequences of one mitochondrial (cytochrome c oxidase subunit I) and two nuclear markers (histone *H3* gene and the ribosomal large sub-unit 28S) were generated for each individual. Additional sequences from the Superorder Thoracica Darwin, 1854 were retrieved from GenBank (http://www.ncbi.nlm.nih.gov/genbank/) and included in the analyses (Table S2).

Restriction-site Associated DNA sequencing (RAD-seq) (Baird *et al.* 2008) was performed on selected individuals from each morphospecies (Table S1) to obtain a genome-wide set of markers that could be used to infer a robust backbone of the vent barnacle phylogenetic tree, and to compare to topologies obtained from species-tree analyses of traditional Sanger-based markers.

Molecular laboratory methods

Total genomic DNA was extracted from tissue samples by: (1) digesting the tissue in 2 % CTAB buffer (Teknova) with proteinase K and RNAse A/T1 (Fermentas) for 1 hour, (2) separating nucleic acids with chloroform: isoamyl alcohol (24:1) (Fermentas) and phenol: chloroform: isoamyl alcohol (25:24:1, Tris buffered at pH 8.0) (Fermentas), (3) precipitating nucleic acids with 100% ethanol (1:1), and (4) washing the precipitate twice with 70% ethanol. Polymerase chain reactions of traditional Sanger-based markers were prepared to a final volume of 25 μ l (1 μ l of template) resulting in the following final concentrations of reagents and enzymes: 1 X GoTaq Flexi Buffer (Promega), 2.5 X BSA, 1.0 mM dNTPs (0.25 mM each), 2.0 mM MgCl₂, 1 U Taq polymerase (GoTaq, Promega), and 0.3 µM of each primer. Primer pairs used for amplifications were: 28SF 330 5'- CGTGAAGCTGCCAVTATGG-3' (designed in this study) & 28S B (Whiting 2002) for 28S, H3F & H3R (Colgan et al. 1998) for H3, and LCO1490 & HC02198 (Folmer et al. 1994) for coxI. Negative controls were included in every experiment to test for contamination. The reactions were carried out with an initial denaturation step of 5 min at 94 °C, 32 cycles (35 for coxI) of 60 s at 94 °C, 90 s at 48 °C, and 90 s at 72 °C, and a final elongation step of 10 min at 72 °C. PCR products were cleaned using the MinElute PCR Purification Kit (Qiagen) following manufacturer's protocols. Cycle-sequencing reactions were performed using the ABI BigDye Terminator v3.1 kit (Life Technologies Corp.), following manufacturer protocols. Subsequent purification was done through isopropanol precipitation. Automated sequencing was completed using a 3730xl DNA analyzer (Life technologies Corp.) at the Josephine Bay Paul Center of the Marine Biological Laboratory. Complementary chromatograms were assembled and edited using Geneious v6.1.6 (Drummond et al. 2011).

Concentration-normalized genomic DNA was submitted to Floragenex Inc. (Eugene, OR) for library preparation and RAD sequencing. Individual libraries were produced from DNA digested with a high-fidelity SbfI restriction enzyme, which is predicted to cut approximately between 5,000 and 15,000 times in the genome of a thoracican barnacle (Table S3). This predicted range was obtained using the observed frequency of the SbfI recognition sequence, and its probability using a trinucleotide composition model, in the genome of the crustacean *Daphnia pulex* (Herrera *et al.* Chapter 2). Ranges of genome size for barnacles were obtained from the Animal Genome Size Database (http://ww.genomesize.com). Barcode tags were 10-base pairs long. Libraries were sequenced by 96-multiplex on a single lane of an Illumina Hi-Seq 2000 sequencer.

Alignments, saturation analysis and model selection

Each set of sequences for Sanger-based markers was aligned independently using MAFFT (Katoh *et al.* 2002), employing the G-INS-i and Q-INS-i algorithms (gap opening penalty= 1.53, offset value= 0.07) for protein coding and ribosomal regions, respectively. To correct possible mistakes, all alignments of protein coding sequences were visually inspected and translated to amino acids in Geneious v6.1.6 (Drummond *et al.* 2011). No unusual stop codons, misplaced reading frames or suspicious substitutions were identified, indicating amplification of nuclear pseudogenes was unlikely (Lopez *et al.* 1994; Bensasson *et al.* 2001). Possible substitution saturation in the DNA sequences was evaluated by implementing the Xia test (Xia *et al.* 2003), as implemented in DAMBE v5.3.48 (Xia 2013), and by plotting genetic distances (K80 model) against the number of transitions and transversions. Saturation in codon partitions was also evaluated for each coding region.

Phylogenetic Inferences

Non-saturated datasets from individual Sanger-based markers were analyzed in RAxML-HPC2 v8.0 (Stamatakis 2006), as implemented in the CIPRES Science Gateway v3.3 (http://www.phylo.org), for a first-pass phylogenetic inference using the maximum likelihood optimality criterion. Branch support was assessed by 500 bootstrap replicates. A Thoracica-wide concatenated dataset was also analysed in this program. Only outgroups with data for at least two of the three markers were included in the concatenated dataset. Phylogenetic estimation through Bayesian inference was performed with these datasets in MrBayes v3.2.2 (Ronquist *et al.* 2012), as implemented in the CIPRES Science Gateway v3.3. Models of nucleotide substitution were selected for each non-saturated gene region using JModeltest v2.0 (Darriba *et al.* 2012), following the Bayesian Information Criterion (Table S4). Four independent analyses of 200 million Markov Chain Monte Carlo (MCMC) generations (4 chains) were run with a sampling frequency

of 20 thousand generations (burn-in = 25%). Combined analyses were performed with explicit character partitions for each concatenated region, along with their independently selected models of evolution. State frequencies were allowed to vary under a flat Dirichlet prior distribution to account for the rate variation among partitions. Nucleotide frequencies, substitution rates, gamma shape and invariant site proportion parameters were unlinked across partitions. Default prior distribution settings were assumed for all other parameters. MCMC runs were analyzed with the programs Tracer v1.5 (Rambaut & Drummond 2007) and AWTY (http://ceb.csit.fsu.edu/awty) (Nylander *et al.* 2008). Convergence among independent runs was supported by observed values of standard deviation of partition frequencies (<0.01), potential scale reduction factors (PSRF) (ca. 1.00), and effective sample sizes (EES) (>200), in addition to high correlations between runs and the flat shapes of the stationary posterior distribution traces of each parameter.

Topological Hypothesis Testing

To test the hypothesis that barnacles from deep-sea hydrothermal vents form a monophyletic group, we performed a Bayes factor comparison (Kass & Raftery 1995) between this topological hypothesis and the alternative hypothesis of non-monophyly of the group using the Thoracica-wide concatenated dataset. The marginal likelihood for each topology model was estimated through the stepping-stone method (Fan *et al.* 2011; Xie *et al.* 2011) in MrBayes using 50 steps. The estimation was performed in two independent runs of 100 million generations, with a diagnostic frequency of 1 million generations, for each topology model. All other parameters were set to default. Convergence among runs was diagnosed by the standard deviation of partition frequencies (<0.01).

Divergence Time Estimations

Time-calibration of the phylogenetic hypothesis was carried out through a Bayesian-MCMC joint estimation of phylogeny and divergence times in BEAST v1.7.5 (Drummond *et al.* 2012), as implemented in the CIPRES Science Gateway v3.3, using the Thoracica-wide concatenated Sanger-based markers dataset. Variation in mutation rates among branches was allowed by assuming an uncorrelated relaxed lognormal molecular clock model. The Yule constant speciation rate model and no extinction (Yule 1925), the Birth-Death constant speciation and extinction rates model (Gernhard 2008), and the Birth-Death constant speciation and extinction rates with incomplete taxonomic sampling model (Stadler 2009) were tested as tree priors. Unlinked character partitions were set for each concatenated region, along with their independently selected models of evolution. Three fossil calibration points (C1, C2, and C7) were selected from the studies by Pérez-Losada *et al.* (2008) and Linse *et al.* (2013) based on well-supported topological congruencies with our phylogenetic hypothesis. Fossil ages were used as lower boundary

constraints assuming prior exponential distributions with mean values of 25 my. Default prior distribution settings were assumed for all other parameters. Three independent MCMC analyses were run for 200 million generations with a sampling frequency of 20 thousand. Convergence diagnostics were examined for the combined runs in Tracer as mentioned above. Most probable trees, after 25% burn-in, were summarized into a maximum clade credibility tree with median node heights using TreeAnnotator v1.7.1 (Drummond *et al.* 2012).

Historical Biogeography

To infer historical patterns of dispersal in deep-sea hydrothermal vent barnacle lineages, we performed a Bayesian reconstruction of discrete character states of geographic location for ancestral nodes (Lemey *et al.* 2009) using BEAST v1.7.5 (Heled & Drummond 2010). In this framework, the geographical sampling locations were mapped to the time-scaled phylogenetic tree. Parameters for tree inference were as described above.

Species Delimitation

To identify species boundaries for vent barnacles in Clade A (see Results section), we employed generalized mixed Yule-coalescent (GMYC) likelihood method (Pons *et al.* 2006; Monaghan *et al.* 2009; Fujisawa & Barraclough 2013), with a single threshold, as implemented in the SPLITS R-package (available from http://r-forge.r-project.org/ projects/splits/). This method estimates species boundaries by identifying increases in branching rates that are characteristic of transition points between interspecific speciation–extinction processes and intraspecific coalescent processes.

Species Tree Estimation

Bayesian analyses of species-trees estimation for vent barnacle species identified in Clade A (see Results section) were carried out using data from the Sanger-based markers in the program *BEAST v1.7.5 (Heled & Drummond 2010). This approach was employed to take into account evolutionary coalescent processes and gene tree heterogeneity, and to evaluate the effects of gene-concatenation on the phylogenetic inference (Brito & Edwards 2008; Edwards 2008). Species were defined after the species delimitation analyses. Unlinked character, clock, and tree partitions were set for each marker, along with their independently selected models of evolution. We assumed a piecewise linear and constant root population size model. Other parameters for tree inference were as described above.



Figure 2. Maximum clade credibility ultrametric time-scaled tree, generated under the Birth-Death model tree prior, for the Thoracica-wide concatenated dataset. Red square indicates hydrothermal vent Clade A. Yellow square

indicates hydrothermal vent Clade B. Node bars represent the 95% highest posterior density intervals. Branch labels show posterior probabilities. Blue circles in nodes indicate fossil calibration points as in (Pérez-Losada *et al.* 2008; Linse *et al.* 2013). Suborders belonging to the order Pedunculata (stalked or gooseneck barnacles) are indicated with (P). Suborders belonging to the order Sessilia (acorn barnacles) are indicated with (S). *Indicates data generated in this study.

RAD-seq data quality control and loci clustering

Sequence reads were de-multiplexed and quality filtered with the process_radtags program from the package Stacks v1.19 (Catchen *et al.* 2011; Catchen *et al.* 2013). Barcodes and Illumina adapters were excluded from each read and length was truncated to 90bp (-t 90) Reads with ambiguous bases were discarded (-c). Reads with an average quality score below 10 (-s 10) within a sliding window of 15% of the read length (-w 0.15) were discarded (-r). The rescue barcodes and RAD-tags algorithm was enabled (-r). Additional filtering, and the clustering within and between individuals to identify loci was performed using the program *py*RAD v2.01 (Eaton 2014). Reads with more than 33 bases with a quality score below 20 were also discarded. The minimum depth of coverage required to build a cluster was 5 (d 5). As in Hipp *et al.* (2014), three different clustering thresholds were explored (c 0.80, 0.85 and 0.90). Similarly, four different values for the minimum taxon coverage in a given locus were explored (m 4, 6, 8, and 10). The maximum number of shared polymorphic sites in a locus was set to 3 (p 3). Loci were concatenated into combined RAD-seq matrices.

RAD Phylogenetics

Phylogenetic inferences of RAD-seq matrices, built with *py*RAD under each combination of clustering threshold and minimum taxon coverage parameters (as outlined above), were carried out in RAxML-HPC2 v8.0. We assumed a generalized time-reversible DNA substitution model with a gamma-distributed rate variation across sites. Branch support was assessed by 500 bootstrap replicates.

RESULTS

Complete Sanger-based marker datasets were obtained for all 94 individuals, except for 2 specimens of *Vulcanolepas osheai*. Sequences are stored in the GenBank database of the U.S. National Center for Biotechnology Information. Approximate sequence lengths for each marker were 700 bp for 28S, 657 bp for *coxI*, and 327 bp for *H3*. Xia tests indicated substantial saturation at the Thoracica-wide level at third codon positions of *coxI* (Table S5). Little saturation was found in all other partitions. Maximum likelihood and Bayesian phylogenetic inferences from each Sanger-based marker produced mostly congruent trees that varied in the degree of resolution yet all showed poorly supported branches (i.e.,
posterior probability < 80, bootstrap support < 80) (supplementary electronic material). Analyses of the Thoracica-wide concatenated dataset generated a better-supported and -resolved phylogeny overall (Fig. 2, supplementary electronic material). The topologies of these trees were congruent with previously published phylogenetic hypotheses for the Thoracica (Pérez-Losada *et al.* 2008; Linse *et al.* 2013).



Figure 3. Clade A combined 28S, *H3*, and *coxI* maximum clade credibility ultrametric time-scaled tree generated under the Birth-Death model. Branch colours show the most probable location states: western Pacific in blue, eastern Pacific in green, Southern Ocean south of the Atlantic in yellow and Indian Ocean in orange. Pie charts show the posterior probabilities of location states for each ancestral node (total pie area = 1). Branch labels show posterior probabilities. Purple vertical dashed line indicates the maximum likelihood-inferred time for the speciation-coalescent threshold for species delimitation (GYMC). Vertical dotted lines indicate important events in geologic time: Oceanic Anoxic Events (red, OAEg for global and OAEr for regional), Cretaceous-Paleogene mass extinction

(fuchsia, C-P), Paleocene-Eocene Thermal Maximum (brown, P-ETM), opening of the Drake Passage (black, DP), establishment of the Antarctic Circumpolar Current (black, ACC), disruption of the Farallon Pacific Ridge (black, FPR), formation of the East Scotia Rise (black, ESR). Geologic periods and eras are indicated with horizontal bars: upper Cretaceous (Cre (U)), Paleocene (Pal), Eocene (Eoc), Oligocene (Oli), Miocene (Mio), Pliocene (Pli), and Pleistocene (Ple).

RAD-seq datasets were obtained from 13 individuals representing the vent barnacle species in this study (Table S1). An average of 843,541 reads (SD 589,377) were obtained per individual. Reads are stored at the Sequence Read Archive (SRA) of NCBI. The great variability in sequencing yield was largely a product of varying of DNA integrity as some samples had notably degraded DNA (Table S6), as determined by agarose gel electrophoresis. An average of 712,306 reads per individual (SD 546,846), roughly 78% of all reads, were retained after quality filtering steps. In individuals with high-integrity DNA, the number of RAD-tag loci with depth of coverage greater than 4 X was approximately 18,000, per individual. This number is congruent with the expected number of RAD-tags, between 10,000 and 30,000, predicted for a barnacle, using the enzyme SbfI (Table S3). The average depth of coverage per locus was approximately 54 X (SD 13 X). As expected, the number of loci per individual was higher as the clustering threshold was larger (Table S7). Phylogenetic trees obtained from the RAD-seq datasets were completely resolved, highly supported as indicated by bootstrap resampling, and were largely congruent with the trees produced with Sanger-based data.

Phylogenetic Inferences

Analyses of Sanger-based markers revealed that barnacles from deep-sea hydrothermal vents are divided into two well-supported (posterior probability = 1, bootstrap support > 99) main clades (Fig. 2): Clade A contains the genera *Neobrachylepas* Newman & Yamaguchi, 1995 (Order Sessilia, Suborder Brachylepadomorpha), *Neoverruca* Newman, 1989 (Order Sessilia, Suborder Verrucomorpha), *Ashinkailepas* Yamaguchi *el al.* 2004 (Order Pedunculata, Suborder Scalpellomorpha), *Leucolepas* Southward & Jones, 2003 (Suborder Scalpellomorpha), *Vulcanolepas* Southward & Jones, 2003 (Suborder Scalpellomorpha), *Vulcanolepas* Southward & Jones, 2003 (Suborder Scalpellomorpha), and *Neolepas* Newman, 1979 (Suborder Scalpellomorpha); and Clade B was restricted to the genus *Eochionelasmus* Yamaguchi, 1990 (Order Sessilia, Suborder Balanomorpha). Clade A is well supported as the sister taxon to the predominantly deep-sea clade of the Scalpellidae (Pérez-Losada *et al.* 2008; Linse *et al.* 2013). Clade B *Eochionelasmus* is associated with the paraphyletic Balanomorpha group, however the lack of support and resolution within the later group prevents an unambiguous phylogenetic placement.



Figure 4. *Top.* Claudogram of the posterior distribution of species trees. High color density is indicative of areas in the species trees with high topology agreement. Different colors represent different topologies. The maximum clade

credibility species tree is shown with thicker branches. Branch labels show posterior probabilities. Trees with the same topology as the maximum clade credibility species tree are coloured in blue. Trees with different topologies are colored yellow or red. *Bottom*. Maximum likelihood phylogenetic tree inferred with RAD-seq data. The matrix used for this tree was obtained with a clustering threshold of 0.85 and minimum taxon coverage of 6. This matrix contains 828,960 nucleotide sites in 9,766 loci. 76,353 of the sites are variable and 26,955 are parsimony informative. This matrix contains 43.54% missing data. Branch labels show bootstrap support values. Scale bar indicates substitutions per site. Barnacle species images are from individuals included in this study. Species names are followed by the collection regions.

Neobrachylepas and *Neoverruca* appear as the extant representatives of the earliest divergent lineages in Clade A; however, their order of divergence is unclear due to lack of strong branch support. The rest of the genera in Clade A belong to the Family Eolepadidae. The genus *Ashinkailepas* belongs to the earliest divergent lineage in the family (Fig. 3), and contains two sub-clades, one grouping individuals from the Izu-Ogasawara (Bonin) Arc and the Okinawa Trough (identified as *Ashinkailepas seepiophila*), and the second grouping individuals from the Lau Basin, and the Kermadec Arc. The latter sub-clade includes a paratype of *A. kermadecensis*. Neither genus *Vulcanolepas* nor *Neolepas* is monophyletic. The *Vulcanolepas/Leucolepas* from the Kermadec Arc, Lau Basin, and Mariana Arc belong to lineages that appear to have diverged earlier in history with respect to a well-supported and well-resolved clade made up by *N. zevinae/rapanuii* from the East Pacific Rise and its sister sub-clade of *V. scotiaensis* from the East Scotia Ridge and *Neolepas* sp. 1 from the Southwest and Central Indian Ridge.

Topological Hypothesis Testing

None of the phylogenetic hypotheses inferred from the Thoracica-wide concatenated Sanger-based dataset support the monophyly of barnacles from deep-sea hydrothermal vents (Fig. 2). The topological test showed that the hypothesis of monophyly was significantly less probable than the hypothesis of non-monophyly (marginal log-likelihoods -16928.21 and -16908.62 respectively). The large difference in log-likelihoods (> 5) (Kass & Raftery 1995) constitutes strong contradictory evidence against the monophyly of vent barnacles as originally suggested by Pérez-Losada *et al.* (2008).

Divergence Estimates and Biogeographic History

Tree time calibrations of the combined Sanger-based dataset produced divergence estimates slightly older under the Yule tree prior of constant speciation, when compared with the nearly identical estimates obtained under the Birth-Death prior models (Fig. 2 and supplementary electronic material). These divergence estimates are consistent with estimates from Linse *et al.* (2013). The tree obtained under the Birth-Death model had the best likelihood score; however, no significant differences were encountered

among models (log-likelihood < 1). The time to the most recent common ancestor (TMRCA) of Clade A was estimated at 68.0 million years before present (Myr BP) (95% Highest Posterior Density Interval [HPD]: 38.2-105.9) under the Birth-Death models (BD) and 79.3 Myr BP (95% HPD: 47.1-121.5) under the Yule model of constant speciation rate. The TMRCAs of Eolepadidae and the *Neolepas-Vulcanolepas-Leucolepas* sub-clade were estimated at 25.1 Myr BP (95% HPD: 12.1-43.3) and 10.5 Myr BP (95% HPD: 5.4-17.3) under BD, and 31.2 Myr BP (95% HPD: 15.4-53.7) and 13.8 Myr BP (95% HPD: 7.5-23.1) under the Yule model, respectively. Divergence between Pacific and non-Pacific *Neolepas-Vulcanolepas* eolepadids was estimated to have occurred 4.8 Myr BP (95% HPD: 2.3-8.5) and 6.4 Myr BP (95% HPD: 3.0-11.2) under BD and Yule models, respectively. The split between the East Scotia Ridge and the Indian Ocean lineages occurred 1.7 Myr BP (95% HPD: 0.4-3.8) under BD and 2.3 (95% HPD: 1.1-6.7) under the Birth-Death model and 4.2 Myr BP (95% HPD: 1.3-8.8) under Yule. The analysis of historical biogeography suggest with high-probability that hydrothermal vent barnacles from Clade A originated in the western Pacific, and during the late Miocene –Pliocene colonized the Eastern Pacific, the Southern Ocean south of the Atlantic, and the Indian Ocean (Fig. 3).

Species Delimitation

GMYC analyses of Clade A identified a transition point between interspecific speciation-extinction processes and intraspecific coalescent processes at 0.6 Myr BP for the time-scaled combined Sangerbased phylogeny estimated with the Birth-Death model tree prior (Fig. 3). The GMYC model showed a significant ($\alpha = 0.05$) better fit to the data than the null model of uniform coalescent branching rates (likelihood ratio = 25.9, p < 0.0001). There were 12 distinct clusters identified, which largely corresponded to species already described or populations that were presumed to be new species. Genetic distances (*coxI* uncorrected distances) among individuals within clusters ranged between 0 and 0.9%. Genetic distances among individuals from different clusters ranged between 2 and 17.8% (except for the two *Neolepas zevinae/rapanuii* clusters whose maximum distance was 0.9%). Similarly in Clade B *Eochionelasmus* the genetic distances among individuals ranged between 0 and 0.9%.

Species Tree Estimation

The topology of the inferred Sanger-based species tree is fully congruent with the topology of the phylogenetic hypothesis obtained with the concatenated Sanger-based markers dataset, and the branch support values are mostly equal (Fig. 4). Poorly-resolved regions of the tree include the relationships among lineages of *Vulcanolepas/Leucolepas* from the Kermadec Arc, Lau Basin, Mariana Trough, and basally the positions of *Neoverruca* and *Neobrachylepas* within Clade A.

RAD Phylogenetics

RAD-seq matrices resulting from the three explored clustering thresholds (c 0.80, 0.85 and 0.90) contained similar numbers of loci and similar percentages of missing data per clustering parameter value used for the minimum taxon coverage in a given locus (approximately 15,500, 9,600, 3,800, and 600 loci, and 52 %, 44 %, 33 % and 21 % missing data, for m 4, 6, 8, and 10 respectively; see Table S8 for details). The percentages of variable sites and parsimony informative sites across matrices ranged between 6.81 -13.18 % and 2.26 - 4.22 % respectively, being higher with smaller values of clustering thresholds and larger values of minimum taxon coverage. The tree topologies obtained from phylogenetic inferences of each matrix were identical to each other (supplementary electronic material). These topologies from RAD-seq matrices were also similar to the species tree obtained with Sanger-based markers (Fig. 4), only differing in the position of Leucolepas, appearing in the RAD-based trees as sister to the clade made up by N. zevinae/rapanuii from the East Pacific Rise, V. scotiaensis from the East Scotia Ridge and Neolepas sp. 1 from the Southwest and Central Indian Ridge. RAD-based trees topologies were highly supported with bootstrap values of 100 for all branches, except for the ones from matrices generated with a minimum taxon coverage parameter of m10. In these cases, the branches supporting the clades of Vulcanolepas from the Lau Basin and the Kermadec Arc, and of Leucolepas-N. zevinae/rapanuii-V. scotiaensis- Neolepas sp. 1 have bootstrap support values greater than 94 and 71, respectively.

DISCUSSION

Are vent barnacles monophyletic?

The inferred evolutionary history of hydrothermal vent barnacles is not consistent with the hypothesis of monophyly (single ancestry) as proposed by the smaller taxon-sampling studies of Pérez-Losada *et al.* (2008) and Linse *et al.* (2013), which included only two of the four families of vent barnacles. Our analyses of a significantly expanded dataset indicate that there are two main clades (Clade A and Clade B) (Fig. 2), thus suggesting that barnacles have colonized deep-sea hydrothermal vents at least twice in history. The results from a concurrent study by Perez-Losada *et al.* (2014) provide support to this inference by placing Clade B (*Eochionelasmus* ohtai) nested within the balanomorph barnacles, although the hypothesis of monophyly of vent barnacles was not explicitly tested in that study.

Deep-sea hydrothermal vent barnacle Clade A is the more diverse of the two, containing six of the seven genera included in this study. This clade also contains a remarkable diversity of morphologies, including asymmetric (Neoverrucidae) and symmetric (Neobrachylepadidae), pedunculate (Eolepadidae) and sessile (Neoverrucidae and Neobrachylepadidae) forms (Fig. 4) (note that neoverrucid barnacles have a

pedunculated stage during early ontogenesis (Newman & Hessler 1989)). The sister relationship of Clade A and the deep-sea pedunculate Scalpellidae (Fig. 2) (Pérez-Losada et al. 2008; Linse et al. 2013) suggests that the sessile state of the Neoverrucidae and Neobrachylepadidae is a derived state. This observation is consistent with the mounting evidence that the characters used to define higher taxonomic groups in Cirripedia need to be revised in light of multilocus molecular phylogenetic hypotheses (Pérez-Losada et al. 2008; Linse et al. 2013). A noteworthy example of this taxonomic and phylogenetic incongruence is the phylogenetic placement of N. relica nested in Clade A. N. relica is the sole living brachilepadoform species and until now was considered the most 'primitive' lineage of sessilian barnacles (Newman & Yamaguchi 1995). Clade B only contains the genus Eochionelasmus. Despite its morphological and phylogenetic affinities with the Balanomorpha, the phylogenetic position of *Eochionelasmus* is this study is unstable. Similarly, Perez-Losada *et al.* (2014) found low support for the branches resolving the position Eochionelasmus ohtai within the balanomorphs. This instability is likely caused by the long branch supporting this clade, which may indicate a rapid evolutionary rate, old divergence, or taxonomic undersampling (Fig. 2, supplementary electronic material). Further taxonomic sampling of related genera and careful review of character use for systematics should help resolve its systematics.

Deep-sea hydrothermal vent environments have been characterized as being patchy and ephemeral habitats with extreme spatial and temporal gradients of temperature, reduced chemicals, oxygen and food supply (Van Dover 2000). These conditions present significant physiological and ecological challenges to organisms and act as environmental filters that promote the evolution and distribution of species with specialized adaptations (Tunnicliffe *et al.* 2003; Fisher *et al.* 2007). The widespread persistence of vent chemosynthetic environments throughout earth's geologic history (Shock *et al.* 1995) has likely been an important factor enabling the independent colonization by multiple lineages of barnacles, as well as of other taxa, e.g., mussels (Lorion *et al.* 2013) and decapods (Yang *et al.* 2013). Clade A is nested within a predominantly deep-sea clade Linse *et al.* (2013), suggesting a colonization of hydrothermal vents at depth. The nested position within Clade A of *A. seepiophila* – the only barnacle species known to live in both cold-seep and hydrothermal vent environments – indicates a single colonization of seep environments by vent ancestors. This pattern contrasts with the stepwise colonization scenario of deep-sea chemosynthetic environments, starting in organic substrates or cold-seeps and then moving to hydrothermal vents, as suggested for other taxonomic groups, e.g., mussels (Lorion *et al.* 2013).

Historical Biogeography

The most common recent ancestor of hydrothermal-vent barnacles from Clade A likely lived in the late Mesozoic or early Cenozoic. The time to the most recent common ancestor inferred in this study is consistent with the timing inferred by Linse et al. (2013), but contrasts with the lower Cretaceous origin proposed by Pérez-Losada et al. (2008) and with classic hypotheses of antiquity of vent taxa, which proposed that hydrothermal-vent barnacles were mid-Mesozoic relict taxa (Newman 1979; Newman 1985). The discrepancy with the results from Pérez-Losada et al. (2008) is due to the exclusion of fossil calibration points because of uncertainty in the phylogenetic placement as described by Linse et al. (2013). The timing of radiation of Clade A during the Cenozoic is comparable to the estimates of origin and radiation in other chemosynthetic taxa, e.g., radiation of bresiliid shrimp 6.7-11.7 Myr BP (Shank et al. 1999); origin of siboglinid tubeworms ca. 60 Myr BP (Chevaldonne et al. 2002); radiation of chemosynthetic mussels at ca. 45 Myr BP (Lorion et al. 2013); radiation of kiwaid yeti crabs starting at ca. 30 Myr BP; also see reviews by Little and Vrijenhoek (2003) and Vrijenhoek (2013). A recent origin and radiation of most modern vent taxa and many other deep-sea taxa (Little & Vrijenhoek 2003; Smith & Stockley 2005; Strugnell et al. 2008) is consistent with the inference of a major deep-sea mass extinction event during the Cretaceous-Paleogene period boundary (Raup & Sepkoski 1982; Horne 1999; Harnik et al. 2012) (see Fig. 3). Several smaller-scale extinction events linked to regional Oceanic Anoxic Events, ocean acidification and temperature changes also occurred during the Cretaceous period and at the Paleocene-Eocene epoch boundary (Jacobs & Lindberg 1998; Rogers 2000; Harnik et al. 2012).

The most probable place of origin of the modern vent barnacle lineage from Clade A is the western Pacific, as indicated and highly supported by Bayesian ancestral state reconstruction. This is also the region where the oldest lineages and the highest diversity are found. The heterogeneity of depths in hydrothermal vent systems in the western Pacific, and the close proximity to other chemosynthetic ecosystems such as cold seeps and organic enrichments, both shallow and deep, have been suggested as important factors driving the re-colonization of vent environments and subsequent diversification (Moalic *et al.* 2011). Our analyses suggest that the most probable path of dispersal out of the western Pacific was a migration eastward during the Miocene epoch, possibly following hydrothermal vent habitats along the Pacific-Antarctic Ridge, and colonization of the eastern Pacific. The neolepadids from the East Pacific Rise have a coalescence point that is posterior to the Oligocene disruption of the Pacific-Farallon Ridge by subduction under the North American Plate, *ca.* 28.5 Myr BP (Fig. 3) (Atwater 1989), which can explain why barnacles are absent from the north-eastern Pacific vents along the Juan de Fuca Ridge. A spreading through the southern hemisphere likely followed to the East Scotia Ridge and South Sandwich Arc during the late Miocene epoch, reaching the Southwest Indian Ridge and Central Indian Ridge during

the Pliocene/Pleistocene epochs. No vent barnacle species have been found at Mid Atlantic Ridge hydrothermal vents, although the southern portion of this major mid-ocean ridge remains largely unexplored.

The proposed history of dispersal is congruent with the timing of opening of the Drake Passage during the mid Eocene epoch, ca. 41 Myr BP (Scher 2006), the late Eocene establishment of the eastward-flowing Antarctic Circumpolar Current (ACC), ca. 34 Myr BP (Scher 2006), and the mid Miocene formation of the East Scotia Rise, ca. 15 Myr BP (Livermore 2003) (see Fig. 3). Hydrothermal vent yeti crabs (Decapoda: Anomura: Kiwaidae) share an almost identical pattern of historical dispersal from the eastern Pacific to the East Scotia Ridge and the Southwest Indian Ridge (see Roterman et al. (2013) for a detailed hypothesis of vicariance in this group). A likely origin in the western or northwestern Pacific followed by migration and colonization eastward throughout the southern hemisphere during the Miocene epoch has also been inferred for other non-vent deep-sea taxa such as the octocoral Paragorgia arborea (Herrera et al. 2012), and other marine taxa such as the spiny dogfish Squalus acanthias (Verissimo et al. 2010) and the bryozoan Membranipora membranacea (Schwaninger 2008). These observations provide support for the biogeographic hypothesis proposed by Moalic et al. (2011) that the western Pacific was a centre of origin of modern vent fauna from which most taxa dispersed globally. However, our data do not support the idea of direct links between the western Pacific communities and the Indian Ocean, but rather a stepping-stone mode of dispersal in the southern hemisphere following the direction of the dominant ACC. We suggest that the geological processes and dispersal mechanisms discussed here can explain the current distribution patterns of many other marine taxa and have played an important role shaping extant deep-sea faunal diversity

The history of Clade B is not well resolved. The phylogenetic hypothesis here presented suggests that the divergence of this lineage within the Balanomorpha occurred in the Mesozoic era (Fig. 2). However this inferred antiquity is likely to be an artifact caused by taxonomic undersampling in this group. Additional data from other *Echionelasmus* populations, e.g., *E. paquensis* from the eastern Pacific, as well as from confamilial species and related groups would provide greater resolution of the evolution of Clade B.

Species Delimitation and Relationships

Inferences of species boundaries in Clade A, based on the generalized mixed Yule-coalescent method, are largely congruent with descriptions of putative morphospecies. The identified species clusters are well-constrained geographically by mid-ocean spreading ridge system and neighboring island arc basins (Figs. 3 and 4). Divergences among congeners in *Ashinkailepas* and *Neoverruca* are largely consistent with the

biogeographic boundary between the northwest and southwest Pacific, inclusive of the Mariana arc, proposed by Bachraty *et al.* (2009). Relationships among *Vulcanolepas, Leucolepas and Neolepas* species clusters remain contentious due to the non-monophyly of all three genera as defined by Buckeridge *et al.* (2013) and thus require substantial revision.

There is a lack of overlap in genetic distances for the *coxI* barcode marker within and among inferred species clusters. The maximum genetic distance within species clusters of 0.9 %, and the minimum distance among species clusters of 2 % are consistent with the proposed threshold value of *ca.* 2 % to define species boundaries through DNA barcoding in Crustacea (Hebert *et al.* 2003; Lefebure *et al.* 2006). Similarly, the maximum genetic distance among individuals of *Echionelasmus ohtai* is 0.9%. The only exception to this pattern is found in the *Neolepas zevinae/rapanuii* species cluster pair, where the maximum distance between clusters is 0.9 %. There is no phylogenetic support for this split or geographic segregation between specimens from the East Pacific Rise and southern East Pacific Rise, thus suggesting that the division of *Neolepas zevinae/rapanuii* is not indicative of species-level differentiation. The barcoding gap within and among species has been consistently found in other barnacle taxa (Tsang *et al.* 2008; Tsang *et al.* 2009; Yoshida *et al.* 2011) and in crustaceans in general (Costa *et al.* 2007; Matzen da Silva *et al.* 2011), thus our *coxI* genetic distance data provides further support to the species delimitations proposed for Clade A. The species delimitation framework developed will enable rapid species assignments as specimens from newly explored geographical regions become available.

RAD phylogenetics

Several sources of uncertainty have been associated with the use of the few traditional sequence markers available for non-model organisms (e.g., mitochondrial and ribosomal genes), including low variability, biased loci sampling, poor genomic representation, low statistical power, and inclusion of pseudogenes, among others. The effects of these are often hard to identify due to the paucity of multi-locus genome-wide comparative datasets. Such problems have been recognized and accounted for in model organisms by comparing large numbers of genomic DNA sequences from various individuals and identifying thousands of variable regions across the genome (Rokas *et al.* 2003; Clark *et al.* 2007). Recent technological and methodological developments in next-generation sequencing platforms and methodologies, such as RAD-seq, have made genomic resources increasingly accessible and available for phylogenetics in non-model organisms (Wagner *et al.* 2012; Eaton & Ree 2013; Jones *et al.* 2013; Reitzel *et al.* 2013), thus offering a great opportunity to overcome the difficulties inherent to the use of traditional approaches in many taxa.

In this study, we demonstrated that RAD-seq data provide strong support for the overall evolutionary history of vent barnacles inferred with traditional Sanger-based markers, and allow the inference of a fully resolved and supported phylogenetic tree. The small difference in topology between the species tree inferred with Sanger-based markers and the RAD-seq trees does not alter any of the conclusions regarding the biogeographical history or species delimitation of vent barnacles, but does have taxonomic implications. Further sampling and a follow-up morphological taxonomic revision would be needed to clarify the validity of the currently described genera. This study demonstrates the utility of comparative Sanger-based and RAD sequencing as a means of comparative phylogenetic inference validation in poorly known taxa such as deep-sea invertebrates.

CONCLUSIONS

Phylogenetic inferences and topology tests indicate that hydrothermal vent barnacles are not a monophyletic group. The likely timing of barnacle radiation in hydrothermal vent ecosystems was during the late Cenozoic, consistent with the timing of other specific deep-sea taxa, and correlated to the occurrence of major extinction events. Our analyses suggest that the western Pacific was the place of origin of the major hydrothermal vent barnacle lineage, followed by circumglobal colonization eastward along the southern hemisphere during the Neogene period. Inferences of species boundaries based on generalized mixed Yule-coalescent methods and DNA barcoding are largely congruent with morphological descriptions of putative species. RAD-seq data provide strong support for the overall evolutionary history inferred from Sanger-based markers and a fully resolved backbone of the vent barnacle phylogenetic tree. These results also constitute critical baseline data with which to assess potential effects of anthropogenic disturbances on deep-sea ecosystems.

AUTHOR CONTRIBUTIONS

SH and TMS conceived and designed research. SH performed research. SH analyzed data. SH, TMS and HW collected and contributed samples and reagents. SH wrote the paper with comments from TMS and HW.

DATA ACCESSIBILITY

Supplementary electronic material

Nucleotide alignments, input files, are tree files are available in Dryad DOI: doi:10.5061/dryad.7kn5k

Raw data

Raw DNA sequences are available at the U.S. National Center for Biotechnology Information (NCBI) GenBank and Sequence Read Archive (SRA accession number SRP051026).

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SUPPORTING INFORMATION

Table S1. Collection and sequence information for the specimens used in this study.

Species	Region	Locality	Cruise	Collection Platform	Collection Event	Date of Collection	Depth (m)	Lat.	Long.	ID sample
Ashinkailepas seepiophila	Izu-Ogasawara Arc	Myojin Knoll	NT97-10	HOV Shinkai 2000	Dive 952	19970622	1268	32.100	139.875	AsOg1
Ashinkailepas seepiophila	Izu-Ogasawara Arc	Myojin Knoll	NT97-10	HOV Shinkai 2000	Dive 952	19970622	1268	32.100	139.875	AsOg2
Ashinkailepas seepiophila	Izu-Ogasawara Arc	Myojin Knoll	NT97-10	HOV Shinkai 2000	Dive 952	19970622	1268	32.100	139.875	AsOg3
Ashinkailepas seepiophila	Izu-Ogasawara Arc	Myojin Knoll	NT97-10	HOV Shinkai 2000	Dive 952	19970622	1268	32.100	139.875	AsOg4
Ashinkailepas seepiophila	Okinawa Trough	Iheya Ridge	NT97-14	HOV Shinkai 2000	Dive 977	19970924	1396	27.545	126.972	AsOk1
Ashinkailepas seepiophila	Okinawa Trough	Iheya Ridge	NT97-14	HOV Shinkai 2000	Dive 977	19970924	1396	27.545	126.972	AsOK2
Ashinkailepas seepiophila	Okinawa Trough	Iheya Ridge	NT97-14	HOV Shinkai 2000	Dive 977	19970924	1396	27.545	126.972	AsOK3
Ashinkailepas_kermadecensis	Kermader Arc	Wright Seamount	KOK0505	HOV Pisces V	Dive 621	20050418	1165	-31.861	-179.188	18008
Ashinkailepas_kermadecensis	Lau Basin	Niua Seamount	RR1211	ROV Quest 4000	Dive 330	20120922	723	-15.081	-173.553	AsNi1
Ashinkailepas_kermadecensis	Tonga Arc	Niua Seamount	RR1211	ROV Quest 4000	Dive 330	20120922	723	-15.081	-173.553	AsNiN1
Ashinkailepas_kermadecensis	Tonga Arc	Niua Seamount	RR1211	ROV Quest 4000	Dive 330	20120922	723	-15.081	-173.553	AsNiN2
Ashinkailepas_kermadecensis	Tonga Arc	Niua Seamount	RR1211	ROV Quest 4000	Dive 330	20120922	723	-15.081	-173.553	AsNiN3
Ashinkailepas_kermadecensis	Tonga Arc	Niua Seamount	RR1211	ROV Quest 4000	Dive 330	20120922	723	-15.081	-173.553	AsNiN4
Eochionelasmus ohtai	Lau Basin	ELSC	KM0417	Scripps-type dredge	Dredge 52	20041009	2640	-20.050	-176.134	bar11
Eochionelasmus ohtai	Lau Basin	ELSC	KM0417	Scripps-type dredge	Dredge 52	20041009	2640	-20.050	-176.134	bar13
Eochionelasmus ohtai	Lau Basin	ELSC	KM0417	Scripps-type dredge	Dredge 52	20041009	2640	-20.050	-176.134	bar14
Eochionelasmus ohtai	Lau Basin	ELSC	TN236	ROV Jason 2	Dive 444	20090627	2232	-20.682	-176.183	bar20
Eochionelasmus ohtai	Lau Basin	ELSC	TN236	ROV Jason 2	Dive 444	20090627	2232	-20.682	-176.183	bar21
Eochionelasmus ohtai	Lau Basin	Fonualei South Volcano	RR1211	ROV Quest 4000	Dive 323	20120913	956	-17.542	-174,576	EoFoS1
Eochionelasmus ohtai	Lau Basin	Fonualei South Volcano	RR1211	ROV Quest 4000	Dive 323	20120913	956	-17.542	-174,576	EoFoS2
Eochionelasmus ohtai	Lau Basin	Fonualei South Volcano	RR1211	ROV Quest 4000	Dive 323	20120913	956	-17.542	-174,576	EoFoS3
Eochionelasmus ohtai	Lau Basin	NELSC	RR1211	ROV Quest 4000	Dive 323	20120913	956	-17.542	-174.576	EoFoS4
Eochionelasmus ohtai	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1617	-15.383	-174.245	EoNE1
Eochionelasmus ohtai	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1617	-15.383	-174.245	EoNE3
Eochionelasmus ohtai	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1617	-15.383	-174.245	EoNE4
Eochionelasmus ohtai	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	EoMaU1
Eochionelasmus ohtai	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	EoMaU2
Eochionelasmus ohtai	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173,788	EoMaU3
Eochionelasmus ohtai	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	EoMaU4
Eochionelasmus ohtai manusensis	Manus Basin	Vienna Woods	YK06-13	HOV Shinkai 6500	Dive 980	20060918	2477	-3.163	150.279	EoMa1
Eochionelasmus ohtai manusensis	Manus Basin	Vienna Woods	YK06-13	HOV Shinkai 6500	Dive 980	20060918	2477	-3.163	150.279	EoMa2
Eochionelasmus ohtai manusensis	Manus Basin	Vienna Woods	YK06-13	HOV Shinkai 6500	Dive 980	20060918	2477	-3.163	150.279	EoMa3
Eochionelasmus ohtai manusensis	Manus Basin	Vienna Woods	YK06-13	HOV Shinkai 6500	Dive 980	20060918	2477	-3.163	150.279	EoMa4
Leucolepas longa	Mariana Arc	TOTO Caldera	KM0912	HROV Nereus	Dive 015	20090604	2949	12.711	143,543	har01
Leucolepas longa	Mariana Arc	TOTO Caldera	KM0912	HROV Nereus	Dive 015	20090604	2949	12.711	143.543	har02
Leucolepas longa	Mariana Arc	TOTO Caldera	KM0912	HROV Nereus	Dive 015	20090604	2949	12.711	143.543	VuTO1

Leucolepas longa	Mariana Arc	TOTO Caldera	KM0912	HROV Nereus	Dive 015	20090604	2949	12.711	143.543	VuTO2
Leucolepas longa	Mariana Arc	TOTO Caldera	KM0912	HROV Nereus	Dive 015	20090604	2949	12.711	143.543	VuTO3
Leucolepas longa	Mariana Arc	TOTO Caldera	KM0912	HROV Nereus	Dive 015	20090604	2949	12.711	143.543	VuTO4
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	bar03
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	bar04
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	bar05
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	bar06
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	bar07
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	bar08
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	EoNE2
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NeNE1
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NeNE2
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NeNE5
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NeNE6
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NeNE7
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NeNE8
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NerNE3
Neobrachylepas relica	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1618	-15.383	-174.245	NerNE4
Neolepas sp. 1	CIR	Kairei Field	YK09-13	HOV Shinkai 6500	Dive 1175	20091113	2422	-25.320	70.040	NeIn1
Neolepas sp. 1	CIR	Kairei Field	YK09-13	HOV Shinkai 6500	Dive 1175	20091113	2422	-25.320	70.040	NeIn2
Neolepas sp. 1	CIR	Kairei Field	YK09-13	HOV Shinkai 6500	Dive 1175	20091113	2422	-25.320	70.040	NeIn3
Neolepas sp. 1	CIR	Kairei Field	YK09-13	HOV Shinkai 6500	Dive 1175	20091113	2422	-25.320	70.040	NeIn4
Neolepas sp. 1	SWIR	Dragon Field	JC067	ROV Kiel 6000	Dive 1	20111127	2770	-37.784	49.649	barJC6731B11
Neolepas sp. 1	SWIR	Dragon Field	JC067	ROV Kiel 6000	Dive 1	20111127	2770	-37.784	49.649	barJC6731B12
Neolepas sp. 1	SWIR	Dragon Field	JC067	ROV Kiel 6000	Dive 1	20111127	2770	-37.784	49.649	barJC6731B13
Neolepas zevinae/rapanuii	East Pacific Rise	9 50'N	AT07-06	HOV Alvin	Dive 3754	20020116	2499	9.827	-104.292	bar22
Neolepas zevinae/rapanuii	East Pacific Rise	9 50'N	AT07-06	HOV Alvin	Dive 3754	20020118	2490	9.848	-104.289	bar23
Neolepas zevinae/rapanuii	SEPR	17 S	AT03-28	HOV Alvin	Dive 3294	19981025	2573	-17.418	-113.204	bar15
Neolepas zevinae/rapanuii	SEPR	17 S	AT03-28	HOV Alvin	Dive 3294	19981025	2573	-17.418	-113.204	SEPR1
Neolepas zevinae/rapanuii	SEPR	17 S	AT03-28	HOV Alvin	Dive 3294	19981025	2573	-17.418	-113.204	SEPR2
Neolepas zevinae/rapanuii	SEPR	17 S	AT03-28	HOV Alvin	Dive 3294	19981025	2573	-17.418	-113.204	SEPR3
Neolepas zevinae/rapanuii	SEPR	17 S	AT03-28	HOV Alvin	Dive 3294	19981025	2573	-17.418	-113.204	SEPR4
Neoverruca sp. 1	Izu-Ogasawara Arc	Myojin Knoll	NT99-09	HOV Shinkai 2000	Dive 1112	19990629	1340	32.105	139.867	NeOg1
Neoverruca sp. 1	Izu-Ogasawara Arc	Myojin Knoll	NT99-09	HOV Shinkai 2000	Dive 1112	19990629	1340	32.105	1 39.86 7	NeOg2
Neoverruca sp. 1	Izu-Ogasawara Arc	Myojin Knoll	NT99-09	HOV Shinkai 2000	Dive 1112	19990629	1340	32.105	1 39.86 7	NeOg3
Neoverruca sp. 1	Izu-Ogasawara Arc	Myojin Knoll	NT99-09	HOV Shinkai 2000	Dive 1112	19990629	1340	32.105	139.867	NeOg4
Vulcanolepas osheai	Kermader Arc	Brothers Seamount	TAN1007	Epibenthic Sledge	Station 080	20100604	1342	-34.879	1 79.07 0	595B2
Vulcanolepas osheai	Kermader Arc	Brothers Seamount	TAN1007	Epibenthic Sledge	Station 080	20100604	1342	-34.879	1 79.07 0	595G
Vulcanolepas osheai	Kermader Arc	Brothers Seamount	TAN1007	Epibenthic Sledge	Station 079	20100604	1437	-34.878	1 79.0 71	bar592C
Vulcanolepas osheai	Kermader Arc	Brothers Seamount	TAN1007	Epibenthic Sledge	Station 080	20100604	1342	-34.879	179.070	bar595D4
Vulcanolepas osheai	Kermader Arc	Clark Seamount	KOK0506	HOV Pisces V	Dive 623	20050428	884	-36.447	177.839	337451
Vulcanolepas osheai	Kermader Arc	Clark Seamount	KOK0506	HOV Pisces V	Dive 623	20050428	884	-36.447	177.839	337452

Vulcanolepas osheai	Kermader Arc	Clark Seamount	KOK0506	HOV Pisces V	Dive 623	20050428	884	-36.447	177.839	337453
Vulcanolepas osheai	Kermader Arc	Clark Seamount	KOK0506	HOV Pisces V	Dive 623	20050428	884	-36.447	177.839	337454
Vulcanolepas osheai	Kermader Arc	Healy Seamount	TAN1104	Epibenthic Sledge	Station 073	20110313	1255	-35.014	178.980	72638_22
Vulcanolepas osheai	Kermader Arc	Tangaroa Seamount	TAN1206	Epibenthic Sledge	Station 017	20120416	682	-36.325	178.031	82121_15
Vulcanolepas scotiaensis	East Scotia Rise	E2	JC042	ROV Isis	Dive 133	20100123	2700	-56.060	-30.330	16640
Vulcanolepas scotiaensis	East Scotia Rise	E2	JC042	ROV Isis	Dive 133	20100123	2700	-56.060	-30.330	16641
Vulcanolepas scotiaensis	East Scotia Rise	E2	JC042	ROV Isis	Dive 133	20100123	2700	-56.060	-30.330	16642
Vulcanolepas scotiaensis	East Scotia Rise	E2	JC042	ROV Isis	Dive 133	20100123	2700	-56.060	-30.330	16643
Vulcanolepas scotiaensis	East Scotia Rise	E9	JC042	ROV Isis	Dive 141	20100130	2400	-60.050	-29.930	46923
Vulcanolepas scotiaensis	East Scotia Rise	E9	JC042	ROV Isis	Dive 141	20100130	2400	-60.050	-29.930	46924
Vulcanolepas scotiaensis	East Scotia Rise	E9	JC042	ROV Isis	Dive 141	20100130	2400	-60.050	-29.930	46925
Vulcanolepas scotiaensis	East Scotia Rise	E9	JC042	ROV Isis	Dive 141	20100130	2400	-60.050	-29.930	46926
Vulcanolepas sp. 1	Lau Basin	NELSC	TN234	ROV Jason 2	Dive 415	20090508	1617	-15.383	-174.245	bar09
Vulcanolepas sp. 1	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	VuMaU1
Vulcanolepas sp. 1	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	VuMaU2
Vulcanolepas sp. 1	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	VuMaU3
Vulcanolepas sp. 1	Tonga Arc	Mata Ua Volcano	RR1211	ROV Quest 4000	Dive 328	20120920	2391	-15.017	-173.788	VuMaU4

Abbreviations: Central Indian Ridge (CIR), Eastern Lau Spreading Center (ELSC), North-East Lau Spreading Center (NELSC), southern East Pacific Rise (SEPR), Southwest Indian Ridge (SWIR)

Table S2. Accession numbers for sequences from the Superorder Thoracica retrieved from GenBank

cox1

CUXI			
Taxon ID	GenBank Accession		
Conchoderma hunteri	KC138462	Amphibalanus variegatus	KC138446
Conchoderma hunteri	KC138463	Amphibalanus variegatus	KC138447
Conchoderma virgatum	KC138464	Galkinia indica	JO946272
Lenas testudinata	KC138477	Galkinia indica	JO946215
Lepas testudinata	KC138478	Galkinia indica	JO946238
Paralenas en HNC2013	KC138502	Galkinia indica	IO946273
I anas anserifera	KC138474	Galkinia eaus	JO946276
Heteralenas ianonica	F1694788	Galkinia equus	JO946251
Heteralenas sn I MT2008	FU884170	Galkinia tabulatus	10946255
Heteralenas sp IMT2000	FU884172	Galkinia altianiculus	10946270
Heteralenas ianonica	FU884154	Galkinia altiapiculus	IO946287
Heteralenas japonica	FU884153	Galkinia altianiculus	IO946271
Heteralenas japonica	EU884169	Galkinia altianiculus	10946286
Paralanas sn I MT2008	EU884174	Galkinia depressa	10946289
Hatavalanas sp LM12000	EU884171	Galkinia depressa	10946285
Resudentemaria sulanta	KC138503	Galkinia depressa Galkinia depressa	10946288
Pseudociomeris suicaia	KC138503	Armatohalanus allium	KC138450
Chthamalua an UNC2012	KC138465	Armaiobaianas antam Cantallius hoogi	FI37031A
Chinamalus sp HNC2013	KC130405	Cantellius hoegi	KC138453
University of the state of the	KC130400	Cantellius hoegi	F1270215
Hexechamaesipho pilsbryi	KC130470 VC129471	Cantellius hoegi	KC138454
Callinia damaga	IC046262	Cantellius noegi	E1270211
Gaikinia aepressa	JQ940202	Cantellius sextus	FJ373311 KC128456
Megabalanus ajax	KC120400	Cantellius sextus	E1270212
Megabalanus ajax	KC1304/9	Cantellius sextus	FJ575512 KC128457
Megabalanus occator	KC120403	Vahiennus sextus Nahia grandia	E1270219
Megabalanus occalor	KC130404	Nobia grandia Nobia grandia	FJ575510 VC128406
Megabalanus zebra	KC130491	Nobia grandis Nobia grandis	KC130490
Megabalanus zebra	KC130492	Nobia granais Palama trigoma	KC130457
Megabalanus linlinnabulum	KC130407	Dalanus trigonus Palanus trigonus	KC130451
Megabalanus liniinnabulum	KC130400	Balanus irigonus Wanalla millanonaa	EE565204
Megabalanus volcano	KC138489	Wanella milleporde	EF J0J204
Megabalanus voicano	KC130490	Wanella milleporde Wanella milleporde	KC136514 KC128515
Megabalanus rosa	KC130403	Amphihalamus amphitrita	KC138J15 KC138445
Megabalanus rosa	KC130400	Amphiloalanus amphilrite	KC13044J VC128469
Megabalanus coccopoma	KC130401	Fistulobalanus albicostatus	KC130400
Megabalanus coccopoma	KC130402	Fisitiobalanus albicostatus	KC130403
Ampnibalanus znujlangensis	IC000064	Memoranobalanus longirosirum Cantollius pallidus	E1270217
Darwiniella conjugalum	JQ966904	Wanalla millenonae	FJ5/751/ FE565206
Nobla conjugatum	EF 309337	Totraolitalla karandai	EF 303200 KC138510
Darwiniella conjugatum	JQ988945	Tetraciliella karandei	KC138510
Darwiniella conjugatum	JQ988940 IO080022	Tetrachiella shinansis	KC138506
Darwiniella conjugalum	JQ989055 EE560556	Vamaguahialla an HNC2013	KC138510
Nobia conjugatum	EF309330	Vamaguchiella an HNC2013	KC130512 KC130513
Cionophorus soongi	JQ940277	Navaguchiella an HNC2013	KC136313
Cionophorus soongi	JQ940278	Newmanella sp HNC2013	KC130494
Cionophorus soongi	JQ940279	Cholonibia natula	IE822664
Darwiniella angularis	JQ9889/4	Chelonibia patula	JF 823004 JE 823661
Darwiniella angularis	1000000	Chelonibia tastudinaria	JF 023001 VC120120
Darwiniella angularis	10046344	Chelonibia natula	TE272662
riroa studdingsi Himo a studdingsi	JQ940244 IO046946	Chelonibia patula Chelonibia patula	JE 023003
niroa stubbingsi Uinoa stubbingsi	JQ940240 IO046927	Chelonibia tastudinaria	JI 023002
niroa siuooingsi Himoo saubbingsi	JQ940237	Chelonibia testudinaria	KU130401 IE032720
riiroa siuooingsi	JQ946245	Chevonioia iestuainaria	JE 023/39

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Chelonibia testudinaria	JF823740	Tetraclita formosana	DQ647761
Tetraclita kuroshioensis	KC138505	Tetraclita japonica	DO647704
Tetraclitella chinensis	KC138507	Tetraclita japonica	DO647707
Darwiniella angularis	JQ988981	Tetraclita formosana	DO647764
Capitulum mitella	KC138458	Tetraclita japonica	DO647706
Capitulum mitella	KC138459	Tetraclita formosana	DO647763
Ibla cumingi	KC138472	Tetraclita japonica	DO647705
Ibla cumingi	KC138473	Tetraclita formosana	DO647762
Galkinia equus	JQ946226	Tetraclita singaporensis	EF035167
Tetraclitella divisa	KC138508	Tesseropora rosea	FJ516076
Tetraclitella divisa	KC138509	Tesseropora rosea	FJ516077
Chthamalus sp 1 JDZ2005	AY823025	Tesseropora rosea	FJ516074
Tesseropora wireni pacifica	AY823032	Tesseropora rosea	FJ516075
Amphibalanus improvisus	FJ845840	Tetraclita rubescens	GU381926
Trevathana mizrachae	FJ620805	Catomerus polymerus	FJ516172
Trevathana mizrachae	FJ620802	Catomerus polymerus	FJ516173
Trevathana mizrachae	FJ620803	Catomerus polymerus	FJ516175
Trevathana mizrachae	FJ620804	Catomerus polymerus	FI516174
Trevathana margaretae	FJ620810	Tetraclita rubescens	GU381927
Trevathana margaretae	FJ620811	Chthamalus cf challengeri JD72005	AY823019
Trevathana margaretae	FJ620812	Tetrachthamalus sinensis	IO755178
Trevathana dentata	FI620820	Chinochthamalus scutelliformis	10755182
Trevathana dentata	FJ620821	Chthamalus fragilis	10755179
Trevathana dentata	FI620822	Chthamalus hisimatus	F1845840
Trevathana jensi	F1620823	Cononea sp 4 DCS2011	10066201
Trevathana jensi	FI620824	Tetraclita rubescens	GU381025
Trevathana jensi	FI620825	Chthamalus moro	FU304448
Trevathana jensi	FI620826	Tetraclitella nurnurascens	EU304446
Trevathana sarae	FI620798	Tetraclitella purpurascens	FI516110
Trevathana sarae	FI620800	Tetraclitella purpurascens	FI516127
Savignium crenatum	FI620792	Tetraclitella nurnurascens	FI516125
Savignium crenatum	FI620789	Herechamaesinho nilshni	KC806275
Savignium crenatum	FI620791	Pollicines pollicines	FE467050
Neotrevathana elongatum	FI620831	Pollicines pollicines	EF462950
Trevathana sarae	FI620830	Pseudoctomeris sulcata	IV022951
Trevathana sarae	FI620799	Chamaesinho tasmanica	JA063603 IV082867
Savignium crenctum	FI620799	Capitulum mitalla	JA063607
Trevathana margaretae	FI620813	Capitatian milena Chthamalus angustitargum	JAJU2990 E1946933
Semihalanus halanoides	F1845815	Chinamalus angustitergum	F1945932
Tetraclita rubescens	GU381028	Chinamalus angustitergum Chthamalus angustitergum	FJ04J0JU
I enas anatifera	GU003580	Chinamalus angustitergum Chihamalus angustitergum	FJ04J039 E1046022
Lepus unatifera Lepus anatifera	GU003500	Chinamalus angusinergum Chihamalus bisimuatus	FJ04J033
Lepas anatifera	GU003501	Chinamalus bisinuatus	FJ04303U
Lepas nectinata	GU003645	Chinamalus bisinuatus	FJ04J04J E1045046
Lepus pectinata Lepus pectinata	GU003644	Microeuraphia depressa	FJ04J040
Lepus pectinata	GU1003650	Microeuraphia depressa Microeuraphia depressa	EE005150
Lepas pecínica Lepas anserifera	GU993630	Microeuraphia depressa Microeuraphia depressa	EF095159
Lepus unserifera Lepus anserifera	GU003620	Chthamalus montagui	EF095100
Lepas australis	GU003638	Chinamalus montagui	FJ0J0001
Lepus dustralis Lepus australis	GU003630	Chinamalus montagui Chthamalus montagui	FJ858063
Lepas australis	GU003641	Chinamatus montagui	F1020002
I enas mistralis	GU002640	Chinanalus stollatur	FJ0J0UU
Lopas austrationa	GI 1002 499	Chthamalus stellatur	EU09924/
Tetrachthamalus ablittaratus	AV/20012	Viununuus sienaus Miaroauraphia shirortoraa	EU099241
Chamaesinho sn ROI DAAW6872	DO860060	Microeuraphia phirophorae	ГЈ843804 Гј846846
Communication of DOLDAAN 00/2 Tetraclita singanoreusis	5005005 FEU36166	Microeuraphia phirophorae	ГЈ843803 П1946944
Tetraclita singaporensis	EF035100 EF025164	Furanhia aastronaceusia	ГЈ843800 Г1946961
Tetraelita singaporensis	EF033104 EF025145	Europhia costropacensis	FJ843831
ien actita singaporensis	ELO33103	Euraphia eastropacensis	rj843832

Europhia aastropacansis	F1845858	Chthamalus hedgecocki	FJ857987
Chthamalus neglectus	F1858077	Chthamalus hedgecocki	FJ857990
Chinamatus neglectus Chthamalus neglectus	F1858079	Chthamalus southwardorum complex sp B	FJ857992
Chinamatus neglectus Chthamalus neglectus	FJ858080	Chthamalus southwardorum complex sp B	FJ857997
Chinamatus neglectus Chthamatus neglectus	F1858078	Chthamalus southwardorum complex sp B	FJ857999
Chinamatus negrecius Chthamalus malayansis	F1845828	Chthamalus southwardorum complex sp B	FJ857993
Chinamatus malayensis Chthamalus malayensis	F1845830	Chthamalus fragilis	AF234807
Chinamatus malayensis Chinamatus malayensis	F1845831	Chthamalus fragilis	AF234813
Chinamatus malayensis Chinamatus malayensis	F1845829	Chthamalus anisopoma	AF234816
Chinamatus maiayensis Chihamatus dontatus	F1858084	Chthamalus proteus	FJ858023
Chinamatus dentatus	F1858086	Chthamalus proteus	FJ858025
Chinamatus dentatus Chihamatus dentatus	F1858087	Chthamalus proteus	FJ858039
Chinamatus dentatus	F1858088	Chthamalus proteus	FJ858027
Chinamatus aetaanatus	IX083870	Chthamalus mexicanus	AF234804
Crimamanus amerinanus	F1845857	Chthamalus mexicanus	AF234805
Chthamalus contegianus	AF734810	Chthamalus mexicanus	AF234803
Chihamalus cortezianus	AF234811	Capitulum mitella	IX502999
Chihamalus cortezianus	AF234812	Chthamalus hedgecocki	F1857989
Chinamatus cortezianus Mierocuranhia sp. 2 MPI 2012	IY083873	Fistulobalarus albicostatus	IX503003
Microeuraphia sp 2 MPL2012	JA003073 IV092972	Chthamalus anisonoma	AF234818
Euraphia sp 2 MFL2012	JA003072	Chihamalus anisopoma	AF234817
Chinamatus chattengeri	FJ050007	Chinamatus anisopoma	ΔΕ234819
Chinamalus challengeri	FJ0J00/J E1050074	Chinamanus anisopoma Miaroauraphia daprassa	HO224881
Chinamalus challengeri	ГЈОЈОU/4 Е1050075	Chthamalus stallatus	FU600240
Chinamaius challengeri	FJ0J0U/J	Chinamalus stellatus	EU699243
Hexechamaesipno pilsoryi	JAU03000 IV002071	Chinanalus Sielialus Miaroauranhia witharsi	AV430814
Cauaoeuraphia cauaala	JAU030/1	Chamaasinha huunnaa	AV/30811
Megatrema anglicum	FJ/15101	Lablius cirratus	GU126095
Nesochinamalus intertextus	JA003009	Jennus cirraius	GU126000
Megabalanus rosa	JAJ03004	Jenitus cirratus	GU126087
Megabalanus rosa	JAJ03003	Jenitus cirratus	GU126007
Balanus gianaula	EF094392	Determentia angulosa	AV/280/0
Balanus glanaula	EF094390	Notochthamalus soghrosus	E19/5971
Balanus glanaula	EF094394	Notochthamalus scabrosus	F1845822
Balanus glanaula	EF094393	Notochthamalus scabrosus	GU125776
Darwiniella sp HNC2013	NC130407	Notochinamatus scaorosus Dollicinos nolumenus	GUA2485
Conopea galeata	JQ900287	Pollicing polymerus	GU442485
Conopea galeata	JQ900288 IV092966	Pollicipes polymerus Pollicipes polymerus	GU442491
Chamaesipho columna	JA003000	Pollicing polymerus	GU142492
Semibalanus carlosus	GQ902333	Pollicing achoustdaysis	HM563665
Semibalanus carlosus	GQ902333	Pollicipes caboverdensis Pollicipes caboverdensis	HM563666
Semidalanus carlosus	GQ902332	Pollicings caboyerdensis	HM563667
Semidalanus carlosus	GQ902334	Pollicing achoustdansis	HM563668
Semibalanus balanoides	FJ04J019	Pollicing pollicing	HM563675
Semibalanus balanoides	FJ043010	Pollicipes pollicipes	HM563676
Semibalanus balanolaes	GQ320904	Pollicines pollicines	HM563678
Austrobalanus imperator	EU423232	Pollicines pollicines	HM563677
Tetraciita kurosnioensis	JA100409	Caltinia adamantana	IV083100
Amphibalanus improvisus	FJ04J041 E1045044	Oatomoria brunnag	AV430812
Amphibalanus improvisus	FJ043044	Caltinia decima	10046213
Amphibalanus improvisus	FJ04J042	Cantollius continuus	KC138455
Verruca stroemia	JA003003	Cumentus septimus Totraalita paoifiaa	DO363694
Catophragmus impricatus	JAV03004 E1016063	Ten uchu puchicu Tetraelita pacifica	DU383803
Microeurapnia rnizopnorae	FJ04J00J E1057074	Tetraolita pacifica Totraolita pacifica	DO363603
Chinamalus panamensis	FJ0J/7/0 E1057074	Tetraolita pacifica Totraolita pacifica	D0363604
Chinamalus panamensis	FJ0J/9/4	Ten acina pacifica Tetraelita sauaresa	DO362704
Chinamalus panamensis	ГЈ8Ј/УОU Б1967044	Tetraolita squamosa	DQ363704
Chinamaius panamensis	ГЈ0 <i>Ј /</i> 900 Г19 67 096	Tetraolita squamosa Totraolita squamosa	DO363703
Cninamalus hedgecocki	L192 / 790	tetracitta squamosa	20100200

Totraclita aguamora	D0262706	Alt:	10105607
Terracina squamosa	DQ303700	Altiverruca sp N1020/&08	AB195607
Conopea calceola	HQ290142	Metaverruca recta	JX083861
Conopea calceola	HQ290143	Octolasmis angulata	KC138498
Conopea sp B DCS2011	HQ290136	Octolasmis cor	KC138499
Conopea sp B DCS2011	HQ290141	Octolasmis cor	KC138500
Conopea sp B DCS2011	HQ290139	Octolasmis orthogonia	EU884173
Conopea sp B DCS2011	HQ290140	Octolasmis warwickii	KC138501
Conopea caiceola	HQ290134	Neoverruca brachylepadoformis	AB195606
Conopea sp A DCS2011	HQ290135	Neoverruca sp Ok8	AB195598
Conopea sp A DCS2011	HQ290136	Neoverruca sp Ok14	AB195604
Conopea sp A DCS2011	HQ290137	Neoverruca sp Ok6	AB195596
Conopea cf galeata USA DCS2011	HQ290146	Neoverruca sp Ok9	AB195599
Conopea cf galeata Panama DCS2011	HQ290130	Neoverruca sp Okl	AB195591
Conopea cf galeata Brazil DCS2011	HQ290133	Neoverruca sp Ok5	AB195595
Conopea cf galeata Galapagos DCS2011	HQ290144	Neoverruca sp Ok15	AB195605
Chthamalus moro	HM135959	Neoverruca sp Ok7	AB195597
Chthamalus moro	HM135960	Neoverruca sp Ok11	AB195601
Tetraclita stalactifera	JN589833	Neoverruca sp Ok13	AB195603
Chelonibia caretta	JN589810	Neoverruca sp Ok10	AB195600
Chelonibia caretta	JN589812	Neoverruca sp Og5	AB195586
Chelonibia caretta	JN589811	Neoverruca sp Og1	AB195582
Chthamalus dalli	AY795282	Neoverruca sp 0g4	AB195585
Chthamalus dalli	AY795283	Neoverruca sp Og3	AB195584
Chthamalus dalli	AY795285	Neoverruca sp Oge	AB195587
Chthamalus dalli	AY795284	Neoverruca sp 0g0 Neoverruca sp 0g7	AB105588
Chthamalus fissus	DO538424	Neoverruca sp Og	AB105500
Chthamalus fissus	DQ538424	Neovernica sp Ogg	AD195590
Chthamalus fissus	DQ538422	Neovernuca sp Ogo	AD195504
Chinanaius fissus	DQ538421	Neovernuca sp Ok4	AD195594
Chinamatus jissus	DQ538425	Neoverrueg on Ok2	AD195002
Chinamatus sp kino Chihamatus sp kino	DQ336449	Neoverruca sp Oks	AB195593
Chinamatus sp kino	DQ538448	Neoverruca sp Og2	AB195583
Chinamatus sp kino	DQ538447	Neoverruca sp Ok2	AB195592
Chinamaius sp kino Teteredite e bassi	DQ538446	Chinamalus sp 2 JDZ2005	AY823028
Tetracitta ensani	JX180290	Chthamalus sp 2 JDZ2005	AY823030
Ietraclita ensani	JX186295	Chthamalus sp 2 JDZ2005	AY823029
Ietraclita ehsani	JX186297	Chthamalus sp 2 JDZ2005	AY823031
Tetraclita ehsani	JX186298	Pollicipes sp JQ2009	GQ472625
Tetraclita serrata	JX186199	Pollicipes sp JQ2009	GQ472627
Tetraclita serrata	JX186201	Pollicipes sp JQ2009	GQ472628
Tetraclita serrata	JX186200	Pollicipes sp JQ2009	GQ472626
Tetraclita serrata	JX186202	Pollicipes elegans	GQ472619
Tetraclita sp n LMT2012	JX186368	Pollicipes elegans	GQ472614
Tetraclita sp n LMT2012	JX186366	Pollicipes elegans	GQ472615
Tetraclita sp n LMT2012	JX186369	Pollicipes elegans	GQ472618
Tetraclita sp n LMT2012	JX186367	Notochthamalus scabrosus	FJ845824
Tetraclita reni	JX186294	Calantica spinosa	AY428047
Tetraclita reni	JX186365	Notochthamalus scabrosus	FJ84582
Tetraclita achituvi	JX186290	Calantica spinosa	AY428047
Tetraclita achituvi	JX186291		
Tetraclita achituvi	JX186289		
Tetraclita achituvi	IX186288		
Lenas anserifera	KC138475		
Lenas nectinata	KC138476		
Vermuca laguigata	TV092967		
Armatohalamis allium	JAU0J002 KC139440		
Armaioodianus dillum Postratovarmiaa kmiaari	AD105600		
Kosiraioverruca krugeri	AD193009		
verruca sp CJS2008	EU439973		
Brochiverruca sp K10203a	AB195608		

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Taxon ID	Accession		
Altiverruca sp KACb00436	EU082300	Microeuraphia depressa	JX083927
Metaverruca recta	EU082297	Octomeris brunnea	JX083917
Verruca laevigata	EU082296	Hexechamaesipho pilsbryi	JX083918
Verruca stroemia	AY520615	Pachylasma japonicum	AB723931
Rostratoverruca sp KACb00435	EU082298	Stephanolepas muricata	AB723918
Rostratoverruca krugeri	EU082299	Tetraclitella divisa	AY520603
Chelonibia caretta	AB723915	Tetraclitella purpurascens	AY520604
Chelonibia manati	AB723917	Semibalanus balanoides	AY520592
Chelonibia patula	EU082295	Semibalanus balanoides	EU370440
Megabalanus occator	AB723916	Semibalanus cariosus	AY 520593
Megabalanus tintinnabulum	AY520597	Microeuraphia rhizophorae	JX083929
Megabalanus californicus	AY520598	Lithotrya valentiana	EU082301
Megabalanus californicus	AY859588	Lithotrya sp KACb00393	EU082302
Austromegabalanus psittacus	AY520600	Balanus crenatus	AY520590
Megabalanus spinosus	AY520599	Balanus glandula	AY 520591
Balanus perforatus	AY520595	Balanus balanus	AY520594
Menesiniella aquila	AY520596	Pollicipes pollicipes	AY520616
Stomatolepas praegustator	AB723919	Pollicipes pollicipes	EU370441
Stomatolepas sp RH2012	AB723921	Lepas sp Lep 1	EU914256
Stomatolepas transversa	AB723920	Pollicipes polymerus	AY520617
Tubicinella cheloniae	AB723922	Capitulum mitella	AY520618
Platylepas decorata	AB723923	Calantica spinosa	EU082303
Cylindrolepas sinica	AB723926	Calantica sp KACb00087	EU082304
Xenobalanus globicipitis	AB723927	Smilium peronii	EU082305
Coronula diadema	AB723928	Litoscalpellum discoveryi	EU4 89824
Cryptolepas rhachianecti	AB723929	Trianguloscalpellum regium	EU082308
Tetraclita squamosa	AY520605	Litoscalpellum regina	AY520619
Tetraclita japonica	AY520606	Scalpellum scalpellum	EU082307
Platylepas hexastylos	AB723924	Arcoscalpellum sp CJS2008	EU489828
Cylindrolepas darwiniana	AB723925	Arcoscalpellum sp CJS2008	EU489829
Austrobalanus imperator	AB723930	Arcoscalpellum africanum	EU489831
Catomerus polymerus	AY520614	Arcoscalpellum africanum	EU489834
Catophragmus imbricatus	JX083911	Arcoscalpellum africanum	EU489835
Pseudoctomeris sulcata	JX083912	Arcoscalpellum africanum	EU489833
Elminius modestus	AY520601	Litoscalpellum sp CJS2008	EU489837
Elminius kingii	AY520602	Litoscalpellum sp CJS2008	EU489839
Notochthamalus scabrosus	AY520612	Scalpellum sp CJS2008	EU489830
Octomeris angulosa	JX083916	Ornatoscalpellum stroemii	EU082306
Chamaesipho brunnea	JX083915	Arcoscalpellum beuveti	EU489827
Chamaesipho tasmanica	AY520613	Leucolepas longa	EU082311
Chamaesipho sp MPL2012	JX083914	Vulcanolepas sp KACb00419	EU082312
Nesochthamalus intertextus	JX083919	Vulcanolepas osheai	EU082313
Chamaesipho columna	JX083913	Neolepas rapanuii	EU082309
Chthamalus stellatus	AY520607	Neolepas zevinae	EU082310
Chthamalus montagui	AY520608	Ashinkailepas seepiophila	EU082314
Chthamalus challengeri	AY520609	Neoverruca sp KACb00361	EU082315
Jehlius cirratus	AY520611	Neoverruca sp KACb00389	EU082316
Microeuraphia withersi	JX083928	Neoverruca brachylepadoformis	EU082317
Chthamalus bisinuatus	AY 520610	Ibla quadrivalvis	AY 520621
Tetrachthamalus oblitteratus	JX083925	Ibla cumingi	EUU82332
Caudoeuraphia caudata	JX083926	Poecilasma inaequilaterale	AY 520620
Chthamalus malayensis	JX083922	Poecilasma kaempjeri	EUU82329
Chthamalus dentatus	JX083923	Megalasma striatum	EUU82330
Chthamalus anisopoma	JX083924	Octolasmis cor	EUU82320
Chthamalus antennatus	JX083920	Octolasmis sp KACDUUU04	EUU8232/

Octolasmis warwickii	EU082328
Conchoderma auritum	EU082320
Conchoderma virgatum	EU082321
Lepas pectinata	EU082322
Lepas anserifera	EU082323
Lepas australis	EU082324
Lepas testudinata	EU082325
Paralepas dannevigi	EU082318
Oxynaspis celata	EU082331
Heteralepadomorpha sp KACb00398	EU082319

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Taxon ID

Vulcanolepas sp KACb00419 Vulcanolepas osheai Neolepas zevinae Leucolepas longa Neolepas rapanuii Ashinkailepas seepiophila Vulcanolepas sp East Scotia Ridge Neoverruca sp KACb00389 Neoverruca sp KACb00361 Neoverruca brachylepadoformis Lithotrya sp KACb00393 Lithotrya valentiana Tesseropora rosea Tetraclita achituvi Tetraclita rufotincta Tetraclita sp n LMT2012 Tetraclita sp n LMT2012 Tetraclita sp n LMT2012 Tetraclita sp n LMT2012 Tetraclita ehsani Tetraclita kuroshioensis Tetraclita serrata Tetraclita serrata Tetraclita serrata Trianguloscalpellum regium Chelonibia patula Metaverruca recta Verruca laevigata Conopea calceola Conopea calceola Conopea calceola Conopea fidelis Conopea sp B DCS2011 Conopea fidelis Conopea sp B DCS2011 Conopea sp B DCS2011 Conopea fidelis Conopea sp B DCS2011 Conopea calceola Conopea calceola Conopea sp A DCS2011 Conopea saotomensis Conopea saotomensis Conopea saotomensis

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HQ290158

HQ290157

KC349905

HQ290151

HQ290152

HQ290149

HQ290160

KC349911

KC349906

KC349913

Cono	pea saotomensis	KC349909
Cono	pea saotomensis	KC349904
Cono	pea sp A DCS2011	HQ290159
Cono	pea saotomensis	KC349907
Cono	pea saotomensis	KC349908
Rostro	atoverruca krugeri	EU082341
Rostra	atoverruca sp KACb00435	EU082340
Mega	lasma striatum	EU082370
Octole	asmis sp KACb00064	EU082367
Lepas	anserifera	EU082363
Lepas	australis	EU082364
Lepas	testudinata	EU082365
Lepas	pectinata	EU082362
Octole	asmis cor	EU082366
Oxynd	aspis celata	EU082371
Hetero	alepadomorpha sp KACb00398	EU082359
Octold	asmis warwickii	EU082368
Scalpe	ellum scalpellum	EU082347
Calan	tica sp KACb00087	EU082345
Smiliu	m peronii	EU082346
Calan	tica spinosa	EU0 8234 4
Pyrgo	psella sp SMB2011	JN800715
Savigr	nium crenatum	JN800716
Conch	noderma auritum	EU082360
Conch	oderma virgatum	EU082361
Conop	pea cf galeata Galapagos DCS2011	HQ290162
Conop	ea cf galeata Galapagos DCS2011	HQ290161
Conop	ea cf galeata USA DCS2011	HQ290147
Conop	ea galeata	JQ966286
Conop	ea galeata	JQ966283
Conop	ea galeata	JQ966284
Conop	ea galeata	JQ966285
Parale	epas dannevigi	EU082358
Pollici	ipes elegans	HM142348
Pollici	pes sp RJVS2010	HM142349
Poecil	asma kaempferi	EU082369
Ibla cı	ımingi	EU082372

Table S3. Predictions of number of RAD-tags in thoraciacan barnacles using Sbfl. Data for *Daphnia pulex* obtained from the U.S. National Center for Biotechnology Information (NCBI) WGS database. Observed frequency of recognition sequences and calculated probability based on a trinucleotide genome composition model were generated following the methodology described by Herrera *et al.* (Herrera *et al.* 2014). Data for known barnacle genome sizes obtained from the Animal Genome Size Database (http://ww.genomesize.com). C-value is the amount of DNA in picograms in the nucleous, where the genome size in Mbp = 978 x C-value

Species	Common name	C-value	Genome size (Mbp)	Observed frequency of SbfI recognition sites per bp	Probability of SbfI recognition site per bp		
Genebank WGS							
Daphnia pulex	Water flea		158.61	7.48E-06	5.90E-06		
Species	Common name	C-value	Genome size (Mbp)	Predicted number of SbfI recognition sites based on <i>D.</i> <i>pulex</i> observed frequency	Predicted number of Sbfl RAD-tags based on <i>D.</i> <i>pulex</i> observed frequency	Predicted number of Sbfl recognition sites based on <i>D. pulex</i> trinucleotide genome composition probability	Predicted number of Sbfl RAD-tags based on <i>D.</i> <i>pulex</i> trinucleotide genome composition probability
Animal Genome Size	e Database						
Balanus amphitrite	Striped barnacle	0.74	723.72	5,411.68	10,823.35	4,267.22	8,534.43
Balanus amphitrite	Striped barnacle	1.4	1,369.20	10,238.31	20,476.61	8,073.11	16,146.22
Balanus cariosus	Thatched barnacle	1.4	1,369.20	10,238.31	20,476.61	8,073.11	16,146.22
Balanus eburneus	Ivory barnacle	1.26	1,232.28	9,214.48	18,428.95	7,265.80	14,531.60
Chthalamus sp.	Acorn barnacle	1.23	1,202.94	8,995.08	17,990.17	7,092.80	14,185.61
Unknown sp.	Goose barnacle	1.46	1,427.88	10,677.09	21,354.18	8,419.10	16,838.20
Mitella polymerus	Pacific goose barnacle	0.9	880.20	6,581.77	13,163.54	5,189.86	10,379.71
Tetraclita rubescens	Volcano barnacle	2.6	2,542.80	19,014.00	38,027.99	14,992.92	29,985.84

Table S4. Nucleotide substitution models for each Sanger-based genetic marker, as selected by the BIC criterion in jModeltest.

Dataset	coxI_codon12	H3_codon12	H3_codon3	28 S
Thoracica	GTR+I+G	HKY	+I+G	GTR+I+G
Clade A	HKY+I	HK	Y+I	HKY+G

Table S5. Results from Xia saturation test for each Sanger-based genetic marker.

Dataset	coxI_codon12	H3_codon12	H3_codon3	288
Thoracica	Little saturation	Substantial saturation	Little saturation	Little saturation
C1	Little saturation	Substantial saturation	Little saturation	Little saturation

Table S6. RAD sequencing results and filtering statistics.

	·	5		S	TACKS filteri	ng	p	yRAD filterin	g
ID Species	ID sample	DNA integrity, HMW band intensity, and shear level (agarose geł)*	Total sequenced reads	Reads discarded due to absence of RAD-tag	Reads discarded due to low quality	Retained reads	Reads that passed quality filtering	Total number of retained reads	Percentage of retained reads after filtering steps
Vulcanolepas_scotiaensis_ESR_E2	16640	H; Strong HMW; NS	1,079,806.00	45,540.00	37,560.00	996,706.00	955,965.00	955,965.00	88.53
Vulcanolepas_osheai_Kermadec_Healy	72638_22	H; Strong HMW; NS	1,647,804.00	35,100.00	64,050.00	1,548,654.00	1,476,024.00	1,476,024.00	89.58
Vulcanolepas_osheai_Kermadec_Tangaroa	82121_15	H; Strong HMW; LS	2,133,810.00	47,251.00	95,743.00	1,990,816.00	1,875,914.00	1,875,914.00	87.91
Ashinkailepas_kermadecensis_Lau_Basin_Niua_N	AsNiN2	H; Strong HMW; LS	1,091,017.00	43,293.00	41,739.00	1,005,985.00	954,951.00	954,951.00	87.53
Ashinkailepas_seepiophila_Okinawa_Trough	AsOk3	M; Moderate HMW, SS	580,247.00	107,879.00	14,673.00	457,695.00	438,974.00	438,974.00	75.65
Neolepas_zevinae_EPR_Tica2	bar22	M; Moderate HMW, SS	643,213.00	53,267.00	20,219.00	569,727.00	549,871.00	549,871.00	85.49
Neobrachylepas_relica_Lau_Basin_NELSC	bar06	H; Strong HMW; LS	1,004,066.00	69,318.00	35,161.00	899,587.00	856,092.00	856,092.00	85.26
Neolepas_sp1_SWIR_Dragon_vent	JC6731B11	H; Strong HMW; LS	984,409.00	65,379.00	27,756.00	891,274.00	862,579.00	862,579.00	87.62
Neolepas_sp1_CIR	NeIn2	L; No HMW, SS	232,285.00	142,801.00	6,243.00	83,241.00	76,759.00	76,759.00	33.05
Neoverruca_sp1_Ogasawara_Arc	NeOg1	L; No HMW, SS	104,531.00	16,470.00	5,000.00	83,061.00	77,248.00	77,248.00	73.90
Neolepas_zevinae_SEPR	SEPR3	L; No HMW, SS	236,557.00	70,581.00	7,952.00	158,024.00	149,850.00	149,850.00	63.35
Vulcanolepas_sp1_Lau_Basin_Mata_Ua	VuMaU1	H; Strong HMW; LS	923,322.00	55,368.00	41,035.00	826,919.00	783,374.00	783,374.00	84.84
Leucolepas_longa_Mariana_Arc_TOTO_vent	VuTO2	M; Moderate HMW, SS	304,975.00	87,453.00	7,744.00	209,778.00	202,382.00	202,382.00	66.36

.

*Abbreviations: High integrity (H), Medium integrity (M), Low integrity (L), High Molecular Weight (HMW), No Shear (NS), Low Shear (LS), Strong Shear (SS)

Table S7. RAD clustering statistics.

Standard deviation Standard Number of clusters Mean depth of of clusters with Total number Mean depth of deviation of with depth greater clusters with depth depth greater than ID sample of clusters clusters cluster depth than 5 greater than 5 5 16640 21,256 43.61 127.54 18,198 50.62 136.60 72638_22 25,837 54.66 219.37 22,289 63.05 235.10 82121_15 26,882 66.65 279.28 23,015 77.53 300.47 AsNiN2 22,793 38.14 189.94 18,847 45.70 208.09 17,057 AsOk3 24.22 131.27 12,610 32.08 151.89 bar22 18,745 28.32 88.47 14,987 34.93 97.83 bar06 20,640 39.36 160.01 17,346 46.47 173.64 JC6731B11 20,340 40.74 148.45 17,359 47.39 159.75 NeIn2 6,133 12.25 25.82 3,706 19.07 31.39 6,055 NeOg1 12.36 18.57 3,797 18.61 21.09 SEPR3 11,611 29.25 12.73 24.74 18.96 7,380 VuMaU1 20,884 95.06 35.76 17,346 42.65 102.96 VuTO2 10,607 18.50 59.61 7,508 25.33 69.71

pyRAD clustering (c85)

pyRAD clustering (c80)

ID sample	Total number of clusters	Mean depth of clusters	Standard deviation of cluster depth	Number of clusters with depth greater than 5	Mean depth of clusters with depth greater than 5	Standard deviation of clusters with depth greater than 5
16640	19,884	44.17	101.89	17,283	50.52	107.86
72638_22	24,008	57.15	228.03	21,070	64.84	242.41
82121_15	24,580	69.94	282.61	21,537	79.55	300.68
AsNiN2	21,056	39.13	191.94	17,750	46.05	208.32
AsOk3	15,728	24.50	105.61	11,915	31.73	120.45
bar22	17,698	28.61	83.40	14,291	34.96	91.68
bar06	19,441	40.50	155.99	16,580	47.15	168.02
JC6731B11	19,045	41.52	139.60	16,444	47.77	149.28
NeIn2	5,848	12.63	27.34	3,600	19.35	33.10
NeOg1	5,831	12.76	19.88	3,712	18.99	22.66
SEPR3	11,244	12.81	20.59	7,195	18.96	23.60
VuMaU1	19,413	36.77	94.33	16,375	43.23	101.40
VuTO2	10,045	18.81	59.55	7,194	25.49	69.23

pyRAD clustering (c90)

Total number of clusters	Mean depth of clusters	Standard deviation of cluster depth	Number of clusters with depth greater than 5	Mean depth of clusters with depth greater than 5	Standard deviation of clusters with depth greater than 5
19,181	44.74	103.33	16,828	50.72	108.99
23,039	57.73	211.75	20,417	64.89	223.93
23,544	71.10	287.50	20,872	79.96	304.21
20,266	39.13	174.49	17,184	45.78	188.72
15,095	24.79	106.79	11,585	31.71	121.05
17,166	28.80	84.49	13,943	35.00	92.65
18,912	41.06	158.41	16,206	47.59	170.25
18,387	41.72	137.37	16,015	47.61	146.27
5,679	12.66	25.03	3,542	19.18	29.85
5,692	12.95	20.34	3,661	19.11	23.16
11,021	12.86	19.55	7,089	18.95	22.13
18,692	37.02	94.83	15,904	43.16	101.57
9,761	19.05	60.22	7,043	25.65	69.78
	Total number of clusters 19,181 23,039 23,544 20,266 15,095 17,166 18,912 18,387 5,679 5,692 11,021 18,692 9,761	Total number of clustersMean depth of clusters19,18144.7423,03957.7323,54471.1020,26639.1315,09524.7917,16628.8018,91241.0618,38741.725,67912.665,69212.9511,02112.8618,69237.029,76119.05	Total number of clusters Mean depth of clusters Standard deviation of cluster depth 19,181 44.74 103.33 23,039 57.73 211.75 23,544 71.10 287.50 20,266 39.13 174.49 15,095 24.79 106.79 17,166 28.80 84.49 18,912 41.06 158.41 18,387 41.72 137.37 5,679 12.66 25.03 5,692 12.95 20.34 11,021 12.86 19.55 18,692 37.02 94.83 9,761 19.05 60.22	Total number of clustersMean depth of clustersStandard deviation of cluster depthNumber of clusters with depth greater than 519,18144.74103.3316,82823,03957.73211.7520,41723,54471.10287.5020,87220,26639.13174.4917,18415,09524.79106.7911,58517,16628.8084.4913,94318,91241.06158.4116,20618,38741.72137.3716,0155,67912.6625.033,5425,69212.9520.343,66111,02112.8619.557,08918,69237.0294.8315,9049,76119.0560.227,043	Total number of clustersMean depth of clustersStandard deviation of cluster depthNumber of clusters with depth greater than 5Mean depth of clusters with depth greater than 519,18144.74103.3316,82850.7223,03957.73211.7520,41764.8923,54471.10287.5020,87279.9620,26639.13174.4917,18445.7815,09524.79106.7911,58531.7117,16628.8084.4913,94335.0018,91241.06158.4116,20647.5918,38741.72137.3716,01547.615,67912.6625.033,54219.185,69212.9520.343,66119.1111,02112.8619.557,08918.9518,69237.0294.8315,90443.169,76119.0560.227,04325.65

RAD-seq matrix	Number of loci	Number of base pairs	Percentage of missing data	Number of variable sites	Percentage of variable sites	Number of parsimony informative sites	Percentage of parsimony informative sites
c80d5m4p3	15,331	1,303,485	51.15	121,475	9.32	37,722	2.89
c80d5m6p3	9,824	835,710	43.18	85,987	10.29	29,077	3.48
c80d5m8p3	4,039	343,583	33.12	39,877	11.61	14,084	4.10
c80d5m10p3	675	57,406	20.91	7,564	13.18	2,423	4.22
c85d5m4p3	15,499	1,314,995	51.62	109,124	8.30	35,303	2.68
c85d5m6p3	9,766	828,960	43.54	76,353	9.21	26,955	3.25
c85d5m8p3	3,884	329,620	33.37	33,894	10.28	12,689	3.85
c85d5m10p3	618	52,373	21.17	5,984	11.43	2,041	3.90
c90d5m4p3	15,595	1,318,652	52.78	89,781	6.81	29,801	2.26
c90d5m6p3	9,310	787,167	44.29	59,659	7.58	21,828	2.77
c90d5m8p3	3,396	286,979	33.74	23,903	8.33	9,559	3.33
c90d5m10p3	481	40,605	21.38	3,592	8.85	1,357	3.34

Table S8. RAD-seq matrices statistics

CHAPTER 4

No evidence of seamount-driven isolation in deep-sea hydrothermal vent barnacle populations

ABSTRACT

Patterns of spatial genetic population structuring provide insight into the factors that limit dispersal and connectivity of species. Deep-sea hydrothermal vents are the focus of increasing interest for the mining of mineral resources found in polymetallic sulphide deposits. Understanding the genetic diversity and population connectivity of vent species is critical for assessing the potential effects of mining on these ecosystems. Population genetic studies of vent species have mainly focused on mid-ocean ridge systems and back-arc spreading centers. However, vents also occur in active seamounts worldwide. Seamounts are hypothesized to behave as isolated island-like systems, where population connectivity is limited and endemicity is promoted (seamount endemicity hypothesis). In this study, we aim to test this seamount endemicity hypothesis using novel genome-wide restriction-site associated DNA (RAD) sequence data from three hydrothermal vent barnacle species. Comparisons of the genetic diversity and population structuring patterns of barnacle populations from seamounts and spreading ridges revealed patterns of population genetic structuring that do not conform to the predictions from the seamount endemicity hypothesis. The patterns of genetic variation among individuals collected from seamounts and spreading ridges, separated horizontally by hundreds of kilometers and vertically by hundreds of meters, did not reject the null hypothesis of panmixia within each species. These inferences are largely insensitive the to de novo assembly parameters used to identify loci from sequence reads. We suggest that the seamount endemicity hypothesis warrants further testing using high-resolution genetic markers in other vent organisms with differing life history strategies (e.g., brooders) that may limit their dispersal potential, as well as in non-vent organisms, which are not exposed to evolutionary pressures imposed by the dynamic nature of hydrothermal vent systems.

INTRODUCTION

Patterns of spatial genetic population structuring provide insight into the factors that limit dispersal and connectivity throughout a species' range. Empirical evidence has revealed the fundamental importance of intrinsic and extrinsic factors, such as habitat discontinuity (Reitzel *et al.* 2013; D'Aloia *et al.* 2014), depth (Prada & Hellberg 2013), currents (Kelly & Palumbi 2010; White *et al.* 2010), distance (Alberto *et al.* 2011), larval developmental mode (Kelly & Palumbi 2010), and symbiotic associations (Beinart *et al.* 2012), acting as barriers to gene flow in marine species and ultimately driving biodiversity patterns in the ocean. Understanding how populations are spatially and temporally interconnected is critical due to the need for marine biodiversity and ecosystems protection against increasing threats created by human activities (Christensen *et al.* 1996).

Deep-sea hydrothermal vents are the focus of increasing interest for the mining of mineral resources found in polymetallic sulphide deposits (Boschen *et al.* 2013), which form by precipitation during mixing of metal-rich vent fluids with bottom seawater. Deep-sea hydrothermal vents (hereafter simply referred to as vents) host some of the most spectacular and unique ecosystems on earth that thrive on *in situ* primary productivity derived from chemosynthesis. Vents present a sharp contrast to other ecosystems in the deep sea, due to their marked patchiness, extremely steep chemical and thermal gradients over centimeter scales, and relatively high frequency of disturbances given their occurrence on highly dynamic geological settings (Van Dover 2000). Consequently, vent environments present extreme selective pressures, evolutionary speaking, and are characterized by their low biodiversity and high endemicity. These characteristics make vent ecosystems potentially susceptible to disturbances caused by deep-sea mining. Although vent organisms have adapted to the natural dynamics and ephemerality of hydrothermal vents, the potential disturbances from mining are likely to have multiplicative harmful effects at unprecedented scales (Van Dover 2010). Understanding the genetic diversity and population connectivity of vent species is critical to assessments of the potential effects of mining on these ecosystems.

Population genetic studies of vent species from spreading ridge systems commonly reveal patterns of genetic diversity consistent with high gene flow along ridge axes extending for hundreds to thousands of kilometers, yet factors such as depth and ocean currents have been identified as barriers to dispersal at regional scales (see review by Vrijenhoek (2010)). The geomorphology of spreading ridges is known to modify local current regimes and generate significant current flows along ridge axes (Thurnherr *et al.* 2011; Lavelle *et al.* 2012), which are hypothesized to facilitate dispersal among spatially separated vent fields over long distances. Deep-sea hydrothermal vents also occur on active volcanic seamouts located

on hotspots or island arcs. However, the population structuring and genetic connectivity dynamics in these systems are much less understood.

Seamounts are hypothesized to behave as isolated island-like systems, where population connectivity, and therefore gene flow, are limited (commonly referred to as the 'seamount endemicity hypothesis', de Forges et al. 2000). Such isolation might arise due to the patchiness of the seamounts habitats (particularly of active seamounts as most seamounts are inactive), combined with larval behaviors, physical oceanographic phenomena, such as Taylor columns, that could cause larvae retention around seamount summits (Wilson & Kaufmann 1986; Mullineaux & Mills 1997; de Forges et al. 2000; Metaxas 2011), or a combination of these factors. The seamount endemicity hypothesis predicts that there will be relatively high levels of endemicity of genetic variants and significant structure among populations from different seamounts. Consistent with this hypothesis Smith et al. (2004) found heterogeneity in allozyme frequencies and significant genetic differentiation between populations of *Bathymodiolus* vent mussels from two active seamounts of the Kermadec arc. Contrastingly, Tunnicliffe et al. (2010) and Watanabe et al. (2005) found significant differentiation in mitochondrial sequences between volcanic arc basins, but no differentiation among seamount populations of vent flatfish and barnacle species within each arc basin. Other studies in non-vent seamount fauna have found similar patterns of absence of population structuring among seamount populations of deep-sea corals (Thoma et al. 2009), clams (Clague et al. 2012), and ophiuroids (Cho & Shank 2010), when examining mitochondrial markers. However, the mitochondrion has a intermediate mutation rate compared to other autosomal markers such as microsatellites and single nucleotide polymorphisms (SNPs), and mounting evidence indicates that significant amounts of genetic variability in populations and species can go undetected based solely on mitochondrial data (Emerson et al. 2010; Herrera & Shank Chapter 5). For example, a study on a species of Patagonian toothfish revealed significant genetic structure between eastern and western Southern Ocean populations using data from microsatellites, when a lack of genetic structure had been previously concluded using mitochondrial data only (Rogers et al. 2006). Therefore, interpretations of nonendemicity in seamounts based on mitochondrial DNA data have been challenged (Baco & Cairns 2012).

In this study we aim to test the seamount endemicity hypothesis in deep-sea hydrothermal-vent fauna using novel data from thousands of genome-wide SNPs obtained from restriction-site associated DNA (RAD) markers (Baird *et al.* 2008). Our approach is to compare the genetic diversity and population structuring patterns of seamount and spreading ridge populations in three hydrothermal vent barnacle species: 1) *Vulcanoleapas scotiaensis* from East Scotia ridge; 2) *Vulcanolepas osheai* from the Kermadec arc; and 3) *Eochionelasmus ohtai* from the Lau Basin. Under the seamount endemicity hypothesis, we

expect to find higher population genetic structuring between seamount populations than between populations from spreading ridges.

METHODS

Barnacle pecimens were collected from deep-sea hydrothermal vents at 2-3 localities each within the East Scotia ridge (E2 and E9 vent sites), the Kermadec arc (Healy and Tangaroa seamounts), and the Lau Basin (Tahi Moana vent site, and the Mata Ua and Founalei South seamounts). Morphological identifications were performed on using stereo-microscopy and species descriptions as references (Table 1, Fig. 1). Species were identified as *Vulcanoleapas scotiaensis, Vulcanolepas osheai*, and *Eochionelasmus ohtai* respectively. Species identifications of representative specimens were confirmed through multilocus phylogenetic analyses by Herrera *et al.* (Chapter 3).

Table 1. Collection information for the specimens used in this study.

Species	Region	Population locality	Collection event	Depth (m)	Lat.	Lon.	Collection ID
Vulcanolepas scotiaensis	East Scotia ridge	E2 vent field	JC042	2700	-56.06	-30.33	WHOI-NEOE2
Vulcanolepas scotiaensis	East Scotia ridge	E9 vent field	JC042	2400	-60.05	-29.93	WHOI-NEOE9
Vulcanolepas osheai	Kermadec arc	Healy volcano	TAN1104/073	1255	-35.01	178.98	NIWA-72638
Vulcanolepas osheai	Kermadec arc	Tangaroa volcano	TAN1206/017	682	-36.32	178.03	NIWA-82121
Eochionelasmus ohtai	Lau Basin	Tahi Moana vent field	TN236/J2-444	2232	-20.68	-176.18	WHOI-13288
Eochionelasmus ohtai	Lau Basin	Founalei South volcano	RR1211/Q4K-323	956	-17.54	-174.58	WHOI-20665
Eochionelasmus ohtai	Lau Basin	Mata Ua volcano	RR1211/Q4K-328	2391	-15.02	-173.79	WHOI-20791

Molecular laboratory methods

Total genomic DNA was purified from specimens by: (1) digesting the tissue in 2% CTAB buffer (Teknova) with proteinase K and RNAse A/T1 (Fermentas) for 1 hour, (2) separating nucleic acids with chloroform: isoamyl alcohol (24:1) (Fermentas) and phenol: chloroform: isoamyl alcohol (25:24:1, Tris buffered at pH 8.0) (Fermentas), (3) precipitating nucleic acids with 100% ethanol (1:1 volume ratio), and (4) washing the precipitate twice with 70% ethanol.

Concentration-normalized genomic DNA was submitted to Floragenex Inc. (Eugene, OR) for library preparation and RAD sequencing. In short, the RAD sequencing method consists of: 1) the digestion of genomic DNA for each individual with a restriction enzyme; 2) ligation of the resulting fragments to sequencing adapters with unique barcodes for each individual; 3) size-selection and enrichment of the
fragments successfully ligated to the adapters; and 4) sequencing via a high-throughput platform. Individual libraries for *Vulcanolepas osehai* and *Eochionelasmus ohtai* were produced from DNA digested with a high-fidelity SbfI restriction enzyme, which is predicted to cut between 5,000 and 15,000 times in the genome of a thoracican barnacle (Table S1) (Herrera *et al.* Chapter 3). Libraries for *Vulcanolepas scotiaensis* were produced from DNA digested with a high-fidelity SgrAI restriction enzyme, which is predicted to cut between 20,000 and 60,000 times in the genome of a thoracican barnacle. The cut-frequency ranges for both SbfI and SgrAI were predicted by the observed recognition sequence frequencies and their estimated probability using trinucleotide composition models in the genome of the crustacean *Daphnia pulex* using the software predRAD (Herrera *et al.* Chapter 2). Ranges of genome size for barnacles were obtained from the Animal Genome Size Database (http://ww.genomesize.com). Barcode tags were 10-base pairs long. Libraries were sequenced by multiplex on an Illumina Hi-Seq 2000 sequencer.



Figure 1. Geographic location of vent barnacle populations for each species. Bathymetry maps show the specific areas where samples have been obtained. Purple labels indicate locations of collections of populations included in this study, but marked to provide context. Yellow labels indicate locations of known populations not included in this study for context. Red dots indicate locations of other active hydrothermal vent sites. Color scale indicates corresponding depths in meters. Bottom right bars indicate distance scale in kilometers. Small maps inset within the bottom right of each subplot indicate region location in a global perspective. Base bathymetry maps were generated using the program GeoMapApp (<u>http://www.geomapapp.org</u>) with data from the Global Multi-Resolution Topography (GMRT) Synthesis (Ryan *et al.* 2009).

Data filtering

Sequence reads were de-multiplexed and quality filtered with the process_radtags program from the package Stacks v1.19 (Catchen *et al.* 2013b). Barcodes and Illumina adapters were excluded from each read, and length was truncated to 70bp (-t 70). Reads with ambiguous bases were discarded (-c). Reads with an average quality score below 10 (-s 10) within a sliding window of 15% of the read length (-w 0.15) were discarded (-q). The rescue barcodes and RAD-tags algorithm was enabled (-r).

Tests of parameters for de novo loci assembly

To explore the effects of parameter choice on *de novo* assemblies of RAD loci using the *denovo_map* pipeline in Stacks, we tested an array of parameters following the guidelines and modified R scripts by Mastretta-Yanes *et al.* (2014). For each species dataset we independently performed multiple *de novo* loci assemblies modifying the following individual parameter values, while keeping default values for the rest: the minimum number of identical raw reads required to create a stack of identical unique sequences for each individual (*-m* 2, 4, 6, 8, 10 for *V. scotiaensis*; and *-m* 2-8, 10, and 12 for *V. osheai and E. ohtai*), the maximum number of stacks at a single locus for each individual (*-max_locus_stacks* 2-5 for *V. scotiaensis*; and *--max_locus_stacks* 2-6 for *V. osheai and E. ohtai*), the number of mismatches allowed between loci for each individual (*-M* 2, 3, 4 for *V. scotiaensis*; and *-M* 2-6, 8 for *V. osheai and E. ohtai*), and the number of mismatches allowed between loci when building the catalog of all loci in a species (*-n* 0-4 for *V. scotiaensis*; and *-n* 0-5 for *V. osheai and E. ohtai*). High-confidence SNP calls in STACKS are performed using a maximum-likelihood framework that accounts for sequencing error and variable depth of coverage among loci (Hohenlohe *et al.* 2010; Catchen *et al.* 2013b). The results from each assembly were loaded onto a MySQL database and indexed through the *load_radtags* and *index_radtags* tools.

We calculated the number of reads and loci coverage per individual from each assembly, and filtered the data to exclude individuals having less than 50% of the mean number of loci per individual and keep only loci present in at least 80% of the individuals. We executed the *populations* program of Stacks after each *de novo* assembly using only individuals and loci that passed the aforementioned filter (whitelist *-W*). To explore the influence of different assembly parameters on population differentiation estimates, we calculated population F_{ST} values from SNPs utilizing a *p_value* filter (*-f*) to keep only significant estimates (α =0.05). To examine the effect of different assembly parameters on the inferred genetic variability within and among populations, we calculated Euclidean distances among individuals from exported SNP data in *plink* format, and performed Neighbor Joining similarity and principal component analyses in R. Optimal *de novo* assembly parameter settings were chosen conservatively, aiming to

maximize the number of SNPs recovered while minimizing the within-population genetic distances, following guidelines by Mastretta-Yanes *et al.* (2014).

Demographic inferences

We performed *de novo* assemblies of RAD loci using optimal parameter settings for each species in Stacks, as explained above. To estimate population genetic descriptive statistics per SNP (genetic diversity π , proportion of polymorphic loci, observed heterozygosity, minor allele frequency, number of private alleles, inbreeding index F_{IS}, and population differentiation index F_{ST}) we executed the *populations* program of Stacks, using only individuals and loci that passed aforementioned filters. We only analyzed loci that were present in all populations of each species (*-p*) and in at least 75% of individuals per population (*-r*). As before, we calculated population F_{ST} values from SNPs utilizing a *p_value* filter (*-f*) correction (α =0.05). We exported SNP data in *plink* format, keeping only one SNP per loci to avoid violating the assumption of independence among markers. To detect possible population structuring we constructed Neighbor Joining similarity dendrograms and performed principal component analyses with data from individuals for each species, as described above. To explore possible non-equilibrium signals in populations we examined frequency distribution plots of minor allele frequencies and inbreeding indices (F_{IS}) in R.

RESULTS

Sequencing results

We generated RAD-seq data for 117 individuals from three species of vent barnacles: 28 individuals from two populations of *Vulcanolepas scotiaensis*, E2 and E9, from the East Scotia rise; 36 individuals from two populations of *Vulcanolepas osheai*, Healy and Tangaroa, from the Kermadec arc; and 53 individuals from three populations of *Eochionelasmus ohtai*, Tahi Moana, Founalei South, and Mata Ua, from the Lau Basin (Table 2). We obtained approximately 3.1 ± 1.5 (mean \pm standard deviation) million reads per individual for *V. scotiaensis*, with individual values ranging from 1.1 to 7.2 million reads. For *V. osheai* we obtained approximately 1.4 ± 0.4 million reads per individual, with individual values ranging from 0.8 to 4.8 million reads. Overall, more than 88% of sequence reads were retained after quality filtering.

Table 2. RAD sequencing results, filtering and *de novo* assembly statistics. N indicates the number of specimens per population. The numbers of raw and retained reads are shown as population mean \pm standard deviation.

Species Population		Sequence IDs	Ν	Raw reads	Retained reads	% reads retained	
Vulcanolepas scotiaensis	E2	NEO_* E2	14	3,531,674±1,490,665	3,058,341±1,086,450	88±6	
Vulcanolepas scotiaensis	E9	NEO * E9	14	2,744,428±1,441,426	2,369,012±1,087,019	88±6	
Vulcanolepas osheai	Healy	72638 *	18	1,254,447±245,024	1,134,690±249,516	90±6	
Vulcanolepas osheai	Tangaroa	82121 *	18	1,500,636±424,632	1,370,937±418,941	91±5	
Eochionelasmus ohtai	Tahi Moana	13288 *	17	2,170,591±906,494	1,978,867±876,981	90±5	
Eochionelasmus ohtai	Founalei South	EoFo *	18	2,438,023±728,551	2,322,302±742,900	95±3	
Eochionelasmus ohtai	Mata Ua	EoMaU_*	18	2,167,746±1,180,479	2,147,705±1,161,193	91±5	

* represents individual IDs

De novo loci assembly parameter examination

De novo loci assembly parameter variation tests produced between 18-55 thousand loci for *V. scotiaensis* containing on average 1.9-2.0 SNPs per locus (Fig. 2). These tests produced 4.1-5.5 and 10.5-13.0 thousand loci, containing on average 2.38-2.43 and 2.17-2.37 SNPs per locus, for *V. scotiaensis* and *E. ohtai*, respectively. Overall, the parameter controlling the minimum number of identical raw reads required to create a stack of identical unique sequences for each individual (-m) had the largest influence on the number of markers obtained, particularly in the dataset from *V. scotiaensis* with a roughly linear decrease rate of 5 thousand loci per parameter unit increase (compared to a decrease rate of nearly 200 and 130 loci per parameter unit increase in *V. osheai* and *E. ohtai*, respectively). The parameter -*M*, which limits the number of mismatches allowed between loci for each individual, also produced a general decrease in the number of loci and SNPs as parameter value increased. The parameter -*n*, which limits the number of mismatches allowed between loci when building the catalog of all loci in a species, produced a similar decreasing effect to parameters -*m* and -*M*, after the number of loci and SNPs peaked at -*n* 1. The parameter -*-max_locus_stacks*, which limits the maximum number of stacks at a single locus for each individual, did not have a substantial effect on the number of loci or SNPs.

All individuals passed the imposed requirement of having more than 50% of the mean number of loci per individual per species. Mean F_{ST} population differentiation values were consistently low, ranging from 0.0008 to 0.0020 (Fig. 3). Parameter *-n* had the greatest influence in the mean F_{ST} values, which tended to increase and plateau with the parameter values (except in *V. osheai*, where mean F_{ST} values oscillated). Increasing values of parameters *-M* and *-m* had opposite effects to parameter *-n*, causing a general drop in mean F_{ST} values. Parameter *--max_locus_stacks* had negligible effects on the estimates of mean F_{ST} values.



Figure 2. *De novo* loci assembly parameter influence on number of obtained loci, SNPs, and genetic distances for each vent barnacle dataset. Species are indicated at the top of each column. **a**) Plots of the value of each of the examined STACKS core parameters against the number of RAD-loci generated. **b**) Plots of the value of each of the examined STACKS core parameters against the number of SNPs generated. Line colors and legend indicate evaluated core parameters: the minimum number of identical raw reads required to create a stack of identical unique sequences for each individual (-*m* in green), the number of mismatches allowed between loci for each individual (-*M* in red); the number of mismatches allowed between loci when building the catalog of all loci in a species (-*n* in

blue); and the maximum number of stacks at a single locus for each individual (--*max_locus_stacks* abbreviated as mls in yellow). c) Box plots of Euclidean genetic distances among individuals, per population, calculated using SNP data from each *de novo* loci assembly parameter examination. Colors and legend indicate evaluated core parameters.

Overall, Euclidean genetic distances among individuals within each population, calculated from SNPs in loci that were present in at least 80% of the individuals for each one of the different *de novo* loci assemblies, were markedly elevated (above 0.8) (Fig. 4). No significant differences in distance distributions were observed among populations or among loci assemblies using different parameters. Neighbor-Joining similarity dendrograms and principal component analyses performed with SNP data from each of the *de novo* loci assemblies were remarkably similar within species, and none of them revealed patterns indicative of population genetic structuring (Fig. 4).

Table 3. Summary statistics of *de novo* RAD stacks assembly. Stacks are defined as clusters of identical reads. N indicates the number of individuals per population. Values per population are shown as mean \pm standard deviation.

Species	Population	Ν	Number of Stacks	Mean coverage depth	S.D. of coverage depth	Maximum coverage depth
Vulcanolepas scotiaensis	E2	14	99,758±10,413	25±5	152±49	28,284±11,158
Vulcanolepas scotiaensis	E9	14	85,349±20,098	23±7	130±56	23,204±13,867
Vulcanolepas osheai	Healy	18	28,533±3,869	38±10	139±34	12,327±3,133
Vulcanolepas osheai	Tangaroa	18	28,993±2,620	44±14	177±62	15,851±5,834
Eochionelasmus ohtai	Tahi Moana	17	28,074±2,449	65±24	324±151	37,552±20,832
Eochionelasmus ohtai	Founalei South	18	27,764±2,524	76±21	386±125	42,658±17,162
Eochionelasmus ohtai	Mata Ua	18	26,504±3,069	71±33	371±198	42,108±25,723

De novo loci assembly with selected parameters

We selected *de novo* assembly parameters conservatively, aiming to prevent pronounced losses of loci and SNPs while ensuring that there were enough reads to make high-confidence SNP calls (-m>3), avoiding the formation of paralogs (-M<4), avoiding calling fixed alleles as separate loci (-n>0), and conforming to biological expectations (-- $max_locus_stacks = 2$, as usually no more than 2 alleles are found in SNP markers of diploid organisms). *De novo* loci assemblies with selected parameters (-m 4 - M3 -n 4 -- $max_locus_stacks 2$) produced approximately 92±17 thousand unique sequence stacks per individual of *V. scotiaensis*, with a mean coverage depth of 24±6x using the restriction enzyme SgrAI (Table 3). The restriction enzyme SbfI was used with both *V. osheai* and *E. ohtai*. Approximately 28±3 thousand stacks per individual were produced for *V. osheai*, with a mean coverage depth of 41±13x (optimal assembly parameters: -m 4 - M 3 - n 3 -- $max_locus_stacks 2$). Lastly, approximately 27±3 thousand stacks per individual were produced for *E. ohtai*, with a mean coverage depth of $73\pm28x$ (optimal assembly parameters: -*m* 5 -*M* 3 -*n* 3 -*max_locus_stacks* 2).



Figure 3. *De novo* loci assembly parameter influence on F_{ST} estimation for vent barnacle datasets. Species are indicated at the top of each subplot. Plots show the value of each one of the examined STACKS core parameters against the estimated mean F_{ST} . Line colors and legend indicate evaluated core parameters: the minimum number of identical raw reads required to create a stack of identical unique sequences for each individual (*-m* in green); the number of mismatches allowed between loci for each individual (*-M* in red); the number of mismatches allowed between loci in a species (*-n* in blue); and the maximum number of stacks at a single locus for each individual (*-max_locus_stacks*, abbreviated as mls, in yellow).

There were 33,507 loci (mean coverage depth of $20\pm6x$) shared among 75% of individuals in all populations of *V. scotiaensis*, of which 28,270 contained one or more SNPs. Significantly fewer markers were obtained from *V. osheai*, with 4,384 loci (mean coverage depth of $34\pm11x$), but a larger fraction, 4,205, contained one or more SNPs. Lastly, there were 9,966 loci for *E. ohtai* (mean coverage depth of $60\pm22x$), of which 9,430 contained one or more SNPs.

Population demographic inferences

Summary population genetic statistics revealed virtually identical levels of genetic diversity among populations of each species – in terms of percentage of polymorphic sites, mean observed heterozygosity and mean nucleotide diversity (Table 4). Overall, *V. scotiaensis* had the greatest genetic diversity, followed by *V. osheai* and *E. ohtai*. These patterns were maintained when summary statistics were calculated from variant positions alone, and from variant and fixed positions combined.

Table 4. Population genetic statistics calculated only from variant positions, and from both variant and fixed positions. Values indicate means \pm standard deviation

Variant positions

Population	Private alleles	Variant sites	% polym. sites	Major allele frequency	Observed heterozygosity	Nucleotide diversity (π)	Fıs
E2	21,327	55,411	64.63	0.9403 ± 0.0949	0.0880 ± 0.1311	0.0981 ± 0.1257	0.0504±0.2356
E9	19,600	55,415	61.51	0.9411±0.0938	0.0823 ± 0.1245	0.0973 ± 0.1261	0.0667±0.2544
Healy	3,446	10,274	63.29	0.9419 ± 0.0970	0.0671 ± 0.1034	0.0935 ± 0.1285	0.1263 ± 0.3041
Tangaroa	3,772	10,274	66.46	0.9422 ± 0.0949	0.0723 ± 0.1082	0.0938 ± 0.1257	0.1058±0.2851
Tahi Moana	4,977	23,177	50.83	0.9503 ± 0.0922	$0.0523{\pm}0.0938$	0.0800 ± 0.1241	0.1305±0.3071
Founalei South	4,849	23,167	50.09	0.9507±0.0917	$0.0555 {\pm} 0.0980$	0.0795 ± 0.1241	0.1078±0.2846
Mata Ua	5,274	23,161	52.48	$0.9504{\pm}0.0917$	0.0523±0.0922	0.0798±0.1229	0.1336±0.3106

All positions (variant and fixed)

Population	Private alleles	Sites	% polym. sites	Major allele frequency	Observed heterozygosity	Nucleotide diversity (π)	Fis
E2	21,327	1,728,859	2.07	0.9981±0.0200	0.0028 ± 0.0283	0.0031 ± 0.0283	0.0016±0.0436
E9	19,600	1,728,867	1.97	0.9981±0.0200	0.0026 ± 0.0265	0.0031 ± 0.0283	0.0021±0.0469
Healy	3,446	382,282	1.70	0.9984±0.0173	$0.0018 {\pm} 0.0200$	0.0025 ± 0.0265	0.0034 ± 0.0539
Tangaroa	3,772	382,279	1.79	0.9984±0.0173	0.0019 ± 0.0224	0.0025 ± 0.0265	0.0028 ± 0.0500
Tahi Moana	4,977	853,175	1.38	0.9986±0.0173	$0.0014{\pm}0.0173$	0.0022 ± 0.0245	0.0035 ± 0.0548
Founalei South	4,849	853,168	1.36	0.9987±0.0173	0.0015 ± 0.0173	0.0022 ± 0.0245	0.0029 ± 0.0500
Mata Ua	5,274	853,151	1.42	0.9987±0.0173	$0.0014{\pm}0.0173$	0.0022 ± 0.0245	0.0036 ± 0.0557

Mean pairwise F_{ST} values indicate low differentiation among populations within each species (0.0011 in *V. scotiaensis*; 0.0011 in *V. osheai*; and in *E. ohtai* 0.0018 between Tahi Moana and Founalei S., 0.0021 between Tahi Moana and Mata Ua, and 0.0017 between Founalei S. and Mata Ua). Globally, neither Neighbor-Joining similarity dendrograms nor principal component analyzes of SNP data produced clustering patterns of genetic variation consistent with population genetic structuring within species. In all cases, none of the axes in the principal component analyses explained more than 7% of the observed variance in the SNP data.

Minor allele frequency spectra show that a majority of the alleles in each population have low frequencies (Fig. 5), as expected for stable populations near mutation-drift equilibrium. F_{IS} distributions for all populations were centered on zero indicating random mating and lack of population-sub structuring (Fig. 6). A small fraction of F_{IS} values were positive, including noticeable clusters of values near 1, could be indicative of markers falling on non-recombinant genomic regions, such as sex-determining regions

(Catchen *et al.* 2013a). Both minor allele frequency spectra and F_{IS} distributions were remarkably similar among populations of each species.



Figure 4. Genetic variability of SNP data from individuals for each vent barnacle species. Species are indicated in each column. a) Calibrated Neighbor-Joining dendrograms of genetic similarity from Euclidean distances among individuals per species. b) Plots of the first two axes of genetic variation among individuals per species found through principal component analyses. c) Plots of the third and fourth axes of genetic variation among individuals per species found through principal component analyses. Labels and colors indicate the source population per

species: V. scotiaensis E2 in red and E9 in blue; V. osheai Healy (HE) in red and Tangaroa (TA) in blue; E. ohtai Founalei South (FS) in red, Mata Ua (MU) in blue, and Tahi Moana (TM) in yellow.

DISCUSSION

Effects of de novo loci assembly parameter selection

Consistently with previous analyses by Mastretta-Yanes *et al.* (2014) and Catchen *et al.* (2013b), we found that the core parameters for *de novo* loci assembly in STACKS -m, -n, and -M, which limit the minimum number of identical raw reads required to create a stack of identical unique sequences for each individual, the maximum number of mismatches allowed between loci for each individual, and the maximum number of mismatches allowed between loci when building the catalog of all loci in a species, respectively, had the largest effects on the amount of loci and SNP generated from RAD sequence data. Although in this study we do not have access to a reference genome or sequencing replicates, with which parameter influences on *de novo* loci assembly error rates can be estimated, our analyses indicate that overall the population genetic inferences from datasets derived using a variety of assembly parameters were largely insensitive to parameter variations. This provides high confidence in the presented results, as the patterns of population genetic diversity here identified are most likely the result of true biological and ecological processes in the examined vent barnacle species, rather than methodological artifacts.

No evidence supporting seamount-driven isolation in vent barnacle populations

Altogether, we find no support for the seamount endemicity hypothesis in deep-sea hydrothermal-vent barnacles after examining thousands of genome-wide SNPs obtained from RAD-seq data. The patterns of genetic variation among individuals collected from seamount and spreading ridges, separated horizontally by hundreds of kilometers and vertically by hundreds of meters, did not reject the null hypothesis of panmixia within each species. Contrary to the predictions from the seamount endemicity hypothesis, we did not find higher population genetic structuring between seamount populations than between populations from spreading ridges.

Non-equilibrium processes, such as population expansion and recent colonization, can also lead to a lack of population structuring and apparently high gene flow. Evidence suggestive of non-equilibrium dynamics has been presented in populations of some of vent species, particularly in populations from fast-spreading ridges (Vrijenhoek 2010). Commonly invoked process to explain negative Tajima's D and Fu's F_s values and star-like haplotype networks in mitochondrial data, include bottlenecks and founder events resulting from processes such as catastrophic eruptions, vent formations and disappearances (Won *et al.*

2003; Hurtado *et al.* 2004; Plouviez *et al.* 2009; Teixeira *et al.* 2011; Thaler *et al.* 2011; Beedessee *et al.* 2013; Thaler *et al.* 2014). Nonetheless, minor allele frequency spectra for vent barnacle populations in this study were consistent with mutation-drift equilibrium, as a majority of the alleles in each population have low frequencies (Luikart *et al.* 1998), with a long tail of alleles at intermediate frequencies approaching zero proportions. Therefore, we conclude that the lack of population structuring and high gene flow among examined vent barnacle populations does not seem to be caused by non-equilibrium processes, but rather by sufficient genetic exchange among among populations.



Figure 5. Minor allele frequency spectra from SNP loci for each vent barnacle population, per species. Labels and colors indicate the source population per species: **a**) *V. scotiaensis* – E2 in red and E9 in blue; **b**) *V. osheai* – Healy in red and Tangaroa in blue; **c**) *E. ohtai* – Founalei South in red, Mata Ua in blue, and Tahi Moana in yellow.

Larval development strategies have been long been hypothesized to be determinant of the dispersal potential of a species. Species in the vent barnacle clade A (sensu Herrera et al. (Chapter 3), to which V. osheai and V. scotiaensis belong, have large lecithotrophic naupliar larvae (Tunnicliffe & Southward 2004; Watanabe et al. 2004). E. osheai is known to produce large eggs (Newman et al. 2006), and thus presumably also has lecithotrophic larvae. It is often assumed that the duration in the water column for lecithotrophic larvae is limited by yolk reserves and thus lecithotrophic larvae are expected to have a shorter development time than planktotrophic larvae, and therefore smaller dispersal potential overall. However, significant evidence has accumulated against this hypothesis. For example, results from a colonization study in the East Pacific Rise suggest that the large lecithotrophic larvae can successfully disperse over long distances exceeding 300 kilometers (Mullineaux et al. 2010) (the horizontal distances among barnacle populations in this study range between ca. 100 and 400 kilometers). Additionally, temperatures in the deep ocean are known to reduce metabolic and development rates and therefore enhance dispersal (O'connor et al. 2007). Large yolk reserves, combined with cold sea-bottom temperatures are known to yield pelagic larval durations of up to 100 days in species in the vent barnacle clade A (Yorisue et al. 2013). Taken together, these characteristics provide a potential mechanism for enabling dispersal over long distances in vent barnacles and may partially account for the absence of population genetic structuring.

Local connectivity patterns in the East Scotia Ridge

Consistent with our observations of genetic variability in *V. scotianensis*, recent population genetic studies of other endemic species from hydrothermal vents in the East Scotia Ridge (ESR) –namely the yeti crab *Kiwa* sp. and two gastropod species – show patterns of no population differentiation along the ridge, between the E2 and E9 sites (Roterman 2013). Interestingly, one of the gastropod species, the limpet *Leptodrilus* sp., also occurs in the neighboring Kemp caldera in of the South Sandwich Arc, a region from which *V. scotiaensis* is also known (Rogers *et al.* 2012). Roterman (2013) found significant differentiation between the ridge and Kemp caldera limpet populations, which is separated from the ridge axis by only approximately 96 kilometers to the east, whereas the E2 and E9 sites are separated from each other by approximately 440 kilometers. Nonetheless E2 and E9 sites are a located at similar depths of approximately 2,500 meters, whereas the Kemp caldera vents are approximately 1,000 meters shallower, which could explain the observed differentiation. Testing for population differentiation in *V. scotiaensis*

between the ridge axis and the Kemp caldera would provide a good test for the generality of the observed depth isolation pattern in the limpet species.



Vulcanolepas scotiaensis

Figure 6. Frequency spectra of Wright's inbreeding index (F_{IS}) values from SNP loci for each vent barnacle population, per species. Labels and colors indicate the source population per species: **a**) *V. scotiaensis* – E2 in red and E9 in blue; **b**) *V. osheai* – Healy in red and Tangaroa in blue; **c**) *E. ohtai* –Founalei South in red, Mata Ua in blue, and Tahi Moana in yellow.

Although mean water fluxes in the ESR have an easterly or north-easterly direction, largely influenced by the Antarctic Circumpolar Current (Meredith *et al.* 2008), the local hydrography and flows on the E2, E9

and Kemp vent sites remain unknown. The ESR has an intermediate spreading rate of approximately 62 to 70 millimeters per year (Livermore 2003), which causes the formation of deep axial valleys with high walls that likely modify local current regimes. Mid-ocean ridges with similar characteristics are known to generate significant flows along and on the flanks of the ridge axes (Thurnherr *et al.* 2008; Thurnherr *et al.* 2011; Lavelle *et al.* 2012), thus potentially facilitating dispersal among spatially separated vent fields. Although hydrothermal vents at ridges with intermediate and low spreading rates – like the East Scotia Ridge – occur spatially less frequently along the ridge than vents at faster spreading systems – like the East Pacific Rise – (Baker & German 2004), they tend to be significantly more stable over time, in some cases for thousands to hundreds of thousands of years (Lalou *et al.* 1995; Jamieson *et al.* 2013). Therefore, local vent formations and extinctions in these systems may not be recurrent enough to significantly influence genetic variability patterns in vent fauna. Finally, potential undiscovered vent sites along the East Scotia Ridge axis (German *et al.* 2000) could act as stepping-stones for vent barnacles and other vent species, and thus facilitate dispersal and gene-flow.

Local connectivity patterns in the Kermadec arc

The only previous population genetic study in seamount populations from the Kermadec arc (Smith *et al.* 2004) found heterogeneity in allozyme allelic frequencies and significant genetic differentiation between populations of *Bathymodiolus* vent mussels from two active seamounts of the Kermadec arc, Rumble V and Rumble III, which were separated horizontally by approximately 50 km. However, this same study failed to reproduce the pattern of population structuring between seamount mussel populations when examining mitochondrial DNA sequences; therefore, it is possible that this pattern was an artifact of the allozyme markers. The use of other nuclear markers, such as SNP would clarify this discrepancy. Consistent with our results indicating high-gene flow among barnacle populations from Kermadec arc seamounts, Herrera *et al.* (Chapter 3) examined the phylogenetic relationships of vent barnacles and found no evidence of cryptic differentiation among *V. oseahi* individuals from Brothers, Clarck, Healy or Tangaroa seamounts.

Physical oceanographic processes could facilitate high non-directional dispersal and gene flow among seamounts in the southern Kermadec arc. Recent studies focusing on Brothers (Lavelle *et al.* 2008) and Rumble III seamounts (Stevens *et al.* 2014) indicate that circulation patterns in this region are highly dynamic over time, being largely influenced by isotropic tidal flows and mesoscale eddies, such as the East Cape Eddy. These studies indicate that flows speeds over these seamounts can range between 5 and 10 centimeters per second, which could potentially disperse larvae between 4 and 8 kilometers per day, and more than 130 to 260 kilometers in just a month. Futhermore, Stevens *et al.* (2014) found no evidence

of Taylor column formation on the Rumble III seamount that could cause larval or particle retention around the summits.

The Kermadec arc is a region of high volcanic activity (de Ronde *et al.* 2007), which likely leads to catastrophic local extinctions, and frequent formation and disappearance of vents. Just between 2006 and 2012 there were at least three reported eruptions in active seamounts (Dodge 2009; Watts *et al.* 2012; Jutzeler *et al.* 2014). However, active seamounts usually host multiple vent fields, which are likely not affected by volcanic and venting dynamics simultaneously or at the same rates. We hypothesize that presence of multiple populations on other active neighboring seamounts in the southern Kermadec arc, such as Brothers, Rumble III, and Clark (A. Rowden, M. Clark, and D. Bowden personal communication), combined with potential high dispersal and gene-flow rates, may jointly act as buffers that prevent significant losses of genetic diversity due to vent instability. Samples from additional populations and further comparative studies with multiple taxa would allow us to elucidate more general patterns for vent populations in these seamounts.

Local connectivity patterns in the Lau basin

Consistent with our results from Lau basin populations, Plouviez *et al.* (2013) found no signatures of genetic differentiation or potential population structuring at horizontal scales from hundreds of meters up to 50 kilometers in *Eochionelasmus ohtai* from the Manus Basin, using mitochondrial markers. Although populations from the Manus basin have been considered a potentially different sub-species from the one inhabiting the Lau and North Fiji basins (Newman *et al.* 2006), a recent phylogenetic study by Herrera *et al.* (Chapter 3) found no evidence of genetic differentiation between Manus and Lau basin individuals of *E. ohtai.* Similar signatures of panmixia have been found in shrimp (*Chorocaris* sp. 2) and gastropod snail (*Ifremeria nautilei*) vent species within the Manus and Lau basins (Thaler *et al.* 2011; Thaler *et al.* 2014). Contrastingly Thaler *et al.* (2014) found significant structuring in populations of the squat lobster *Munidopsis lauensis* within Manus basin, which the authors attributed to potential larvae behaviors and water flow directionality.

In addition to the populations included in this study, *E. ohtai* is known from vents in other active systems in the Lau Basin, including the Eastern Lau Spreading Center and the North-East Lau Spreading Center (Herrera *et al.* Chapter 3), which can serve as dispersal stepping-stones for dispersal throughout the basin. The Lau basin has an extremely diverse and active geologic setting, with multiple isolated back-arc spreading ridges, axial volcanoes and active arc and back-arc seamounts hosting hydrothermal vent communities. Volcanic eruptions that create and destroy hydrothermal vent communities are also

common in this region (Resing *et al.* 2011; Bohnenstiehl *et al.* 2014; Embley *et al.* 2014). However, as discussed earlier, we suggest that the presence of well-connected populations within a metapopulation in some vent species can act as buffers that prevent significant losses of genetic diversity due to vent instability. Ocean circulation at the scale of months in the Lau basin is dominated by isotropic flows caused by eddies, reaching speeds of up to 10 centimeters per second (Speer & Thurnherr 2012), which could help explain the lack of directionality or differentiation observed in the genetic pool of *E. otahi* and other vent species within the basin.

CONCLUSIONS

Comparative data from three species of barnacles from deep-sea hydrothermal vents revealed patterns of genetic variation inconsistent with the seamount endemicity hypothesis. These results reinforce the idea that within-basin structuring is rare while between-basin structuring is common, lending insight into the scale at which vent populations are structured. Both the Kermadec Arc and Lau Basin are areas of high interest for polymetallic sulphide mining (Boschen *et al.* 2013). Although these results indicate that populations of vent barnacles in the examined populations could be resilient to potential disturbance from local mining restricted to focal sites, we suggest that further multi-species and time-series studies in these and other target and neighboring populations should take place in order to better assess the potential impacts of this extractive activity. This study constitutes the first use of genome-wide SNP data to examine patterns of population genetic structuring and connectivity among populations of vent species. We suggest that the seamount endemicity hypothesis warrants further testing using high-resolution SNP data in other vent organisms with differing life history strategies (e.g., brooders) that may limit their dispersal potential, as well as in non-vent organisms, which are not exposed to evolutionary pressures imposed by the dynamic nature of hydrothermal vent systems.

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SUPPORTING INFORMATION

Table S1. Predictions of number of RAD-tags in thoraciacan barnacles using SbfI. Data for *Daphnia pulex* obtained from the U.S. National Center for Biotechnology Information (NCBI) WGS database. Observed frequency of recognition sequences and calculated probability based on a trinucleotide genome composition model were generated following the methodology described by Herrera *et al.* (Herrera *et al.* 2014). Data for known barnacle genome sizes obtained from the Animal Genome Size Database (http://ww.genomesize.com). C-value is the amount of DNA in picograms in the nucleous, where the genome size in Mbp = 978 x C-value

Species	Common name	C-value	Genome size (Mbp)	Observed frequency of SbfI recognition sites per bp	Probability of SbfI recognition site per bp		
Genebank WGS Daphnia pulex	Water flea		158.61	7.48E-06	5.90E-06		
Species	Common name	C-value	Genome size (Mbp)	Predicted number of SbfI recognition sites based on <i>D.</i> <i>pulex</i> observed frequency	Predicted number of SbfI RAD-tags based on <i>D.</i> <i>pulex</i> observed frequency	Predicted number of SbfI recognition sites based on <i>D. pulex</i> trinucleotide genome composition probability	Predicted number of SbfI RAD-tags based on <i>D.</i> <i>pulex</i> trinucleotide genome composition probability
Animal Genome Size	e Database						
Balanus amphitrite	Striped barnacle	0.74	723.72	5,411.68	10,823.35	4,267.22	8,534.43
Balanus amphitrite	Striped barnacle	1.4	1,369.20	10,238.31	20,476.61	8,073.11	16,146.22
<u>Balanus cariosus</u>	Thatched barnacle	1.4	1,369.20	10,238.31	20,476.61	8,073.11	16,146.22
Balanus eburneus	Ivory barnacle	1.26	1,232.28	9,214.48	18,428.95	7,265.80	14,531.60
Chthalamus sp.	Acorn barnacle	1.23	1,202.94	8,995.08	17,990.17	7,092.80	14,185.61
Unknown sp.	Goose barnacle	1.46	1,427.88	10,677.09	21,354.18	8,419.10	16,838.20
Mitella polymerus	Pacific goose barnacle	0.9	880.20	6,581.77	13,163.54	5,189.86	10,379.71
Tetraclita rubescens	Volcano barnacle	2.6	2,542.80	19,014.00	38,027.99	14,992.92	29,985.84

CHAPTER 5

RAD sequencing enables unprecedented phylogenetic resolution and objective species delimitation in recalcitrant divergent taxa

ABSTRACT

Species delimitation is problematic in many taxa due to the difficulty of evaluating predictions from species delimitation hypotheses, which chiefly relay on subjective interpretations of morphological observations and/or DNA sequence data. This problem is exacerbated in recalcitrant taxa for which genetic resources are scarce and inadequate to resolve questions regarding evolutionary relationships and uniqueness. In this case study we demonstrate the empirical utility of restriction site associated DNA sequencing (RAD-seq) by unambiguously resolving phylogenetic relationships among recalcitrant octocoral taxa with divergences greater than 80 million years. We objectively infer robust species boundaries in the genus Paragorgia, which contains some of the most important ecosystem engineers in the deep-sea, by testing alternative taxonomy-guided or unguided species delimitation hypotheses using the Bayes factors delimitation method (BFD*) with genome-wide single nucleotide polymorphism data. We present conclusive evidence rejecting the current morphological species delimitation model for the genus Paragorgia and indicating the presence of cryptic species boundaries associated with environmental variables. We argue that the suitability limits of RAD-seq for phylogenetic inferences in divergent taxa cannot be assessed in terms of absolute time, but depend on taxon-specific factors such as mutation rate, generation time and effective population size. We show that classic morphological taxonomy can greatly benefit from integrative approaches that provide objective tests to species delimitation hypothesis. Our results pave the way for addressing further questions in biogeography, species ranges, community ecology, population dynamics, conservation, and evolution in octocorals and other marine taxa.

INTRODUCTION

Species delimitation is problematic in many taxa due to the difficulty of evaluating predictions from species delimitation hypotheses derived using different species concepts. Species concepts set particular expectations of the properties used to support species delimitations (De Queiroz 2007). For example, the classic biological species concept requires intrinsic reproductive isolation between heterospecific organisms and interbreeding among homospecific organisms resulting in viable and fertile descendants (Mayr 1942; Dobzhansky 1970). In many cases, if not the majority, it is difficult to evaluate behavioral, reproductive, and ecological properties due to technical limitations of field or laboratory work, which largely determine the kind of observations and data that can be obtained. In these cases researchers conventionally rely on morphological observations and/or DNA sequence data to generate species delimitation hypotheses.

Although there have been significant attempts at developing statistical methods to objectively identify species-diagnostic morphological discontinuities (e.g., Zapata & Jimenez 2012), most species delimitations continue to be performed subjectively based on assessments made by specialized taxonomists. Molecular phylogenetic analyses of DNA sequences provide an independent way to test these species delimitation hypotheses utilizing a variety of methods, ranging from variability thresholds of barcode sequences (Hebert *et al.* 2003), to probabilistic coalescent-based model methods (Pons *et al.* 2006; Yang & Rannala 2010; Fujisawa & Barraclough 2013; Grummer *et al.* 2014). These molecular methods rely on informative DNA sequence markers, and in many cases on resolved phylogenies.

The sub-class Octocorallia (Phylum Cnidaria), which includes animals known as gorgonians, sea pens, and soft corals, is an example of a recalcitrant group where species delimitations are problematic. Octocorals are predominantly a deep-sea group (Cairns 2007; Roberts & Cairns 2014) and therefore are extremely difficult to observe and collect. Classic morphology-based species delimitation and identification in this group is arduous for non-specialists, and challenging to replicate among taxonomists (Daly *et al.* 2007; McFadden *et al.* 2010b). Variations in octocoral colony architecture and micro-skeletal structures – sclerites – are used as species diagnostic characters (Bayer 1956). However, studies over the last 15 years have shown that in many cases species delimitations and systematics based on these morphological traits keep little to no correspondence with the patterns of genetic diversity and relatedness inferred using mitochondrial and ribosomal DNA sequence markers (McFadden *et al.* 2006; Clark *et al.* 2007; France 2007; Dueñas & Sánchez 2009). A confounding factor when analyzing mitochondrial DNA markers is the fact that anthozoans, including octocorals, have slow rates of sequence evolution relative to

other metazoans (Shearer *et al.* 2002; Hellberg 2006). Furthermore, octocoral mitochondrion is unique among eukaryotes by having a functional DNA mismatch repair gene — mtMutS — which presumably is responsible for the extremely low sequence variability observed in this group (Bilewitch & Degnan 2011). Traditional molecular markers have thus been remarkably insufficient to resolve relationships at all taxonomic levels within the octocorals (Berntson *et al.* 2001; France *et al.* 2002; Mcfadden *et al.* 2004; Smith *et al.* 2004; Thoma *et al.* 2009; Duefas *et al.* 2014). Alternative nuclear markers, such as the ITS2 and *SRP54* have been used to examine interspecific and intraspecific relationships (Aguilar & Sánchez 2007; Concepcion *et al.* 2007; Grajales *et al.* 2007; Herrera *et al.* 2010); however, their application and impact has been limited due to issues regarding intragenomic variability (Sanchez & Dorado 2008) and low sequencing reliability (Mcfadden *et al.* 2010a). These long-standing technical problems have caused fundamental questions in octocorals regarding species differentiation, systematics, diversity, biogeography, and species ranges to remain unanswered.

Technological developments in next-generation sequencing platforms and library preparation methodologies have made genomic resources increasingly accessible and available for the study of nonmodel organisms, thus offering a great opportunity to overcome the difficulties inherent to the use of traditional sequencing approaches. One of these methodologies is restriction-site-associated DNA sequencing (RAD-seq), which combines enzymatic fragmentation of genomic DNA with high-throughput sequencing for the generation of large numbers of markers (Baird *et al.* 2008). RAD-seq has shown great promise to resolve difficult phylogenetic, phylogeographic, and species delimitation questions in diverse groups of eukaryotes (Emerson *et al.* 2010; Nadeau *et al.* 2012; Wagner *et al.* 2012; Eaton & Ree 2013; Jones *et al.* 2013; Cruaud *et al.* 2014; Escudero *et al.* 2014; Hipp *et al.* 2014; Leache *et al.* 2014; Herrera *et al.* 2014). The number of orthologous restriction sites that can be retained across taxa, which decreases as divergence increases, limits the usefulness of RAD-seq for these kinds of studies. *In silico* studies in model organisms indicate that RAD-seq can be used to infer phylogenetic relationships in young groups of species (up to 60 million years old), such as *Drosophila* (Rubin *et al.* 2012; Cariou *et al.* 2013; Seetharam & Stuart 2013); however, the real limits of this technique have not been significantly explored.

In this study we aim to empirically explore the limits of RAD-seq to solve questions in phylogenetics and species delimitation. We focus on the recalcitrant *Anthomastus-Corallium* clade of octocorals (sensu McFadden *et al.* 2006) to test the utility of RAD-seq to resolve phylogenetic relationships among divergent taxa, and to infer objective species boundaries. Corals in the *Anthomastus-Corallium* clade (hereafter referred as the AC clade) are among the most conspicuous, widely distributed, and ecologically

important benthic invertebrates in deep-water ecosystems (Roberts *et al.* 2009; Wating *et al.* 2011). This clade is constituted by more than 100 species defined morphologically, divided in 10 genera, and three families (World Register of Marine Species at <u>http://www.marinespecies.org</u> accessed on 2014-10-10), spanning a divergence time of over 100 million years (Ardila *et al.* 2012; Herrera *et al.* 2012). However, species delimitations and phylogenetic relationships in this clade, as in other octocorals, are controversial and conflictive (Herrera *et al.* 2010; Ardila *et al.* 2012; Herrera *et al.* 2012). Many of the species in this group are considered species indicators of Vulnerable Marine Ecosystems (e.g. ICES 2013), with some of them considered endangered (CITES 2014). Accurate species identifications, as well as complete inventories and knowledge of species ranges, are therefore critical to ensure the effectiveness and appropriateness of conservation and management policies.

RESULTS

Morphological species identifications

Using current species descriptions, colony observations, and scanning electron microscopy of sclerites, we identified a total of 12 putative morphological species among the 44 examined specimens from the AC clade (Table S1). These species correspond to the genera *Paragorgia (P. arborea, P. stephencairnsi, P. johnsoni, P. maunga, P. alisonae, P. kaupeka, and P. coralloides)* and *Sibogagorgia (S. cauliflora)* of the family Paragorgiidae; *Hemicorallium (H. laauense-imperiale)* and *Corallium* of the family Coralliidae; and *Anthomastus* and *Heteropolypus* of the family Alcyoniidae.

Octocorals are amenable to RAD sequencing

We generated a dense genome-wide set of genetic markers from the 44 AC clade specimens via RAD sequencing, using the 6-cutter restriction enzyme PstI, and used them to perform phylogenetic inferences and species delimitation analyses. We obtained roughly 3.9 ± 1.4 million reads (average \pm standard deviation) per individual, of which $74.3 \pm 8.1\%$ were retained after stringent quality filtering steps (Table S2).

Optimization of RAD-loci clustering parameters

To examine the sensitivity of the phylogenetic inference to the clustering parameters used to identify loci and create nucleotide matrices in the program pyRAD (Eaton 2014), we investigated different combinations of clustering thresholds (c 0.80, 0.85 and 0.90) and minimum number of taxa per locus (m 4, 6, and 9) in a reduced **'backbone'** matrix (hereafter matrix names will be highlighted in bold) containing one individual from each of the 12 morphological species. The 9 resulting **backbone** matrices ranged in the total number of loci per matrix from approximately 9 to 60 thousand loci, increasing dramatically as the minimum number of taxa per locus was reduced (Table S3). In contrast, the different clustering thresholds did not have a significant effect on the total number of loci, but rather on the number of variable sites and, most importantly, on the number of phylogenetically informative sites (Table S3). Each resulting **backbone** matrix analyzed in RAxML (Stamatakis 2006) produced identical strongly-supported tree topologies (Fig. S1). We selected c 0.80 (80% similarity among sequences) and m 9 (minimum coverage of taxa per locus of 75%) as the optimal combination of loci-clustering parameters because they minimized the proportion of missing data (0.20) in the matrix while maximizing the fraction of variable sites that were phylogenetically informative (0.24) (Table S3). The proportion of shared loci among individuals of Paragorgiidae and Coralliidae, lineages whose split has been estimated to be between 80-150 million years ago (Ardila *et al.* 2012; Herrera *et al.* 2012), was remarkably high (70-80%) (Fig. 1).



Figure 1. Proportion of loci shared among individuals of the AC clade in the optimal backbone matrix (c 0.80, m 9). Each family is indicated with a different color: red for Paragorgiidae; blue for Coralliidae; and yellow for

Alcyoniidae. Black-filled circles represent the proportion of the total number of loci shared among individuals. Redfilled circles represent the proportion of the total number of loci present in each individual. Circle scale shows the number of loci represented by 1.0 and 0.5 circle sizes. Black vertical bars represent the average proportion of loci shared by each individual. Phylogenetic tree was inferred with RAxML. Stars on the tree represent branch bootstrap support of 100. Smaller bootstrap support values are indicated with numbers. This figure was generated with the package RADami (Hipp *et al.* 2014).

We used the selected optimal loci-clustering parameters to generate the '**PHYLO**' matrix, containing the sequence data of all the 44 octocoral specimens. The use of the parameter value c 0.80 yielded approximately 71 ± 15 thousand loci – with a minimum depth of coverage of 5x and after filtering for paralogs– per specimen (mean depth of clusters used in loci construction was $23 \pm 8x$) (Table S4). The **PHYLO** matrix contained a total of 5,997 loci that contained data for at least 75% of the specimens (after a second paralog removal). There were 85,293 variable sites in this matrix, of which 53,150 were phylogenetically informative.

RAD-seq data support a fully resolved phylogeny

The phylogenetic analysis of the **PHYLO** concatenated RAD-seq matrix produced a completely resolved evolutionary tree of the AC specimens (Fig. 2). In general, all branches were supported by high (greater than 95) bootstrap values, except for the one supporting the clade of *P. johnsoni*, *P. alisonae*, and *P. maunga*. Each one of the morphologically identified families, genera, and species in this dataset were monophyletic. The branching pattern of the tree is consistent with an expected transition between coalescent processes among species and genera (long deep branches), and population processes within species (short shallow branches).

The topology of the tree obtained with a traditional '**mitochondrial**' matrix (711 base pairs of the *mtMutS* gene containing 130 variable sites, of which 101 were phylogenetically informative) was incongruent with the **PHYLO** tree (Fig. 2). The **mitochondrial** tree indicated a well-supported alternative divergence order for *P. coralloides* and *P. kaupeka* in the *Paragorgia* clade. In addition, the families Paragorgiidae (bubblegum corals) and Coralliidae (precious corals) were not monophyletic. The bubblegum coral genus *Sibogagorgia* appeared more closely related to the precious corals than to the other bubblegum coral genus *Paragorgia*, and the genera *Corallium* and *Hemicorallium* did not form a clade. However, these alternative relationships were not significantly supported by the bootstrap analysis,. Indeed, a substantial proportion of branches on the **mitochondrial** tree were poorly supported (bootstrap values smaller than 80%).



Figure 2. Phylogenetic trees of the AC clade based on RAD-seq and mitochondrial data. Left tree based on the RAD-seq concatenated PHYLO matrix. Right tree based on the mtMutS mitochondrial matrix. Each family is indicated with a different branch color: blue red for Paragorgiidaea; blue for Corallidae; and yellow for

Alcyoniidae. Phylogenetic trees were inferred with RAxML. Stars on the trees represent bootstrap support of 100. Smaller bootstrap values are indicated in numbers. Scale bar indicates substitutions per site.

RAD-seq data reveal cryptic genetic diversity

Branch-length differences among individuals, as well as well-supported sub-clades, revealed intraspecific genetic diversity that was undetected by the **mitochondrial** matrix. Two sub-clades were revealed by the phylogenetic analysis of the **PHYLO** matrix in the *P. arborea* and *P. stephencairnsi* clades. The sub-clades in *P. arborea* correspond to a pattern of segregation by geographic location with specimens from the north Pacific in one sub-clade, and specimens from the south Pacific and north Atlantic in the other. Contrastingly, the sub-clades in *P. stephencairnsi* correspond to a pattern of segregation by depth with specimens collected shallower than 350m in one sub-clade, and specimens collected deeper than 1000m in the other.

Current morphological species delimitation is rejected

To evaluate the utility of RAD-seq to perform objective species delimitations in octocorals we focused on specimens the genus *Paragorgia* as it was the best-sampled taxon in our dataset, both in terms of geographic representation and number of morphological species. We used the Bayes Factor Delimitation method with genomic data (BFD*) (Leache *et al.* 2014), which allows for the comparison of conflictive species delimitation models in an explicit multispecies coalescent framework using genome-wide single nucleotide polymorphism (SNP) data. We calculated marginal likelihoods of taxonomy-guided and taxonomy-unguided species delimitation models from a matrix of unlinked SNPs including only specimens of *Paragorgia* ('PARAGORGIA' matrix containing 1,203 SNPs present in all individuals). We compared the marginal likelihood estimates of alternative species delimitation models to the null model 'morphid', which is based on current morphological species descriptions, using Bayesian factors.

The null model, **morphid**, was rejected in favor of alternative species delimitation models for *Paragorgia* (Fig. 3) (**morphid** was ranked 7th among 10 evaluated models in terms of the marginal likelihood estimate). The '**PABSTE'** model, which proposes 9 species based on the 7 morphological species in the dataset plus splits corresponding to the sub-clades in *P. arborea* and in *P. stephencairnsi*, received decisive support from Bayes factors as the best species delimitation model. The taxonomy-unguided model '**geo'** – which splits the specimens based on the geographic location where they were collected – and the models proposed by the Poisson tree processes (PTP) method based on the **mitochondrial** data matrix, were the lowest ranked and most strongly rejected models overall.

		Species models									
		Taxonomy-informed Taxonomy-independent									ent
		morphid	PAB	STE	PAB STE	split PAB	mIPTI	P bPTP mt	mIPTI rad	bPTP rad	geo
*	2036	S. cauliflora	-	-	-	-	-	-	-	-	-
	•1122230	S.cauliflora	-	-	-	-	-	-	-	-	-
*	12016	P. coralloides	1	1	1	1	8	1	11	8	1
	L 06390	P. coralloides	1	1	1	1	7	1	12	9	1
	84804	P. kaupeka	2	2	2	2	1	17	1	1	2
	82260	P. kaupeka	2	2	2	2	1	16	1	1	2
		P. kaupeka	2	2	2	2	1	19	1	1	2
יוו	72152	P. kaupeka	2	2	2	2	1	20	1	1	2
	* TC1603	P. arborea	3	3	3	3	2	10	7	6	1
* *	L .41870	P. arborea	3	3	3	4	2	10	8	7	2
	1007340	P. arborea	4	3	4	5	2	13	6	5	3
	* 0672	P. arborea	4	3	4	5	2	14	6	5	3
×	L-1601	P. arborea	4	3	4	5	2	14	6	5	3
	64980	P. maunga	5	4	5	6	4	3	3	3	2
3	66097	P. alisonae	6	5	6	7	5	4	4	4	2
	20136	P. johnsoni	7	6	7	8	3	2	2	2	1
	*1007316	P. stephencairns	i 8	7	8	9	6	23	9	12	3
1	L 2344	P. stephencairns	i 8	7	8	9	6	9	10	13	3
*	C05	P. stephencairns	<i>i</i> 8	8	9	10	6	24	5	10	3
	FOC25	P. stephencairns	i 8	8	9	10	6	11	5	11	3
*	C03	P. stephencairns	i 8	8	9	10	6	15	5	11	3
	Agam	P. stephencairns	i 8	8	9	10	6	22	5	11	3
	FOC26	P. stephencairns	i 8	8	9	10	6	5	5	11	3
	·····C102	P. stephencairns	i 8	8	9	10	6	8	5	11	3
	·····FOC5	P. stephencairns	i 8	8	9	10	6	24	5	11	3
0.01	1124300	P. stephencairns	i 8	8	9	10	6	7	5	11	3
	101010	P. stephencairns	i 8	8	9	10	6	7	5	11	3
	·····C104	P. stephencairns	i 8	8	9	10	6	12	5	11	3
	C02	P. stephencairns	i 8	8	9	10	6	18	5	11	3
	·····C04	P. stephencairns	i 8	8	9	10	6	23	5	11	3
	FOC30	P. stephencairns	i 8	8	9	10	6	21	5	11	3
	·····C100	P. stephencairns	i 8	8	9	10	6	6	5	11	3
	C101	P. stephencairns	i 8	8	9	10	6	11	5	11	3
Number o	f species	7	8	8	9	10	8	24	12	13	3
Marginal L	ikelihood	-6240	-6221	-6039	-6014	-6028	-6251	-6243	-6050	-6045 -	10296
-	Rank	7	6	3	1	2	9	8	5	4	10
Bayes Factor		-	-38	-401	-451	-424	23	7	-379	-388	8113

Figure 3. Species delineation hypotheses for *Paragorgia*. Table shows the different species delimitation models for *Paragorgia* evaluated with the BFD* method and their results. *Sibogagorgia* was included as outgroup to root the inferences for *Paragorgia*. Each row indicates a different specimen. Each column indicates a different species

delimitation model. The first column, model morphid, indicates the species identifications based on morphology. For all other models, numbers indicate the species assignments. Bottom rows show the total number of species proposed, the marginal likelihood estimate, and rank for each model. The Bayes factor comparisons were calculated with respect to the null morphid model. Phylogenetic tree on the left, shown only for visual reference, was inferred with the RAD-seq concatenated PARAGORGIID matrix in RAxML. Each genus is indicated with a different branch color: pink for *Sibogagorgia*; and dark red for *Paragorgia*. Stars on the trees represent bootstrap support of 100. Smaller bootstrap values are indicated in numbers. Scale bar indicates substitutions per site.

Concatenated and coalescent species tree analyses are congruent

The topology of the species tree inferred using the SNP **PARAGORGIA** matrix was entirely congruent with the topology generated by the maximum likelihood phylogenetic analysis of the concatenated sequence matrices (Fig. 4). The species tree analysis also greatly improved support for the clade of *P. johnsoni*, *P. alisonae*, and *P. maunga*. The posterior distribution of species trees indicated a small fraction of conflictive topologies concentrating in this region of the tree.

DISCUSSION

RAD sequencing enables unprecedented phylogenetic resolution

Our analyses of RAD-seq data provide a robust phylogenetic hypothesis for the recalcitrant octocorals in the *Anthomastus-Corallium* clade, a result never achieved before. Moreover, this study, together with the work by Pante *et al.* (2014) in the octocoral genus *Chrysogorgia*, constitute the first applications of RAD-sequencing for phylogenetics and species delimitation in cnidarians. Only a handful of previous studies, using traditional mitochondrial data and the ITS2 and 28S nuclear markers, have attempted to evaluate phylogenetic relationships in the octocoral AC clade (Herrera *et al.* 2010; Ardila *et al.* 2012; Brockman & McFadden 2012; Herrera *et al.* 2012; McFadden & van Ofwegen 2013; Uda *et al.* 2013; Figueroa & Baco 2014). These studies find support for the monophyly of the genus *Paragorgia*, the family Coralliidae, and the sister relationship between the Paragorgiidae and Coralliidae. However, those data do not provide enough phylogenetic resolution to infer the evolutionary relationships among many of the putative morphological species. Furthermore, significant incongruences between mitochondrial and nuclear ITS2 gene trees from AC taxa have been documented (Herrera *et al.* 2010). Here we reproduce similar incongruences when comparing the trees inferred from mitochondrial and RAD-seq datasets (Fig. 2). Likewise, Pante *et al.* (2014) documented marked incongruence between trees inferred from mitochondrial and RAD-seq data in *Chyrsogorgia.* These observations suggest that processes that can

cause gene tree heterogeneity, such as incomplete lineage sorting and horizontal gene transfer (Maddison 1997; Edwards 2009), may be more prevalent in octocorals than previously recognized.



Figure 4. Species tree of *Paragorgia*. This claudogram illustrates the posterior distribution of species trees inferred with SNAPP based on the best species delimitation model PABSTE. High color density is indicative of areas in the species trees with high topology agreement. Different colors represent different topologies. The maximum clade credibility species tree is shown with thicker branches. Trees with the same topology as the maximum clade credibility species tree are colored in red. Trees with different topologies are colored green or blue. With the exception of the branch leading to the clade of *P. johnsoni*, *P. maunga*, and *P. alisonae*, which has a posterior probability of 0.87, all interior branches have posterior probabilities of 1.0.

All of our analyses based on RAD-seq matrices – varying in taxon coverage, degree of divergence among taxa, proportion of missing data, number of loci, and analysis type (concatenated or species tree) – produced completely congruent trees, which together provide extremely high confidence on the phylogenetic hypothesis inferred for the octocoral AC clade (Figs. 1, 2 and 3). Consequently, we suggest that single marker gene trees in octocorals, particularly from the mitochondria, should not be considered as robust hypotheses of true species phylogenies on their own, without further validation by multiple

informative and independent nuclear loci. We urge systematists to be conservative when making taxonomic rearrangements based on inferences from single-marker data alone.

RAD-seq data is suitable for phylogenetic inference in divergent taxa

Contrary to the currently accepted idea that RAD-seq data are only suitable for taxa with divergence times younger than 60 million years (MY) (Rubin *et al.* 2012), we demonstrate their suitability well beyond this age threshold. Remarkably, we were able to confidently resolve phylogenetic relationships among genera from different families diverging by at least 80 MY in the AC clade. The split between the families Paragorgiidae and Coralliidae has been dated, using coralliid fossils, to be between 80-150 MY old (Ardila *et al.* 2012; Herrera *et al.* 2012). Park *et al.* (2012) estimated the age of the most recent common ancestor of the Coralliidae at approximately 50 MY (25-100 MY 95% confidence region), using independent enidarian fossils for molecular clock calibration. The split with the genera *Anthomastus* and *Heteropolypus* is likely older than 100 MY. It is without question that, due to evolution at restriction sites, the number of RAD loci among taxa for which orthology can be established decreases rapidly as divergence increases. However, we suggest that the suitability limits of RAD-seq for phylogenetics in divergent taxa cannot be assessed in terms absolute time, but depend on taxon-specific factors such as mutation rate, generation time and effective population size.

Bioinformatic studies addressing the issue of extent of the suitability of RAD-seq for phylogenetic inference have focused mainly on *Drosophila* as study model (Rubin *et al.* 2012; Cariou *et al.* 2013). Longer generation times and lower metabolic rates in taxa like deep-sea corals, relative to those in organisms like *Drosophila*, could cause a reduction in mutation rates (see review by Baer *et al.* (2007)), which may in turn decrease the evolutionary rates at restriction sites and allow for phylogenetic inferences using RAD-seq in situations of deeper divergence. Consistent with this hypothesis, we observe a nucleotide diversity (π) calculated across all octocoral specimens from the **PHYLO** matrix of 0.012 ± 0.002 (considered a minimum since RAD-seq can underestimate diversity (Arnold *et al.* 2013); see Table S5 and Table S6 for individual values), which is significantly lower than the nucleotide diversity in many of the *Drosophila* species included in the bioinformatic studies by Cariou *et al.* (2013) and Rubin *et al.* (2012). Nonetheless, there are other important factors known to influence genetic diversity across species – and likely the evolutionary rate as well. These factors include the effective population size, selection, habitat kind, geographic range, and mating system (Leffler *et al.* 2012). To sum up, we argue that RAD-seq can be successfully used to infer phylogenetic relationships in certain taxa with deeper divergences than previously suggested. This is particularly true when the number of RAD loci is maximized through

the choice of restriction enzymes with higher cutting frequencies in the target taxon (Herrera *et al.* Chapter 2).

RAD-seq allows the formulation of robust species delineations

Our study, the first statistical rigorous test of species hypothesis in octocorals, provides conclusive evidence rejecting the current morphological species delimitation model for the genus *Paragorgia*. We find decisive support for a nested model that combines species boundaries from morphological taxonomy with cryptic diversity linked to environmental variables of geographic location and depth (Figs. 3 and 4). This nested model, proposes 9 species among the examined specimens. Five of these species correspond to the morphological species *P. coralloides*, *P. kaupeka*, *P. alisonae*, *P. johnsoni*, and *P. maunga*. Two splits, corresponding to sub-clades in the morphological species *P. arborea* and in *P. stephencairnsi*, indicate cases of cryptic species.

Herrera *et al.* (2012) found significant genetic differentiation of the north Pacific populations of *P. arborea* relative to the south Pacific, Atlantic and Indian ocean populations, and suggested that these populations may represent sub-species. The north Pacific populations of *P. arborea* were previously defined as a separate species, *P. pacifica*, by Verrill (1922) based on gross colony morphology, but later combined into a single species by Grasshoff (1979). Sánchez (2005) suggested potential small differences in medullar sclerite sizes and ornamentation between north Pacific specimens and specimens form elsewhere. However, we were unable to recognize these morphological differences in the few examined specimens in this study. Nonetheless, based on the decisive support for the split of *P. arborea* from analysis of genome-wide SNP makers indicates, we resurrect the species *Paragorgia pacifica* for the north Pacific populations of formerly *P. arborea*. We find no evidence of cryptic speciation between the north Atlantic and south Pacific *P. arborea* and therefore conclude it should be considered a single species as previously suggested by Herrera *et al.* (2012).

Depth is an important factor contributing to genetic differentiation and formation of species in the ocean, both shallow (Carlon & Budd 2002; Prada & Hellberg 2013) and deep (Miller *et al.* 2011; Jennings *et al.* 2013; Quattrini *et al.* 2013; Glazier & Etter 2014). The observed cryptic differentiation between specimens of *P. stephencairnsi* collected shallower than 350m and deeper than 1000m indicates that depth is also a diversifying force in octocorals from the AC clade, which had gone undetected due to the low variability of traditional sequence data (Herrera *et al.* 2012). The holotype of *P. stephencairnsi* was collected from approximately 350m in the Georgia Strait of British Columbia, overlapping in depth range and geographic region with that of most of the specimens from the shallow sub-clade examined in this

study. Therefore, we propose to conserve that name *P. stephencairnsi* for that shallow sub-clade, and consider the deep sub-clade as a new species.

Other recent species delimitation studies in anthozoan corals have also revealed significant incongruences when comparing morphological and single-locus species delimitation hypotheses (particularly from mitochondrial data) with phylogenetic evidence from multi-locus datasets (Pante *et al.* 2014; Prada *et al.* 2014). In line with the findings of Pante *et al.* (2014), we find that specimens of *Paragorgia* sharing identical *mtMutS* haplotypes can belong to more than one species. Contrastingly, Herrera *et al.* (2012) present strong evidence showing significant mitochondrial haplotype diversity in the south Pacific and north Atlantic populations of *Paragorgia arborea.* Our observations, together with those from the aforementioned studies, constitute compelling evidence indicating that there is no solid basis for the widespread assumption that *mtMutS* haplotypes may be equivalent to individual octocoral species, as proposed by Thoma *et al.* (2009). The analysis with RAD-seq, or alternative genomic multi-locus methods, of a larger number of specimens from diverse geographic locations and depth horizons will likely reveal further cryptic diversity not characterized by mitochondrial haplotypes (see Fig. S2, Fig. S3, and Table S8), and thus further illuminates taxonomy and systematics in this an other groups.

CONCLUSIONS

In this case study we demonstrate the empirical utility of RAD-seq to resolve phylogenetic relationships among divergent and recalcitrant taxa and to objectively infer species boundaries by testing alternative delimitation hypotheses. We were able to make use of RAD-seq to overcome long-standing technical difficulties in octocoral genetics, and to resolve fundamental questions in species definitions and systematics. We show that classic morphological taxonomy can greatly benefit from integrative approaches that provide objective tests to species delimitation hypothesis. Our results pave the way for addressing further questions in biogeography, species ranges, community ecology, population dynamics and evolution of octocorals and other marine taxa. The results from this study also represent a valuable reference resource for the development of tools, such as SNP arrays, that can be used to perform accurate species identifications, and generate species inventories that will aid the design and implementation of conservation and management policies.
METHODS

To perform identifications using current morphological species descriptions we performed colony observations and scanning electron microscopy of sclerites on 44 octocoral specimens from the AC clade (Table S1).

To obtain a genome-wide set of markers that could be useful for phylogenetic inferences of deepdivergent taxa and species delimitation in the AC clade (greater than 100 million years) we performed RAD sequencing with a the 6-cutter restriction enzyme PstI, which is predicted to cut between 32,000 and 110,000 times in the genome of an octocoral (Table S7). This predicted range was obtained using the observed frequency of the PstI recognition sequence, and its probability calculated using a trinucleotide composition model, in the genomes of the cnidarians *Nematostella vectensis, Acropora digitifera, Hydra vulgaris,* and *Alatina moseri* (Herrera *et al.* Chapter 2). Genome size range of 0.3-0.5 pg was used based on observations obtained through flow cytometry in gorgoniid octocorals by Luisa Dueñas at the Universidad de los Andes, Bogotá, Colombia (personal communication). Total genomic DNA was purified from specimens following protocols described in Herrera *et al.* (Chapter 3). Concentrationnormalized genomic DNA was submitted to Floragenex Inc (Eugene, OR). for library preparation and RAD sequencing. Libraries were sequenced by 48-multiplex, using 10-base pair barcodes, on a single lane of an Illumina Hi-Seq 2000 sequencer.

To compare the inferences obtained from RAD-seq data with the inferences drawn from traditional genetic barcoding data, we performed targeted sequencing of the mitochondrial *mtMutS* gene — a genetic marker widely used for phylogenetics and species delimitation studies in octocorals. Polymerase chain reactions were carried out following the protocols by Herrera *et al.* (Chapter 3). Primer pairs used for amplifications were AnthoCorMSH (Herrera *et al.* 2010) and Mut-3458R (Sánchez *et al.* 2003). Negative controls were included in every experiment to test for contamination. Purified PCR products were submitted to Eurofins Genomics (Eurofins MWG Operon, Inc.) for sequencing.

RAD-seq data filtering

Sequence reads were de-multiplexed and quality filtered with the process_radtags program from the package Stacks v1.20 (Catchen *et al.* 2013). Barcodes and Illumina adapters were excluded from each read and length was truncated to 91bp (-t 91) Reads with ambiguous bases were discarded (-c). Reads with an average quality score below 10 (-s 10) within a sliding window of 15% of the read length (-w 0.15) were discarded (-r). The rescue barcodes and RAD-tags algorithm was enabled (-r). Additional

filtering, and the clustering within and between individuals to identify loci was performed using the program pyRAD v2.01 (Eaton 2014). Reads with more than 33 bases with a quality score below 20 were discarded.

RAD-seq loci clustering and phylogenetic inference

We investigated different combinations of clustering thresholds (c 0.80, 0.85 and 0.90) and minimum number of taxa per locus (m 4, 6, and 9) in a reduced dataset that included one individual from each of the 12 putative morphological species. The minimum depth of coverage required to build a cluster and the maximum number of shared polymorphic sites in a locus were kept constant at 4 (d) and 3 (p) respectively. Loci sequences were concatenated into combined matrices. We refer to these 9 resulting matrices as the **'backbone'** matrices. Each of the resulting backbone matrices was analyzed in RAxML-HPC2 v8.0 (Stamatakis 2006) for maximum likelihood (ML) phylogenetic tree inference. For this, and all the other phylogenetic analyses in RAxML, we assumed a generalized time-reversible DNA substitution model with a gamma-distributed rate variation across sites (GTR GAMMA). Branch support was assessed by 500 bootstrap replicates.

We selected an optimal combination of loci clustering parameters as the set of parameters that minimized the number of missing data and maximized the number of phylogenetically informative sites while producing a highly supported phylogenetic tree. The optimal set of parameters chosen was a clustering threshold of 80% similarity among sequences (c 0.80) and a minimum coverage of taxa per locus of 75% (m 9). A concatenated matrix containing the sequence data of all the 44 octocoral specimens, denominated **'PHYLO'**, was built using this parameter combination (c 0.80, m 33) in pyRAD and subsequently analyzed in RAxML.

Phylogenetic inference with traditional genetic barcoding data

To compare the tree topology obtained from the phylogenetic inferences of the **PHYLO** RAD-seq dataset with traditional genetic barcoding data we analyzed the '**mitochondrial**' dataset (containing the *mtMutS* sequences) using RAxML. These two datasets – **PHYLO** and **mitochondrial** – contain data from the same individuals. To place the specimens from this study in a broader phylogenetic context we also analyzed the **mitochondrial** dataset in RAxML with the addition of *mtMutS* data from 233 additional specimens belonging to the AC clade, as well as outgroups (see Table S8, Fig. S2, and Fig. S3).

Testing species delimitation models for Paragorgia

We constructed 5 taxonomy-guided species delimitation models for *Paragorgia*: i) 'morphid' model: 7 species based on current morphological species descriptions (Sánchez 2005); ii) 'PAB' model: 8 species based on the 7 morphological species plus a split of *P. arborea* based on previous evidence of genetic differentiation of north Pacific populations (Herrera *et al.* 2012); iii) 'STE' model: 8 species based on the 7 morphological species plus a split of *P. stephencairnsi* based on depth differences (specimens collected <350m vs. >1000m), as depth is known to be an important structuring variable in marine taxa (Jennings *et al.* 2013; Prada & Hellberg 2013; Quattrini *et al.* 2013); iv) 'PABSTE' model: 9 species based on the 7 morphological species plus the splits of the PAB and STE models; v) 'splitPAB' model: 10 species based on the 7 morphological species plus the split of the STE model and an additional split in the PAB model where *P. arborea* is split in 3 species corresponding to the ocean basin where the specimens were collected (north Pacific, south Pacific and north Atlantic).

We also generated taxonomy-unguided species delimitation models for *Paragorgia* through Bayesian and ML implementations of the Poisson tree processes model (PTP) (available at http://species.h-its.org/ptp/). PTP estimates the number of speciation events in a rooted phylogenetic tree in terms of nucleotide substitutions (Zhang et al. 2013). We used PTP to analyze the trees obtained from phylogenetic inferences in RAXML of reduced *mtMutS* and RAD-seq datasets that include only members of the family Paragorgiidae (genera Paragorgia and Sibogagorgia). The 'PARAGORGIIDAE' RAD-seq concatenated matrix was generated in pyRAD using a clustering threshold of 80% similarity among sequences (c 0.80) and a minimum coverage of taxa per locus of 100% (m 33). The resulting phylogenetic trees of Paragorgia were rooted with the specimens of Sibogagorgia and analyzed by the PTP method using a Markov Chain Monte Carlo (MCMC) chain length of 500,000 generations (100 thinning, 25% burnin). We assessed convergence by examining the likelihood trace. The combinations of the ML or Bayesian PTP implementations (mIPTP and bPTP, respectively) with the *mtMutS* or RAD-seq trees of Paragorgia resulted in four species delimitation models: i) 'mIPTPmt' model; ii) 'bPTPmt' model; iii) 'mlPTPrad' model; and iv) 'bPTPrad' model. Lastly, because deep-sea corals are known to show genetic differentiation at ocean basin/regional scales (Miller et al. 2011; Morrison et al. 2011; Herrera et al. 2012), we constructed an additional taxonomy-unguided species delimitation model – the 'geo' model - based on the geographic location where the specimens were collected (north Pacific, south Pacific or north Atlantic ocean basins).

To estimate the marginal likelihood of each species delimitation model we generated a matrix including only specimens of *Paragorgia*, denominated '**PARAGORGIA'** using a clustering threshold of 80% similarity among sequences (c 0.80) and a minimum coverage of taxa per locus of 100% (m 31) in

pyRAD. In contrast to the **backbone**, **PHYLO**, and **PARAGORGIIDAE** RAD-seq matrices, this matrix contained the data of one SNP per locus and not the entire locus sequence. We analyzed these data using the implementation of BFD* in the SNAPP (Bryant *et al.* 2012) plug-in for the program BEAST v2.1.3 (Bouckaert *et al.* 2014). We performed a path-sampling of 48 steps, with a MCMC chain length of 100,000 (10,000 pre-burnin), following the guidelines from Leache *et al.* (2014). Bayesian factors were calculated from the marginal likelihood estimates for each model and compared using the framework proposed by Kass and Raftery (1995)

Species tree inference

To test the tree topology in the genus *Paragorgia* obtained by the phylogenetic analysis of the **PHYLO** and **PARAGORGIIDAE** concatenated matrices we performed a species tree inference from the SNP data in the **PARAGORGIA** matrix using the program SNAPP. This program allows the inference of species trees from unlinked SNP data (only one SNP per locus retained) bypassing the inference of individual gene trees (Bryant *et al.* 2012). We performed 3 independent runs using a MCMC chain length of 10,000,000 (sampling every 1,000 generations; pre-burnin of 1,000) with default prior distributions for coalescence rate, mutation rate and ancestral population size parameters. We assessed convergence to stationary distributions and effective sample sizes >200 after 10% burnin in the program TRACER (Rambaut & Drummond 2007). Species trees in the posterior distribution were summarized with the program DENSITREE v2.01 (Bouckaert 2010).

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SUPPLEMENTARY MATERIALS

Minimum taxa coverage per locus



Figure S1. Phylogenetic trees based on backbone matrices. Trees inferred from the 9 backbone RAD-seq matrices built with different parameters of clustering threshold (c 0.80, 0.85 and 0.90; indicated by vertical labels) and minimum number of taxa per locus (m 4, 6, and 9; indicated by horizontal labels). Each family is indicated with a different branch color: red for Paragorgiidae; blue for Coralliidae; and yellow for Alcyoniidae. Trees were inferred with RAxML. All interior branches have bootstrap support values of 100, except for those shown. Scale bars indicate substitutions per site.



	Paragorgia_sp_USNM_1075752	
	Paragorgia_regalis_USNM_	1014743_GQ293307_Sanchez2005
	Paragorgia_sp_USNM_107	1233
	Paragorgia_sp_USNM_107	1440
	Paragorgia_regalis_JQ241	244
	Paragorgia_sp_YPM_3691	00
	Paragorgia sp YPM 2729	4
	Paragorgia_sp_YPM_3538	31
	Paragorgia_sp_YPM_3863	6
	Paragorgia_arborea_MINH Paragorgia_alisonae_Valo	0046
	Paragorgia_sp_KC788261	0010
	Paragorgia_johnsoni_USNN	1_100898_Sanchez2005
	Paragorgia_sp_YPM_28905	66007 EXAMINED
	Paragorgia where NIWA 3	436 TYPE
Constanting and	Paragorgia_alisonae_NIWA	_3317_TYPE
	Paragorgia_alisonae_NIWA	_3312_JX128349_TYPE
	Paragorgia_alisonae_NIWA	_3313_11PE _42002
	Paragorgia sp YPM 3678	_+2002
	Paragorgia_alisonae_NIWA	_44606
	Paragorgia_alisonae_NIWA	_3316_TYPE
	Paragorgia_alisonae_NWA	73767 JX128348 Sanchez2005
	Paragorgia_johnsoni_KC7	88262
	Paragorgia_cf_johnsoni_KC	3984606
	Paragorgia_jonnsoni_10-w	HOI_20136_KC984607_EXAMINED
	Paragorgia_maunga_N	WA_3323_TYPE
	Paragorgia_maunga_NIW/	4_3322_TYPE
	Paragorgia_sp_USNM_56	615 A 3325 GO203305 TYPE
	Paragorgia maunga NIW	A 64980 EXAMINED
	Paragorgia_sp_USNM_54	830
	Paragorgia_sp_NIWA_446	07 3326 GO203314 TYPE
	Paragorgia sp USNM 566	90
	Paragorgia_regalis_USN	M_1122239
	Paragorgia_regalis_USN	M_1027063
	Paragorgia sp USNM 11	22235
	Paragorgia_sp_USNM_11	22227
	Paragorgia_sp_USNM_10	11094
	Paragorgia_yutiinux_USNI Paragorgia so USNM 11	22302
	Paragorgia_sp_USNM_11	22303
	Paragorgia_sp_USNM_10	82644
	Paragorgia_sp_USNM_10	15764 SUSNM 98792
	Paragorgia_dendroides_US	SNM_1072362
	Paragorgia_cf_regalis_USN	IM_1072337
	Paragorgia_sp_USNM_112 Paragorgia_sp_USNM_112	2238
	Paragorgia_regalis_USNM	1072338
	Paragorgia_cf_dendroides	USNM_98788
	Paragorgia_regalis_USIN	1_90709 JM 1075392
	Paragorgia_sp_USI	VM_1011093
	Paragorgia_sp_Sto	ne-Sanchez_1200601
	Paragorgia_sp_USI	MM_1122225 MM_1122305_IX128347
	Paragorgia_sp_Sto	ne-Sanchez_1310601
	Paragorgia_sp_USI	VM_1027079
	Paragorgia_sp_USI	VM_1122234 VM_1011095
	Paragorgia_sp_USI	VM_1122231
	Paragorgia_stephencairr	si_WHOI_C101_EXAMINED
	Paragorgia_stephencairr	SI_WHOI_C100_EXAMINED
	Paragorgia stephencairr	isi_USNM_101010_EXAMINED
	Paragorgia_stephencairr	si_WHOI_C104_EXAMINED
	Paragorgia_stephencairr	SI_WHOI_C102_EXAMINED
	Paragorgia_stephencairr	isi_RBCM_2344_EXAMINED
	Paragorgia_stephencairr	si_WHOI_C02_EXAMINED
	Paragorgia_stephencairr	SI_DFO_FOC25_EXAMINED
	Paragorgia_stephencairr	ISI_USINIVI_1092703
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		Paragorgia_stephencaims_OSINM_1075741
		Paragorgia_stephencaimsi_DFO_FOC5_EXAMINED
		Paragorgia stephencairnsi USNM 1124300 EXAMINED
		Paragorgia_stephencaimsi_USNM_1123930
		Paragorgia_stephencairnsi_CAS_190438
		Paragorgia stephencairnsi USNM 1007316 EXAMINED
		Paragorgia stephencairnsi WHOL CO3 EXAMINED
		Paragorgia stephencairnsi WHOL CO4 EXAMINED
		Paragorgia_stephencairnei_Whon_oo4_EXAMINED
		Paragorgia_stephencairnsi_USNM_1124298
		Paragorgia stephencairnsi USNM 1122304
	• Pa	aragorgia arborea USNM 1123936 JX124532
6	P	aragorgia_arborea_WPMNH_IX124557
	P	aragorgia_arborea_USNM_1123935_JX124580
l	P	aragorgia arborea USNM 1123934 JX124555
l	P	aragorgia_arborea_USNM_1123937_JX124604
l	P	aragorgia_arborea_USNM_1011097_JX124545
	P	aragorgia_arborea_USNM_100817_JX124539
	P	aragorgia_arborea_USNM_100846_JX124538
l	P	aragorgia arborea USNM 1011360 JX124582
l	P	aragorgia arborea USNM 100758 JX124584
	P	aragorgia arborea USNM 100818 JX124597
	P	aragorgia arborea USNM 50890 JX124521
1	P	aragorgia arborea USNM 100843 JX124529
	P	aragorgia arborea BStone ZC0706BOV01 JX124598
	P	aragorgia_arborea_USNM_1123938_JX124536
	P	aragorgia arborea NIWA 46318 JX124583
	P	aragorgia arborea NIWA 46319 JX124595
	P	aragorgia arborea NIWA 42001 JX124574
l	P	aragorgia arborea NIWA 44609 JX124534
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	P	aragorgia arborea NIWA 46377 JX124549
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ł	Pa	aragorgia_arborea_USNM_1016320_JX124581
l	Pa	aragorgia_arborea_USNM_1075753_JX124527
Į	Pa	aragorgia_arborea_BStone_20070178B01_JX124587
l	Pa	aragorgia_arborea_USNM_10/5/61_JX1245/6
	Pa	aragorgia_arborea_USNM_1122233_JX124586
ł	Pa	aragorgia_arborea_USNM_110/5/54_JX124593
	Pa	aragorgia_arborea_USNM_1122237_JX124590
l		aragorgia_arborea_USNM_1122240_UX124540
l		aragorgia_arborea_05NNV_1122240_37124522
l		aragorgia_arborea_AAnulews_DAVII_JA1245/7
l		aragorgia_arborea_USNM_1075744_124565
ľ		aragorgia_arborea_USNM_1007340_UX124565 EXAMINED
l	P	aragorgia_arborea_0511M_1007540_07124050_EXAMINED
ļ	P	aragorgia_arborea_HSOM_0072_EXAMINED
	P	aragorgia_arborea_USNM_1027060_UX124520
	Par	ragorgia_arborea_USNM_4091_UX124579
	Par	ragorgia_arborea_MNHN_422_IX124564
	Par	agorgia arborea NIWA 41780 JX124558 EXAMINED
	Par	agorgia arborea USNM 1092764 JX124589
l	Par	agorgia arborea NIWA 28123 JX124546
ł	Par	agorgia arborea NIWA 46320 JX124570
	Par	agorgia arborea USNM 80937 GQ293312
	Par	agorgia arborea NIWA 41854 JX124611
	Par	agorgia_arborea_NIWA_41999_JX124588
	Par	agorgia_arborea_USNM_1120444_JX124547
	Par	agorgia_arborea_USNM_4242_JX124567
	Par	agorgia_arborea_USNM_1092766_JX124554
	Par	ragorgia_arborea_NIWA_28422_JX124601
	Par	agorgia_arborea_NIWA_3309_JX124535_Sanchez2005
	Par	agorgia_arborea_NIWA_3308_JX124552_Sanchez2005
	Par	ragorgia_arborea_NIWA_46316_JX124573
	Par	agorgia_arborea_NIWA_3310_GQ293311_Sanchez2005
	Par	agorgia_arborea_USNM_80936_JX124543
	Par	ragorgia_arborea_NIWA_28156_JX124610
	Par	ragorgia_arborea_NIWA_28154_JX124553
10		LINE CONTRACTOR CONTRACT



0.02

Figure S2. Mitochondrial *mtMutS* gene tree of all available sequences for the clade AC. Tree inferred from *mtMutS* sequence data from specimens examined in this study, GenBank, and additional specimens. Each family is indicated with a different branch color: red for Paragorgiidae; blue for Coralliidae; and yellow for Alcyoniidae. Outgroups are indicated with black branches. Specimens examined in detail in this study are indicated with green labels. Type specimens are labeled TYPE. Specimens examined in Sanchez (2005) are labeled "Sanchez2005". Tree was inferred with RAxML. Scale bars indicate substitutions per site.





	Paragorgia stephenceirpei LISNM 1124300 EXAMINED
	Paragorgia_stephencaimsi_USNM_1124500_LXAMINED
	Paragorgia stephencairnsi WHOI CO4 EXAMINED
	Paragorgia stephencairnsi DFO FOC5 EXAMINED
	Paragorgia_stephencairnsi_USNM_1075741
	Paragorgia_stephencairnsi_USNM_1007316_EXAMINED
	Paragorgia_stephencairnsi_DFO_FOC26_EXAMINED
	Paragorgia_stephencairnsi_WHOI_C02_EXAMINED
	66 Paragorgia_stephencairnsi_DFO_FOC25_EXAMINED
	Paragorgia_stephenoairnsi_CAS_190438_
	Paragorgia_stephencaimsi_WHOT_COS_EXAMINED
	Paragorgia_stephencairnsi_Stone-Sanchez_200601106B01
	Paragorgia stephencairnsi RBCM 2344 EXAMINED
	Paragorgia_stephencairnsi_USNM_1123930
	Paragorgia_stephencairnsi_USNM_1124301
	Paragorgia_stephencairnsi_DFO_FOC30_EXAMINED
	Paragorgia_stephencairnsi_USNM_1124298
	Paragorgia_stephencairnsi_USNM_1122304
	Paragorgia_arborea_MC2_51244_JX124600
	Paragorgia arborea WHOL 03CAN EXAMINED
	Paragorgia arborea NIWA 3311 JX124525 Sanchez2005
	Paragorgia arborea NIWA 28156 JX124610
	Paragorgia_arborea_USNM_33559_JX124559
	Paragorgia_arborea_NIWA_28154_JX124553
	Paragorgia_arborea_USNM_4089_JX124591
	Paragorgia_arborea_USNM_80936_JX124543
	Paragorgia_arborea_NIWA_41780_JX124558_EXAMINED
	Paragorgia_arborea_USINM_4242_JX124567
	Paragorgia arborea NIWA 44608 IX124507
	Paragorgia arborea USNM 1092766 JX124554
	Paragorgia arborea MNHN 412 JX124609
	Paragorgia_arborea_NIWA_46317_JX124603
	Paragorgia_arborea_USNM_33560_JX124572
	Paragorgia_arborea_USNM_1092765_JX124520
	Paragorgia_arborea_NIWA_3308_JX124552_Sanchez2005
	Paragorgia_arborea_NIWA_28392_JX124575
	Paragorgia arborea NIWA 3310 GO293311 Sanchez2005
	Paragorgia arborea USNM 1010787 JX124528
	Paragorgia_arborea_NIWA_46316_JX124573
	Paragorgia_arborea_MCZ_15721_JX124606
	Paragorgia_arborea_NIWA_28123_JX124546
100	Paragorgia_arborea_USINM_4091_JX124579
and the second	Paragorgia arborea LISNM 33561 .IX124562
	Paragorgia arborea NIWA 28157 JX124531
	Paragorgia_arborea_NIWA_46320_JX124570
	Paragorgia_arborea_NIWA_3309_JX124535_Sanchez2005
	Paragorgia_arborea_NIWA_17970_JX124605
	Paragorgia_arborea_NIWA_44156_JX124563
	Paragorgia_arborea_USNM_4238_JX124608
	Paragorgia_arborea_USINM_4509_JX124596
	Paragorgia arborea NIWA 17971 JX124592
	Paragorgia arborea NIWA 41829 JX124585
	Paragorgia_arborea_NIWA_28161_JX124568_Sanchez2005
	Paragorgia_arborea_NIWA_46314_JX124594
	Paragorgia_arborea_USNM_1120444_JX124547
	Paragorgia_arborea_NIWA_25527_JX124548
	Paragorgia_arborea_MCZ_28057_JX124551
	Paragorgia arborea LISNM 1092764 IX124007
	Paragorgia arborea USNM 80937 GQ293312
	Paragorgia arborea NIWA 41854 JX124611
	Paragorgia_arborea_NIWA_28158_JX124566
	Paragorgia_arborea_USNM_1123936_JX124532
	78 Paragorgia_arborea_USNM_1123937_JX124604
	Paragorgia_arborea_USNM_1123934_JX124555
	Paragorgia_arborea_USNM_1123935_UX124590
	Paragorgia arborea NIWA 44609 JX124534
	Paragorgia_arborea_NIWA_76238_JX124533
	Paragorgia_arborea_NIWA_46318_JX124583
	80 Paragorgia_arborea_NIWA_28425_JX124550
	Paragorgia arborea NIWA 46315 IX124569



Figure S3. Mitochondrial *mtMutS* bootstrap support consensus tree of all available sequences for the clade AC. Tree inferred from *mtMutS* sequence data from specimens examined in this study, GenBank, and additional specimens. Each family is indicated with a different branch color: red for Paragorgiidae; blue for Coralliidae; and yellow for Alcyoniidae. Outgroups are indicated with black branches. Specimens examined in detail in this study are indicated with green labels. Type specimens are labeled "TYPE". Specimens examined in Sanchez (2005) are labeled "Sanchez2005". Tree was created with RAxML using a 50% majority consensus from 500 bootstrap replicates.

Table S1. Collection and sequence information for the specimens used in this study.

ID	Mamphala starl Granter	Delimitation		Date	T 11.	Depth		_
ID	Morphological Species	model PABSIE	Collection	Collected	Locality	(m)	Lat.	Lon.
p4	Anthomastus sp		WHOI	2011	Patagonian shelf			
86121	Corallium sp		NIWA	2012	Kermadec Ridge; Colville ridge volcano; TAN1213/18	380	-30.19	179.72
64642	Hemicorallium imperiale-laauense		NIWA	2010	Kermadec Ridge; Rumble II East seamount cone; TAN1007/97	1050	-35.42	178.65
83361	Hemicorallium imperiale-laauense		NIWA	2012	Bay of Plenty; Site SM1b, Matatara Knoll; TAN1206/168	948	-37.19	176.98
82674	Hemicorallium imperiale-laauense		NIWA	2012	Kermadec Ridge; Site SM2aa, summit of Whakatane Seamount; TAN1206/77	878	-36.81	177.47
86234	Hemicorallium imperiale-laauense		NIWA	2012	Kermadec Ridge; Northeast pimple volcano; TAN1213/22	483	-30.08	179.82
86286	Hemicorallium imperiale-laauense		NIWA	2012	Kermadec Ridge; Havre volcano; TAN1213/30	860	-31.13	-179.05
86232	Hemicorallium imperiale-laauense		NIWA	2012	Kermadec Ridge; Northeast pimple volcano; TAN1213/22	483	-30.08	179.82
69654	Heteropolypus sp		NIWA	2011	Chatam Rise; TRIP3306/78	495	-44.33	-177.22
85040	Heteropolypus sp		NIWA	2012	Chatam Rise; TAN1208/22	2098	-42.59	179.42
85662	Heteropolypus sp		NIWA	2012	Chatam Rise; TAN1208/61	1931	-42.59	179.59
66097	Paragorgia alisonae	P. alisonae	NIWA	2008	Cambell Plateau; TRIP2718/50	875	-50.02	175.00
41780	Paragorgia arborea	P. arborea	NIWA	2008	Chatam Rise; TRIP2617/120	600	-44.52	175.78
TC16_03	Paragorgia arborea	P. arborea	WHOI	2013	NE US Canyons; Munson Canyon; TowCam 16	540	40.54	-67.01
0672	Paragorgia arborea	P. pacifica	RBCM	2004	British Columbia; VE14280		53.33	-135.66
1007340	Paragorgia arborea	P. pacifica	USNM	2001	British Columbia; Vancouver Island	1168	48.44	-126.38
1601	Paragorgia arborea	P. pacifica	RBCM	2009	British Columbia; VE14444	695	53.31	-135.58
063902	Paragorgia coralloides	P. coralloides	WHOI	2005	New England Seamounts; Rehobot Seamount; H13	1821	37.46	-59.95
12016	Paragorgia coralloides	P. coralloides	WHOI	2003	New England Seamounts; Manning Seamount, station 4; AD3890	2000	38.23	-60.46
20136	Paragorgia johnsoni	P. johnsoni	TU/WHOI	2009	Gulf of Mexico; MC751; J2-464	438	28.19	-89.80
72152	Paragorgia kaupeka	P. kaupeka	NIWA	2011	Kermadec Ridge; Clark Seamount, chimney field, north cone; TAN1104/13	877	-36.45	177.84
82260	Paragorgia kaupeka	P. kaupeka	NIWA	2012	Kermadec Ridge; Site SM3a, summit of Clark Seamount; TAN1206/34	850	-36.45	177.84
82342	Paragorgia kaupeka	P. kaupeka	NIWA	2012	Kermadec Ridge; Site SM3a, Clark Seamount; TAN1206/40	1100	-36.45	177.84
84804	Paragorgia kaupeka	P. kaupeka	NIWA	2005	Kermadec Ridge; Clark Seamount; KOK0506/12	870	-36.45	177.84
64980	Paragorgia maunga	P. maunga	NIWA	2010	Kermadec Ridge; Silent II seamount; TAN1007/120	772	-35.17	178.89
1007316	Paragorgia stephencairnsi	P. sp. nov	USNM	2001	British Columbia; Vancouver Island	1168	48.44	-126.38
2344	Paragorgia stephencairnsi	P. sp. nov	RBCM	2004	British Columbia; VE13978	1194	53.37	-133.31
		-				•		

101010	Paragorgia stephencairnsi	P. stephencairnsi	USNM	2008	California; Piggy Bank, southern California; DW-026-02	283	33.92	-119.47
1124300	Paragorgia stephencairnsi	P. stephencairnsi	USNM	2006	British Columbia; Vancouver Island, Ohiat Island; OC 06/952	188	48.83	-125.13
Agam	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2012	British Columbia; Agamemnon Channel	32	49.72	-124.05
C02	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03
C03	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03
C04	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03
C05	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03
C100	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67
C101	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67
C102	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67
C104	Paragorgia stephencairnsi	P. stephencairnsi	WHOI	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67
FOC25	Paragorgia stephencairnsi	P. stephencairnsi	DFO	2012	British Columbia; W of Graham Island; 2012-65	204	53.31	-133.03
FOC26	Paragorgia stephencairnsi	P. stephencairnsi	DFO	2012	British Columbia; W of Graham Island; 2012-65	221	53.30	-133.04
FOC30	Paragorgia stephencairnsi	P. stephencairnsi	DFO	2012	British Columbia; W of Graham Island; 2012-65	318	53.48	-133.07
FOC5	Paragorgia stephencairnsi	P. stephencairnsi	DFO	2009	British Columbia; E of Graham Island; 2009-47	201	52.13	-128.90
1122230	Sibogagorgia cauliflora		USNM	2006	California; Davidson seamount; dive 945	2502	35.83	-122.61
2036	Sibogagorgia cauliflora		TU/WHOI	2009	Gulf of Mexico; DC583; J2-454	2440	28.39	-87.39

Table S2. RAD sequencing results and filtering statistics.

STACKS filtering

pyRAD filtering

Morphological Species	Delimitation model PABSTE	RAD-seq data file ID	Total sequenced reads	Reads discarded due to low quality	Retained reads	Reads that passed quality filtering	Trimmed reads due to detection of adapters	Total number of retained reads	% of retained reads after filtering steps
Anthomastus sp		PoC_p4_ARG	2,207,834	151,391	1,312,903	1,076,607	38,570	1,115,177	50.51
Corallium sp		COR_86121_NZ	3,826,317	305,019	3,013,324	2,555,644	22,650	2,578,294	67.38
H. imperiale-laaue	nse	COR_64642_NZ	4,344,702	239,625	4,105,077	3,697,256	21,823	3,719,079	85.60
H. imperiale-laaue	nse	COR_83361_NZ	3,351,944	291,963	2,588,492	2,103,039	37,016	2,140,055	63.85
H. imperiale-laaue	nse	HEM_82674_NZ	4,455,288	364,162	3,777,461	3,216,577	40,911	3,257,488	73.12
H. imperiale-laaue	nse	HEM_86234_NZ	2,408,325	186,372	2,142,057	1,876,187	16,195	1,892,382	78.58
H. imperiale-laaue	nse	HEM_86286_NZ	5,324,532	446,268	4,266,674	3,577,980	42,661	3,620,641	68.00
H. imperiale-laaue	nse	PAR_86232_NZ	6,092,276	503,438	5,206,376	4,454,042	35,518	4,489,560	73.69
Heteropolypus sp		ANT_69654_NZ	4,300,289	349,226	3,645,814	3,139,309	32,395	3,171,704	73.76
Heteropolypus sp		ANT_85040_NZ	4,498,336	393,748	3,882,304	3,333,172	26,687	3,359,859	74.69
Heteropolypus sp		ANT_85662_NZ	3,487,138	283,856	2,812,750	2,382,184	40,135	2,422,319	69.46
P. alisonae	P. alisonae	PAR_66097_NZ	3,007,362	220,209	2,511,717	2,219,265	21,207	2,240,472	74.50
P. arborea	P. arborea	PAR_41780_NZ	6,668,080	367,144	6,300,936	5,655,328	33,758	5,689,086	85.32
P. arborea	P. arborea	PAR_TC16_03_CAN	2,259,880	167,106	1,799,244	1,556,363	17,448	1,573,811	69.64
P. arborea	P. pacifica	PAR_0672_BC	4,348,226	293,526	4,054,700	3,543,642	29,284	3,572,926	82.17
P. arborea	P. pacifica	PAR_1007340_BCD	2,808,882	251,102	2,431,829	2,063,940	25,831	2,089,771	74.40
P. arborea	P. pacifica	PAR_1601_BC	4,202,185	221,386	3,980,799	3,606,334	25,434	3,631,768	86.43
P. coralloides	P. coralloides	PAR_063902_NES	3,244,860	185,122	3,059,738	2,746,228	20,408	2,766,636	85.26
P. coralloides	P. coralloides	PAR_12016_NES	3,838,375	226,253	3,612,122	3,255,440	15,945	3,271,385	85.23
P. johnsoni	P. johnsoni	PAR_20136_GOM	7,751,624	422,203	7,329,421	6,623,446	38,692	6,662,138	85.95
P. kaupeka	P. kaupeka	PAR_72152_NZ	3,909,139	268,871	2,926,960	2,531,796	26,984	2,558,780	65.46
P. kaupeka	P. kaupeka	PAR_82260_NZ	5,511,873	475,292	4,558,291	3,824,683	42,782	3,867,465	70.17
P. kaupeka	P. kaupeka	PAR_82342_NZ	4,056,319	307,425	3,339,653	2,866,786	33,020	2,899,806	71.49
P. kaupeka	P. kaupeka	PAR_84804_NZ	2,828,879	230,493	2,313,997	1,968,495	19,602	1,988,097	70.28
P. maunga	P. maunga	PAR_11369_NZ	2,536,311	252,554	2,008,028	1,619,915	34,477	1,654,392	65.23
P. stephencairnsi	P. sp. nov	PAR_1007316_BCD	2,257,183	151,729	1,606,180	1,371,398	36,692	1,408,090	62.38
P. stephencairnsi	P. sp. nov	PAR_2344_BC	2,858,311	159,073	2,699,238	2,414,787	26,272	2,441,059	85.40
P. stephencairnsi	P. stephencairnsi	PAR_101010_CA	6,642,673	398,317	6,244,356	5,595,342	39,240	5,634,582	84.82
P. stephencairnsi	P. stephencairnsi	PAR_1124300_WA	3,843,250	215,859	3,627,391	3,260,864	27,921	3,288,785	85.57
P. stephencairnsi	P. stephencairnsi	PAR_Agam_BC	4,322,564	260,095	4,062,469	3,619,988	26,669	3,646,657	84.36

P. stephencairnsi	P. stephencairnsi	PAR_C02_BCS	2,149,938	153,978	1,594,818	1,377,002	16,157	1,393,159	64.80
P. stephencairnsi	P. stephencairnsi	PAR_C03_BCS	2,815,327	245,051	2,264,951	1,916,452	19,800	1,936,252	68.78
P. stephencairnsi	P. stephencairnsi	PAR_C04_BCS	3,240,713	267,042	2,754,373	2,379,284	22,261	2,401,545	74.11
P. stephencairnsi	P. stephencairnsi	PAR_C05_BCS	3,465,396	294,547	2,972,548	2,538,577	25,005	2,563,582	73.98
P. stephencairnsi	P. stephencairnsi	PAR_C100_BCS	5,998,914	480,465	4,984,766	4,330,930	38,445	4,369,375	72.84
P. stephencairnsi	P. stephencairnsi	PAR_C101_BCS	5,332,619	446,452	4,652,535	4,030,919	30,684	4,061,603	76.17
P. stephencairnsi	P. stephencairnsi	PAR_C102_BCS	4,347,757	348,662	3,710,381	3,221,934	26,559	3,248,493	74.72
P. stephencairnsi	P. stephencairnsi	PAR_C104_BCS	2,720,994	210,479	2,203,568	1,893,649	19,213	1,912,862	70.30
P. stephencairnsi	P. stephencairnsi	PAR_FOC25_BCD	5,479,109	453,625	4,693,804	4,073,503	34,583	4,108,086	74.98
P. stephencairnsi	P. stephencairnsi	PAR_FOC26_BCD	5,242,949	477,361	4,528,629	3,778,655	37,137	3,815,792	72.78
P. stephencairnsi	P. stephencairnsi	PAR_FOC30_BCD	4,417,520	337,501	3,781,973	3,312,791	32,621	3,345,412	75.73
P. stephencairnsi	P. stephencairnsi	PAR_FOC5_BCD	2,834,377	239,051	2,455,041	2,109,251	23,444	2,132,695	75.24
S. cauliflora		SIB_1122230_DAV	1,701,751	123,755	1,220,077	1,016,203	26,104	1,042,307	61.25
S. cauliflora		SIB_2036_GOM	2,638,533	191,392	2,373,232	2,077,092	20,727	2,097,819	79.51
		AVERAGE	3,944,974	292,232	3,395,704	2,950,279	28,840	2,979,119	74.27
		STD DEV.	1,392,573	105,332	1,352,276	1,219,146	7,968	1,222,799	8.09

 Table S3. RAD-seq backbone clustering and matrix statistics.

Number of loci recovered in final data set for each taxon

Matrix	Min. # of taxa per locus (m)	Cluster thres- hold (c)	Total # loci in matrix	ANT_85040_NZ	COR_86121_NZ	HEM_82674_NZ	PAR_0672_BC	PAR_11369_NZ	PAR_12016_NES	PAR_20136_GOM	PAR_66097_NZ	PAR_82260_NZ	PAR_PR27_BCD	PoC_p4_ARG	SIB_2036_GOM	Total # of variable sites (var)	Total # of phylogenetically informative sites (pis)	pis/v ar	% of missing data
c80d5m4p3	4	80	62,726	10,588	24,303	25,369	41,190	44,891	38,333	46,858	42,472	41,810	45,377	7,154	26,243	601,763	109,290	0.18	49.6
c80d5m6p3	6	80	35,340	8,518	18,512	19,033	28,087	29,555	27,273	30,193	28,098	28,763	29,647	5,665	19,485	388,349	81,574	0.21	38.3
c80d5m9p3	9	80	10,333	5,176	8,607	8,701	9,558	9,778	9,528	9,879	9,477	9,635	9,794	3,458	8,183	132,803	32,017	0.24	20.7
c85d5m4p3	4	85	62,318	8,692	22,647	23,757	40,930	45,124	37,243	46,984	42,725	41,154	45,540	5,918	24,943	519,766	89,806	0.17	50.1
c85d5m6p3	6	85	33,785	7,020	17,146	17,608	27,047	28,504	26,012	29,192	27,197	27,640	28,608	4,715	18,225	328,464	66,419	0.20	38.3
c85d5m9p3	9	85	9,411	4,374	7,947	8,024	8,752	8,922	8,729	9,041	8,666	8,852	8,953	2,960	7,447	109,189	26,048	0.24	20.1
c90d5m4p3	4	90	58,765	5,796	18,775	19,255	37,994	43,101	33,306	44,722	40,899	37,911	43,075	4,073	21,371	402,245	63,938	0.16	51.4
c90d5m6p3	6	90	28,923	4,674	13,826	13,932	23,212	24,715	22,141	25,263	23,694	23,804	24,705	3,226	14,961	236,051	45,285	0.19	38.5
c90d5m9p3	9	90	7,312	2,943	6,277	6,318	6,855	6,955	6,836	7,048	6,779	6,905	6,984	2,029	5,815	72,149	16,695	0.23	19.4

Table S4. RAD-seq PHYLO clustering and matrix statistics.

Morphological	Delimitation	RAD-seq data	Total number	Mean	Std. dev. of	Number of clusters	Mean depth of clusters with	Std. dev. of cluster depth for clusters with	Number	Number of loci with	Number of loci with >5 Number depth of sites		Number of polym	Frequency • of	
Species	model PABSTE	file ID	of clusters	depth	cluster depth	depth greater than 5	depth greater than 5	with depth greater than 5	of loci	>5 depth coverage	and passed paralog filter	across loci	orphic sites	polymorph ic sites	
Anthomastus sp		PoC_p4_ARG	142,551	5.7	11.5	54,060	11.9	16.9	142,551	54,050	46,191	3,917,645	12,866	0.0032841	
Corallium sp		COR_86121_NZ	176,085	9.6	34.3	85,869	17.8	47.8	176,085	85,816	74,563	6,332,203	29,627	0.0046788	
H. imperiale-laauense	2	COR_64642_NZ	177,897	4.3	26.1	29,162	16.7	62.8	177,897	29,125	14,235	1,205,608	6,515	0.0054039	
H. imperiale-laauense	e	COR_83361_NZ	190,447	7.2	63.8	74,091	15.7	101.7	190,447	74,069	63,941	5,426,154	20,057	0.0036964	
H. imperiale-laauense	e	HEM_82674_NZ	219,724	8.9	53.4	92,866	18.8	81.1	219,724	92,810	80,468	6,834,388	29,969	0.0043850	
H. imperiale-laauense	e	HEM_86234_NZ	155,795	7.8	35.4	70,304	14.7	51.8	155,795	70,274	57,724	4,900,969	25,071	0.0051155	
H. imperiale-laauense	e	HEM_86286_NZ	252,395	9.0	76.6	96,846	20.8	122.8	252,395	96,772	81,349	6,912,750	30,928	0.0044741	
H. imperiale-laauense	e	PAR_86232_NZ	219,190	12.6	95.8	106,265	24.0	136.7	219,190	106,145	85,667	7,282,625	32,347	0.0044417	
Heteropolypus sp		ANT_69654_NZ	151,154	15.1	52.2	88,266	24.7	66.7	151,154	88,178	80,757	6,864,699	18,791	0.0027373	
Heteropolypus sp		ANT_85040_NZ	159,357	15.1	42.4	92,201	24.8	53.7	159,357	92,099	83,695	7,115,901	27,473	0.0038608	
Heteropolypus sp		ANT_85662_NZ	165,100	10.5	120.2	82,840	19.2	169.2	165,100	82,793	74,272	6,309,974	23,863	0.0037818	
P. alisonae	P. alisonae	PAR_66097_NZ	134,879	11.1	54.8	79,570	17.5	70.7	134,879	79,531	70,095	5,954,159	23,831	0.0040024	
P. arborea	P. arborea	PAR_41780_NZ	107,360	17.7	78.9	76,515	24.0	92.7	107,360	76,367	68,928	5,855,658	20,899	0.0035690	
P. arborea	P. arborea	PAR_TC16_03_CAN	138,933	7.7	46.4	62,427	14.9	68.6	138,933	62,399	54,495	4,625,417	13,801	0.0029837	
P. arborea	P. pacifica	PAR_0672_BC	116,545	21.4	71.8	85,334	28.5	82.8	116,545	85,165	77,189	6,559,823	25,068	0.0038214	
P. arborea	P. pacifica	PAR_1007340_BCD	149,606	9.9	30.4	76,311	17.6	41.0	149,606	76,259	67,539	5,736,203	22,134	0.0038587	
P. arborea	P. pacifica	PAR_1601_BC	118,243	21.7	65.9	85,263	29.3	76.2	118,243	85,074	77,035	6,547,638	24,879	0.0037997	
P. coralloides	P. coralloides	PAR_063902_NES	132,778	20.1	61.7	87,919	29.4	74.1	132,778	87,756	79,692	6,774,134	26,579	0.0039236	
P. coralloides	P. coralloides	PAR_12016_NES	129,675	36.8	207.3	90,736	51.8	246.3	129,675	90,474	82,903	7,049,554	21,990	0.0031193	
P. johnsoni	P. johnsoni	PAR_20136_GOM	122,432	29.8	487.2	89,665	40.0	568.9	122,432	89,630	81,301	6,911,399	23,857	0.0034518	
P. kaupeka	P. kaupeka	PAR_72152_NZ	132,520	12.7	197.2	77,976	20.2	256.8	132,520	77,929	69,780	5,926,828	16,005	0.0027004	
P. kaupeka	P. kaupeka	PAR_82260_NZ	186,768	13.1	77.9	89,892	25.4	110.9	186,768	89,766	80,556	6,845,659	17,710	0.0025870	
P. kaupeka	P. kaupeka	PAR_82342_NZ	148,202	12.6	43.0	83,517	21.0	55.8	148,202	83,429	74,747	6,349,820	16,935	0.0026670	
P. kaupeka	P. kaupeka	PAR_84804_NZ	124,756	10.5	119.9	69,174	17.3	160.6	124,756	69,119	61,022	5,182,009	15,300	0.0029525	
P. maunga	P. maunga	PAR_11369_NZ	113,838	19.3	161.8	84,649	25.2	187.3	113,838	84,510	76,038	6,461,976	26,079	0.0040358	
P. stephencairnsi	P. sp nov	PAR_1007316_BCD	166,722	5.9	17.7	60,649	12.9	27.9	166,722	60,636	51,368	4,357,427	19,584	0.0044944	

Table

P. stephencairnsi	P. sp nov	PAR_2344_BC	125,366	13.4	51.7	80,271	19.9	63.7	125,366	80,218	72,194	6,132,435	25,912	0.0042254
P. stephencairnsi	P. stephencairnsi	PAR_101010_CA	115,888	33.2	134.3	92,526	41.1	149.3	115,888	92,225	84,367	7,173,308	24,279	0.0033846
P. stephencairnsi	P. stephencairnsi	PAR_1124300_WA	109,466	20.9	103.4	83,749	26.8	117.6	109,466	83,645	76,157	6,471,962	21,532	0.0033270
P. stephencairnsi	P. stephencairnsi	PAR_Agam_BC	131,554	19.3	95.4	87,778	28.1	115.8	131,554	87,648	79,612	6,766,163	23,143	0.0034204
P. stephencairnsi	P. stephencairnsi	PAR_C02_BCS	111,985	9.0	22.7	63,618	14.2	29.0	111,985	63,597	56,380	4,786,614	16,361	0.0034181
P. stephencairnsi	P. stephencairnsi	PAR_C03_BCS	116,493	11.7	41.1	75,396	17.0	50.3	116,493	75,356	67,485	5,732,283	20,104	0.0035072
P. stephencairnsi	P. stephencairnsi	PAR_C04_BCS	128,452	13.0	47.9	81,845	19.4	59.1	128,452	81,792	73,789	6,269,830	21.010	0.0033510
P. stephencairnsi	P. stephencairnsi	PAR_C05_BCS	128,134	13.9	64.4	83,590	20.3	79.0	128,134	83,533	75,306	6,399,562	21,288	0.0033265
P. stephencairnsi	P. stephencairnsi	PAR_C100_BCS	152,749	19.9	55.9	92,858	31.6	69.2	152,749	92,693	84,309	7,168,326	22,907	0.0031956
P. stephencairnsi	P. stephencairnsi	PAR_C101_BCS	152,447	18.4	82.8	92,630	29.2	104.8	152,447	92,487	83,853	7,129,265	23,371	0.0032782
P. stephencairnsi	P. stephencairnsi	PAR_C102_BCS	131,680	17.3	46.6	88,133	24.9	55.3	131,680	88,038	79,983	6,798,340	21,816	0.0032090
P. stephencairnsi	P. stephencairnsi	PAR_C104_BCS	119,764	11.4	29.6	74,773	17.0	36.3	119,764	74,738	67,146	5,703,606	18,976	0.0033270
P. stephencairnsi	P. stephencairnsi	PAR_FOC25_BCD	154,768	18.3	56.3	93,419	29.2	70.4	154,768	93,284	84,570	7,190,050	24,801	0.0034494
P. stephencairnsi	P. stephencairnsi	PAR_FOC26_BCD	440,488	6.3	26.0	107,701	20.5	49.9	440,488	107,621	94,298	8,008,313	28,868	0.0036048
P. stephencairnsi	P. stephencairnsi	PAR_FOC30_BCD	314,181	7.6	44.3	95,360	21.1	78.8	314,181	95,276	84,774	7,202,768	24,900	0.0034570
P. stephencairnsi	P. stephencairnsi	PAR_FOC5_BCD	197,822	7.5	60.1	75,192	16.9	96.8	197,822	75,145	66,755	5,669,791	19,056	0.0033610
S. cauliflora		SIB_1122230_DAV	123,947	5.5	29.3	40,167	12.8	50.7	123,947	40,157	33,058	2,804,036	11,864	0.0042310
S. cauliflora		SIB_2036_GOM	127,391	11.3	36.4	71,374	18.6	47.3	127,391	71,315	63,094	5,358,510	17,312	0.0032307
		AVERAGE	157,171	13.9	76.4	80,744	22.6	99.0	157,171	80,658	71,425	6,068,313	21,810	0.0036393
		STANDARD DEV	59,920	7.2	77.0	15,153	8.0	89.6	59,920	15,119	14,833	1,262,657	5,318	0.0006234

Table S5. Nucleotide diversity and	error rate estimates per specimen	based on the PHYLO matrix
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Morphological Species	PABSTE	RAD-seq data file ID	Nucleotide diversity (Pi)	Error Rate
Anthomastus sp		PoC_p4_ARG	0.01295412	0.0029032
Corallium sp		COR_86121_NZ	0.01467291	0.00258843
H. imperiale-laauense*		COR_64642_NZ	0.04406216	0.01630744
H. imperiale-laauense		COR_83361_NZ	0.01399387	0.00269528
H. imperiale-laauense		HEM_82674_NZ	0.01440962	0.00268885
H. imperiale-laauense		HEM_86234_NZ	0.01676557	0.00254558
H. imperiale-laauense		HEM_86286_NZ	0.01526906	0.00234296
H. imperiale-laauense		PAR_86232_NZ	0.01633107	0.00217237
Heteropolypus sp		ANT_69654_NZ	0.00936378	0.00151989
Heteropolypus sp		ANT_85040_NZ	0.01122064	0.00146665
Heteropolypus sp		ANT 85662 NZ	0.01161827	0.00189506
P. alisonae	P. alisonae	PAR 66097 NZ	0.0127996	0.00205016
P. arborea	P. arborea	PAR 41780 NZ	0.01209999	0.00186767
P. arborea	P. arborea	PAR TC16 03 CAN	0.01293451	0.00241858
P. arborea	P. pacifica	PAR 0672 BC	0.01189447	0.00170062
P. arborea	P. pacifica	PAR 1007340 BCD	0.01285038	0.00223148
P. arborea	P. pacifica	PAR 1601 BC	0.01205019	0.00164477
P. coralloides	P. coralloides	PAR 063902 NES	0.01182997	0.00146894
P coralloides	P. coralloides	PAR 12016 NES	0.01035726	0.00116583
P iohnsoni	P. johnsoni	PAR 20136 GOM	0.01159218	0.00137491
P kauneka	P. kauneka	PAR 72152 NZ	0.01085919	0.00209157
P kaupeka	P. kaupeka	PAR 82260 NZ	0.01077562	0.00211998
P kaupeka	P kaupeka	PAR 82342 NZ	0.01082052	0.00225821
P kauneka	P kaupeka	PAR 84804 NZ	0.01169178	0.00221668
P maunga	P. maunga	PAR 11369 NZ	0.01225252	0.00176518
P stenhencairnsi	P sn nov	PAR 1007316 BCD	0.01488811	0.00295065
P stephencairnsi	P sp nov	PAR 2344 BC	0.01260347	0.00188425
P stephencairnsi	P. stenhencairnsi	PAR 101010 CA	0.01063818	0.00134692
P stephencairnsi	P. stephencairnsi	PAR 1124300 WA	0.01092481	0.00163497
P stephencairnsi	P. stephencairnsi	PAR Agam BC	0.01117934	0.00151902
P stephencairnsi	P. stephencairnsi	PAR CO2 BCS	0.01180863	0.00194975
P stephencairnsi	P. stephencairnsi	PAR C03 BCS	0.01129071	0.00175198
P stephencairnsi	P. stephencairnsi	PAR C04 BCS	0.01098867	0.00171446
P stephencairnsi	P. stephencairnsi	PAR C05 BCS	0.01117442	0.00167389
P stephencairnsi	P. stephencairnsi	PAR C100 BCS	0.01063313	0.00155401
P stephencairnsi	P. stephencairnsi	PAR C101 BCS	0.0109634	0.00160305
P stephencairnsi	P. stephencairnsi	PAR C102 BCS	0.01060609	0.00159815
P stephencairnsi	P stephencairnsi	PAR C104 BCS	0.01124495	0.00175065
P stephencairnsi	P stephencairnsi	PAR FOC25 BCD	0.01104859	0.00154584
P stephencairnsi	P. stephencairnsi	PAR FOC26 BCD	0.01263196	0.00180015
P stephencairnsi	P. stephencairnsi	PAR FOC30 BCD	0.01196811	0.00182162
P stenhencairnsi	P. stephencairnsi	PAR FOC5 BCD	0.0119129	0.00186229
S cauliflora		SIB 1122230 DAV	0.01525699	0.00328691
S. cauliflora		SIB 2036 GOM	0.01214983	0.00221517
		AVERAGE	0.01291599	0.00228532
		SD	0.00513429	0.00223806

*Excluded from calculations due to low number of loci

Table S6. Nucleotide diversity and error rate estimates per species based on the PHYLO matrix

		Nucle	eotide diversit	y (Pi)	Per-site sequence error rate (Epsilon)						
Morphological Species	Delimitation model PABSTE	mean	SD	SE	mean	SD	SE				
Anthomastus sp		0.01295412			0.00290320						
Corallium sp		0.01467291			0.00258843						
H. imperiale-laauense		0.01935775	0.01094005	0.00489254	0.00447727	0.00522005	0.00233448				
Heteropolypus sp		0.01073423	0.00120338	0.00069477	0.00162720	0.00023350	0.00013481				
P. alisonae	P. alisonae	0.01279960			0.00205016						
P. arborea	P. arborea	0.01245029	0.00056582	0.00040010	0.00193813	0.00041487	0.00029336				
P. arborea	P. pacifica	0.01073423	0.00120338	0.00069477	0.00162720	0.00023350	0.00013481				
P. coralloides	P. coralloides	0.01245029	0.00056582	0.00040010	0.00193813	0.00041487	0.00029336				
P. johnsoni	P. johnsoni	0.01159218			0.00137491						
P. kaupeka	P. kaupeka	0.01241323	0.00175196	0.00087598	0.00229768	0.00048923	0.00024462				
P. maunga	P. maunga	0.01225252			0.00176518						
P. stephencairnsi	P. sp. nov	0.01290646	0.00280248	0.00198165	0.00229281	0.00093033	0.00065784				
P. stephencairnsi	P. stephencairnsi	0.01137950	0.00066172	0.00016543	0.00169174	0.00016364	0.00004091				
S. cauliflora		0.01370341	0.00219709	0.00155358	0.00275104	0.00075783	0.00053587				

Table S7. Predictions of # of RAD-tags in octocorals using PstI. Data for *Nematostella vectensis* obtained from the U.S. Joint Genome Institute (JGI-DOE) database. Data for *Acropora digitifera, Hydra vulgaris*, and *Alantina moseri* obtained from the U.S. National Center for Biotechnology Information (NCBI) WGS database. Observed frequency of recognition sequences and calculated probability based on a trinucleotide genome composition model were generated following the methodology described by Herrera et al. (2014). Octocoral genome size ranges were obtained by Luisa Dueñas from gorgoniid octocorals through flow cytomery at the Universidad de los Andes, Bogota, Colombia. Abbreviation: restriction sites (RS).

Cnidarians with sequenced genomes

Species	Common name C-value	Genome size (Mbp) Unambiguous	Observed frequency of PstI RS per bp	Prob. of PstI recognition site per bp based on trinucleotide model
Nematostella vectensis	Starlet Anemone	297.39	0.00016661	0.00019608
Acropora digitifera	Staghorn Coral	364.97	0.00021313	0.00022777
Hydra vulgaris	Hydra	1189.96	0.00010830	0.00010871
Alatina moseri	Sea Wasp	1544.15	0.00020617	0.00021637

Predictions of # of PstI RS in known octocoral genome size range

Octocoral C-value	OctocoralPredicted # of PstI RSoral Genome sizebased on N. vectensisoe(Mbp)observed frequency		Predicted # of PstI RS based on <i>A. digitifera</i> observed frequency	Predicted # of PstI RS based on <i>H. vulgaris</i> observed frequency	Predicted # of PstI RS based on <i>A. moseri</i> observed frequency		
0.3	293.40	48,882.95	62,533.09	31,774.37	60,489.13		
0.5	489.00	81,471.59	104,221.81	52,957.29	100,815.22		
Octocoral C-value	Octocoral Genome size (Mbp)	Predicted # of PstI RS based on <i>N. vectensis</i> trinucleotide genome composition probability	Predicted # of PstI RS based on <i>A. digitifera</i> trinucleotide genome composition probability	Predicted # of PstI RS based on <i>H. vulgaris</i> trinucleotide genome composition probability	Predicted # of PstI RS based on <i>A. moseri</i> trinucleotide genome composition probability		
0.3	293.40	57,529.94	66,828.40	31,895.01	63,482.46		
0.5	489.00	95,883.23	111,380.67	53,158.36	105,804.09		

Predictions of # of PstI RAD-tags in known octocoral genome size range

Octocoral C-value	Octocoral Genome size (Mbp)	Predicted # of PstI RS based on <i>N. vectensis</i> observed frequency	Predicted # of PstI RS based on <i>A. digitifera</i> observed frequency	Predicted # of PstI RS based on <i>H. vulgaris</i> observed frequency	Predicted # of PstI RS based on <i>A. moseri</i> observed frequency
0.3	293.40	97,765.91	125,066.17	63,548.75	120,978.26
0.5	489.00	162,943.18	208,443.62	105,914.58	201,630.43
Octocoral C-value	Octocoral Genome size (Mbp)	Predicted # of PstI RS based on <i>N. vectensis</i> trinucleotide genome composition probability	Predicted # of PstI RS based on <i>A. digitifera</i> trinucleotide genome composition probability	Predicted # of PstI RS based on <i>H. vulgaris</i> trinucleotide genome composition probability	Predicted # of PstI RS based on <i>A. moseri</i> trinucleotide genome composition probability
0.3	293.40	115,059.87	133,656.80	63,790.03	126,964.91
0.5	489.00	191,766,45	222.761.34	106,316,71	211,608,19

		•	Collecti	Catalog		·	Depth			Taxonomic
Species	ID	Genbank	on	Number	Date	Locality	(m)	Lat.	Lon.	remarks
Anthomasttus ritteri		DQ302816	K. McFa	deen	1998	Off Pebble Beach: Californa: USA	300	36.58	-122.10	
Anthomastus cf grandiflorus		KC984603								
Anthomastus robustus delta		KC984604								
Anthomastus sp	p4		WHOI	p4	2011	Patagonian shelf				
Chelidonisis aurantiaca mexicana		KC788274								
Corallium elatius		AB700134								
Corallium elatius		AB700135								
Corallium japonicum		AB595189								
Corallium kishinouyei		GQ293300	USNM	1072441	2003	Off Laysan Island: Hawaii: USA	1490	25.70	-171.45	
Corallium kishinouyei		KC782353								
Corallium konojoi		AB595190								
Corallium regale		AF385321								
Corallium rubrum		AB700136								
Corallium secundum		GQ293303	USNM	1010758	2001	Off Maui: Hawaii: USA	240	20.88	-156.73	
Corallium secundum		KC782347								
Corallium sp	86121		NIWA	86121	2012	Kermadec Ridge; Colville ridge volcano; TAN1213/18	380	-30.19	179.72	
Corallium sp	Coralliumsp56807		USNM	56807	1978	USA, Hawaii, Oahu Island, Makapuu Point	366	21.30	-157.53	
Corallium sp		KC788270								
Corallium tortuosum		GQ293306	USNM	1089600	2003	New Caledonia	470-621	-23.71	168.26	
Hemicorallium ducale		DQ297416								
Hemicorallium ducale		EF060050								
Hemicorallium ducale		EU293805								
Hemicorallium imperiale		KC782352								
Hemicorallium imperiale		KC782355								
Hemicorallium imperiale-laauense	64642		NIWA	64642	2010	Kermadec Ridge; Rumble II East seamount cone; TAN1007/97	1050	-35.42	178.65	
Hemicorallium imperiale-laauense	82674		NIWA	82674	2012	Kermadec Ridge; Site SM2aa, summit of Whakatane Seamoun	878	-36.81	177.47	
Hemicorallium imperiale-laauense	83361		NIWA	83361	2012	Bay of Plenty; Site SM1b, Matatara Knoll; TAN1206/168	948	-37.19	176.98	
Hemicorallium imperiale-laauense	86232		NIWA	86232	2012	Kermadec Ridge; Northeast pimple volcano; TAN1213/22	483	-30.08	179.82	
Hemicorallium imperiale-laauense	86234		NIWA	86234	2012	Kermadec Ridge; Northeast pimple volcano; TAN1213/22	483	-30.08	179.82	
Hemicorallium imperiale-laauense	86286		NIWA	86286	2012	Kermadec Ridge; Havre volcano; TAN1213/30	860	-31.13	-179.05	
Hemicorallium imperiale-laauense	Claauense1072452		USNM	1072452	2003	USA, Hawaii, Laysan Island	1509	25.70	-171.44	
Hemicorallium imperiale-laauense		GQ293301	USNM	1071433	2004	Off Keahole Point: Hawaii Island: Hawaii: USA	867	19.80	-156.13	
Hemicorallium imperiale-laauense		GQ293302	USNM	1075800	2004	Pratt Seamount: Alaska: USA	1627	56.32	-142.44	
Hemicorallium laauense		KC782348								
Hemicorallium niobe		EF060051								
Hemicorallium niobe		KC788267								
Heteropolypus sp	69654		NIWA	69654	2011	Chatam Rise: TRIP3306/78	495	-44 33	-177 22	
Heteropolypus sp	85040		NIWA	85040	2012	Chatam Rise: TAN1208/22	2098	-42.59	179 42	
Heteropolypus sp	85662		NIWA	85662	2012	Chatam Rise; TAN1208/61	1931	-42.59	179.59	
Ideogorgia capensis		GQ342502								
Paragorgia alisonae	66097	•	NIWA	66097	2008	Cambell Plateau: TRIP2718/50	875	-50.02	175.00	
Paragorgia alisonae	Palisonae3312	JX128349	NIWA	3312	1998	New Zealand, Otara Hill, 1171/25, Z9596	980	-48.02	166.08	type

Table S8. Collection information for all specimens in the clade AC with available mtMutS sequences

Paragorgia alisonae	Palisonae3313		NIWA	3313	1998	New Zealand, Otara Hill, 1171/25, Z9596	980	-48.02	166.08	type
Paragorgia alisonae	Palisonae3315		NIWA	3315	1998	New Zealand, Otara Hill, 1171/24, Z9595	940	-48.02	166.10	type
Paragorgia alisonae	Palisonae3316		NIWA	3316	1997	New Zealand, TAN9713/037, Z8981	1041	-44.96	174.19	type
Paragorgia alisonae	Palisonae3317		NIWA	3317	1998	New Zealand, TRIP1171/12, Z9583	935	-48.03	166.10	type
Paragorgia alisonae	Palisonae42002		NIWA	42002	2008	New Zealand, TRIP2551/258	930.0	-44.7850	-176.583	3333333
Paragorgia alisonae	Palisonae44606		NIWA	44606	2007	New Zealand, TRIP2521/9	1068	-44.7433	-177.053	33333
Paragorgia alisonae	ParagospValpo046			Valpo046		Chile, Valparaiso, deep seamount				
Paragorgia aotearoa		GQ293305	NIWA	3325	1996	New Zealand, Mt. Muck (Box Hill Complex), TAN9609/40	700	-42.83	176.92	type
Paragorgia arborea	0672		RBCM	011-00067-002	2004	British Columbia; VE14280		53.33	-135.66	
Paragorgia arborea	1601		RBCM	011-00160-001	2009	British Columbia; VE14444	695	53.31	-135.58	
Paragorgia arborea	41780	JX124558	NIWA	41870	2008	New Zealand, TRIP2617/120	600	-44.52	175.78	
Paragorgia arborea	1007340	JX124556	USNM	1007340	2001	British Columbia; Vancouver Island	1168	48.44	-126.38	
Paragorgia arborea	Paragosp1014919	JX124578	USNM	1014919	2003	Davidson Seamount, California, USA	1313	35.70	-122.70	Sanchez 2005
Paragorgia arborea	Paragosp1027060	JX124560	USNM	1027060	2003	Pioneer Seamount, South of farallon Islands, California, USA	1712	37.40	-123.44	
Paragorgia arborea	Paragosp1075738	JX124526	USNM	1075738	2004	Dickins Seamount, Gulf of Alaska, USA	760	54.55	-136.84	
Paragorgia arborea	Paragosp1075744	JX124565	USNM	1075744	2004	Dickins Seamount, Gulf of Alaska, USA	851	54.51	-136.91	
Paragorgia arborea	Paragosp1075745	JX124602	USNM	1075745	2004	Dickins Seamount, Gulf of Alaska, USA	849	54.51	-136.91	
Paragorgia arborea	Paragosp1075746	JX124540	USNM	1075746	2004	Welker Seamount, Gulf of Alaska, USA	780	55.05	-140.31	
Paragorgia arborea	Paragosp1075753	JX124527	USNM	1075753	2004	Welker Seamount, Gulf of Alaska, USA	1112	55.07	-140.41	
Paragorgia arborea	Paragosp1075754	JX124593	USNM	1075754	2004	Welker Seamount. Gulf of Alaska, USA	1084	55.07	-140.41	
Paragorgia arborea	Paragosp1075760	JX124530	USNM	1075760	2004	Pratt Seamount, Gulf of Alaska, USA	959	56.17	-142.70	
Paragorgia arborea	Paragosp1075761	JX124576	USNM	1075761	2004	Pratt Seamount, Gulf of Alaska, USA	941	56.17	-142.70	
Paragorgia arborea	Paragosp1075766	JX124541	USNM	1075766	2004	Welker Seamount, Gulf of Alaska, USA	1114	55.07	-140.41	
Paragorgia arborea	Paragosp17971	JX124592	NIWA	17971	2004	New Zealand, 1172/06, Z9566	1235	-44.80	-177.12	
Paragorgia arborea	Paragosp200701	JX124587	B. Stone	20070178B01	2004	Gulf of Alaska, USA	867	55.91	-154.02	
Paragorgia arborea	Paragosp44156	JX124563	NIWA	44156	2007	New Zealand, TRIP2416/54	720-741	-47.47	177.02	
Paragorgia arborea	Paragosp46314	JX124594	NIWA	46314	2008	New Zealand, TRIP2324/48	843-998	-50.05	174.73	
Paragorgia arborea	Paragosp46315	JX124569	NIWA	46315	2008	New Zealand, TRIP2571/65	888-101	:-47.55	177.86	
Paragorgia arborea	Paragosp46316	JX124573	NIWA	46316	2007	New Zealand, TRIP2617/120	600	-44.52	175.77	
Paragorgia arborea	Paragosp46317	JX124603	NIWA	46317	2008	New Zealand, TRIP2494/13	931-102	: -4 7.58	177.78	
Paragorgia arborea	Paragosp46318	JX124583	NIWA	46318	2008	New Zealand, TRIP2551/254	794-987	-44.73	-177.04	
Paragorgia arborea	Paragosp46319	JX124595	NIWA	46319	2006	New Zealand, TRIP2614		-49.50	176.00	
Paragorgia arborea	Paragosp46377	JX124549	NIWA	46377	1927	New Zealand, TRIP2571/53	952-111	{ -50.00	176.06	
Paragorgia arborea	Paragosp56389		USNM	56389	1962	USA, Hawaii, Bushnell Seamount	1920	18.55	-155.44	
Paragorgia arborea	ParagospDAVI1	JX124577	A. Andre	r DAVI1	2002	Davidson Seamount, California, USA	1313	35.75	-122.70	
Paragorgia arborea	ParagospDAVI3	JX124542	A. Andre	r DAVI3	2004	Davidson Seamount, California, USA	1313	35.75	-122.70	
Paragorgia arborea	Paragospnizinski	JX124547	USNM	1120444	2008	off Maryland, USA	400	37.06	-74.62	
Paragorgia arborea	Parborea100758	JX124584	USNM	100758	1994	Aleutian Islands		52.00	-170.00	
Paragorgia arborea	Parborea100817	JX124539	USNM	100817	1 994	Atka Island, Andreanof Islands, Aleutian Islands		53.00	-174.00	
Paragorgia arborea	Parborea100818	JX124597	USNM	100818	1994	Semisopochnoi Island, Rat Islands, Aleutian Islands		52.17	179.72	
Paragorgia arborea	Parborea100843	JX124529	USNM	100843	1994	Tanaga Island, Andreanof Islands, Aleutian Islands		52.00	-178.00	
Paragorgia arborea	Parborea100846	JX124538	USNM	100846	1994	Yunaska Island, Islands of Four Mountains, Aleutian Islands		53.00	-171.00	
Paragorgia arborea	Parborea1010787	JX124528	USNM	1010787	2000	Norfolk Canyon, Virginia, USA	375-489	37.07	-74.66	
Paragorgia arborea	Parborea1011097	JX124545	USNM	1011097	2002	Buldir Reef, Rat Islands, Aleutian Islands	160	51.96	176.83	
Paragorgia arborea	Parborea1011360	JX124582	USNM	1011360	2001	off Umnak Island, Fox Islands, Aleutian Islands	102	53.68	-169.11	

Paragorgia arborea	Parborea1092764	JX124589	USNM	1092764	2000	East of Virginia Beach, Virginia, USA	375-489	37.07	-74.66	
Paragorgia arborea	Parborea1092765	JX124520	USNM	1092765	2000	East of Virginia Beach, Virginia, USA	375-489	37.07	-74.66	
Paragorgia arborea	Parborea1092766	JX124554	USNM	1092766	2000	East of Virginia Beach, Virginia, USA	375-489	37.07	-74.66	
Paragorgia arborea	Parborea17969	JX124607	NIWA	17969	2002	New Zealand, 1621/18, Z11010	900	-44.74	-177.19	
Paragorgia arborea	Parborea17970	JX124605	NIWA	17970	2001	New Zealand, AEX0101/80, Z10956	753	-44.74	-177.19	
Paragorgia arborea	Parborea25527	JX124548	NIWA	25527	2004	New Zealand, TAN0408/23	826	-42.83	177.42	
Paragorgia arborea	Parborea28123	JX124546	NIWA	28123	2000	New Zealand, 1390/12, Z11161	872	-47.31	165.83	
Paragorgia arborea	Parborea28154	JX124553	NIWA	28154	1981	New Zealand, T16	427	-43.35	178.66	
Paragorgia arborea	Parborea28156	JX124610	NIWA	28156	2002	New Zealand, 1621/08, Z11008	920	-33.92	167.92	
Paragorgia arborea	Parborea28157	JX124531	NIWA	28157	1999	New Zealand, TRIP1223/29	959	-44.58	-177.88	
Paragorgia arborea	Parborea28158	JX124566	NIWA	28158	2001	New Zealand, AEX0101/80, Z10907	753	-44.74	-177.18	
Paragorgia arborea	Parborea28160	JX124523	NIWA	28160	2001	New Zealand				
Paragorgia arborea	Parborea28161	JX124568	NIWA	28161	2001	New Zealand, AEX0101/80, Z10920	753	-44.74	-176.81	Sanchez 2005
Paragorgia arborea	Parborea28392	JX124575	NIWA	28392	2001	New Zealand, AEX0101/80, Z10956	753	-44.74	-177.19	
Paragorgia arborea	Parborea28422	JX124601	NIWA	28422	2001	New Zealand, AEX0101/80, Z10920	753	-44.74	-176.81	
Paragorgia arborea	Parborea28425	JX124550	NIWA	28425	1997	New Zealand, TAN9713/52, Z8979	858	-44.45	-179.96	
Paragorgia arborea	Parborea3308	JX124552	NIWA	3308	2002	New Zealand, Z10987	1225	-33.93	167.92	Sanchez 2005
Paragorgia arborea	Parborea3309	JX124535	NIWA	3309	2002	New Zealand, Z11009	955	-33.93	167.91	Sanchez 2005
Paragorgia arborea	Parborea3310	GQ293311	NIWA	3310	1999	New Zealand, Z9862	687	-44.75	174.82	Sanchez 2005
Paragorgia arborea	Parborea3311	JX124525	NIWA	3311	1996	New Zealand, Southern Havre trough, X700	1525	-35.84	177.91	Sanchez 2005
Paragorgia arborea	Parborea33559	JX124559	USNM	33559	1878	Fishing Banks, North Carolina, USA	457	36.00	-74.00	
Paragorgia arborea	Parborea33561	JX124562	USNM	33561		Off NE North America, USA				
Paragorgia arborea	Parborea4089	JX124591	USNM	4089		Sable Island, 50 Mile E Of E Light, Nova Scotia, Canada	512	43.90	-58.80	
Paragorgia arborea	Parborea4091	JX124579	USNM	4091	1879	Banquereau Bank, Nova Scotia, Canada	366	44.58	-57.68	
Paragorgia arborea	Parborea4178B	JX124524	USNM	1123932	2002	South of Trinity Islands, Aleutian Islands	746	55.87	-154.06	
Paragorgia arborea	Parborea41829	JX124585	NIWA	41829	2006	New Zealand, TRIP2324/76	1044	-47.25	178.33	
Paragorgia arborea	Parborea41854	JX124611	NIWA	41854	2007	New Zealand, TRIP2494/10	867-986	-47.53	177.87	
Paragorgia arborea	Parborea41999	JX124588	NIWA	41999	2007	New Zealand, TRIP2551/50	1203-121	-44.50	-174.79	
Paragorgia arborea	Parborea42001	JX124574	NIWA	420 01	2007	New Zealand, TRIP2551/55	1283-139	-44.50	-174.82	
Paragorgia arborea	Parborea4238	JX124608	USNM	4238	1879	Banquereau Bank, South Of, Nova Scotia, Canada	457	43.90	-58.67	
Paragorgia arborea	Parborea4242	JX124567	USNM	4242	1879	Grand Banks, W Part Of, Newfoundland, Canada		45.00	-54.00	
Paragorgia arborea	Parborea44608	JX124537	NIWA	44608	2007	New Zealand, TRIP2506/81	1106-13:	-46.91	171.88	
Paragorgia arborea	Parborea44609	JX124534	NIWA	44609	2007	New Zealand, TRIP2506/135	870-967	-47.53	177.92	
Paragorgia arborea	Parborea4569	JX124596	USNM	4569	1879	Sable Island Bank, Nova Scotia, Canada	457	43.42	-60.00	
Paragorgia arborea	Parborea46320	JX124570	NIWA	46320	2008	New Zealand, TRIP2320/70	750-855	-46.48	170.60	
Paragorgia arborea	Parborea50890	JX124521	USNM	50890	1927	Burdwood Bank, S Of Falkland Islands, Scotia Sea		-54.50	-59.10	
Paragorgia arborea	Parborea80838	JX124544	USNM	80838	1979	Baltimore Canyon, Off Eastern Shore, Maryland, USA	480	38.17	-73.84	
Paragorgia arborea	Parborea80936	JX124543	USNM	80936	1979	Lydonia Canyon, Massachusetts, USA	680-370	40.38	-67.66	
Paragorgia arborea	Parborea80937	GQ293312	USNM	80937	1979	Lydonia Canyon, Massachusetts, USA	613-430	40.38	-67.66	
Paragorgia arborea	ParboreaJ2095271	JX124532	USNM	1123936	2004	Amlia Island, Andreanof Islands, Aleutian Islands	843	51.81	-173.83	
Paragorgia arborea	ParboreaJ2095272	JX124580	USNM	1123935	2004	Amlia Island, Andreanof Islands, Aleutian Islands	843	51.81	-173.83	
Paragorgia arborea	ParboreaJ2099211	JX124604	USNM	1123937	2004	Adak Canyon, Andreanof Islands, Aleutian Islands	1269	51.51	-177.04	
Paragorgia arborea	ParboreaJ210441	JX124536	USNM	1123938	2004	Amchitka Pass, Andreanof Islands, Aleutian Islands	747	51.72	-179.58	
Paragorgia arborea	ParboreaJ210462	JX124555	USNM	1123934	2004	Amchitka Pass, Andreanof Islands, Aleutian Islands	857	51.68	-179.58	
Paragorgia arborea	ParboreaMCZ1572	JX124606	MCZ	15721	2002	Atlantic Ocean	480	37.67	-74.65	

Paragorgia arborea
Paragorgia arborea
Paragorgia cf dendroides
Paragorgia cf dendroides
Paragorgia cf dendroides
Paragorgia cf johnsoni
Paragorgia cf regalis
Paragorgia coralloides
Paragorgia coralloides
Paragorgia coralloides
Paragorgia dendroides
Paragorgia johnsoni
Paragorgia johnsoni
Paragorgia johnsoni
Paragorgia johnsoni
Paragorgia kaupeka
Paragorgia maunga
Paragorgia maunga
Paragorgia maunga
Paragorgia maunga
Paragorgia regalis

	ParboreaMCZ2805	JX124551	MCZ	28057	2001	Nantucket Island, Massachussetts, USA	156	41.97	-65.87	
	ParboreaMCZ2869	JX124561	MCZ	28697	1878	off Maine, USA	245	42.60	-65.73	
	ParboreaMCZ5124	JX124600	MCZ	51244	1979	Georges Bank, Massachusetts, USA		41.00	-67.00	
	ParboreaMNHN04	11	MNHN	0411	1997	France, North Altlantic Ocean, France	700	48.83	-11.33	
	ParboreaMNHN04	JX124609	MNHN	412		off Norway		66.70	11.60	
	ParboreaMNHN04	JX124564	MNHN	422	2002	Trondhjems Fjord, Norway		63.50	10.50	
	Parboreap33560	JX124572	USNM	33560		Fishing Banks, North Carolina, USA		45.00	-53.50	
	ParboreaT661A10	JX124590	USNM	1122237	2002	Rodriguez Seamount, California, USA	894.5	34.06	-121.08	
	ParboreaT662A28	JX124522	USNM	1122240	2002	San Juan Seamount, California, USA	1362.9	32.97	-121.04	
	ParboreaT662A29	JX124586	USNM	1122233	2002	San Juan Seamount, California, USA	1360.8	32.97	-121.04	
	ParboreaZ11166	JX124533	NIWA	76238	2007	New Zeland	891	-44.44	175.54	
	ParboreaZC0706	JX124598	B. Stone	ZC0706ROV01		Zhemchung Canyon, Bering Sea	171	-47.54	177.93	
	PcfarboreJapan	JX124557	WPMNH		2005	Off Yaizu-shi, Shijuoka Prof., Japan	760-800	33.00	138.40	
	Ppacifica1016320	JX124581	USNM	1016320	2002	British Columbia, Canada	1152-119	53.70	-133.42	
	TC16 03		WHOI	HB1302 TC016	5 2013	NE US Canyons; Munson Canyon; TowCam 16	540	40.54	-67.01	
des	Pcfdendro98788		USNM	98788	1996	USA, Hawaii, Lanai Island, Keanapapa Point, SW of Point	1007	20.78	-157.15	
des	Pcfdendro98792		USNM	98792	1996	USA, Hawaii, Lanai Island, Keanapapa Point, SW of Point	1018	20.78	-157.15	
ides	PcfdendroT662A3)	USNM	1122228	200?	USA, California, San Juan Seamount	1237.6	32.97	-121.03	
ni		KC984606								
	Pcfregalis1072337		USNM	1072337	2003	USA, Hawaii, Pioneer Bank	1211	25.81	-173.50	
s	06390		WHOI	REH112-5	2005	New England Seamounts: Rehobot Seamount; H13	1821	37.46	-59.95	
s	12016		WHOI	leftover sieving	2003	New England Seamounts: Manning Seamount, station 4: AD38	2000	38.23	-60.46	
les	Pcfcorallo98785	JX128350	USNM	98785	1995	East Pacific Rise, off Mexico	1950	12.73	-102.60	
s	Pdendroid1072362		USNM	1072362	2003	USA, Hawaii, Necker Island, Seamount East of Island	1536	23.30	-163.70	
-	20136	KC984607	TU-WHO	0909-Oct1	2009	Gulf of Mexico; MC751; J2-464	438	28.19	-89.80	
	Paragosp100898		USNM	100898	2000	USA, North Atlantic Ocean, South of Georges Bank		39.86	-67.42	Sanchez 2005
	Pjohnsoni73767	JX128348	USNM	73767	1984	Little Bahama Bank, Bahamas	608	27.10	-79.70	Sanchez 2005
	2	KC788262								
	72152		NIWA	72152	2011	Kermadec Ridge; Clark Seamount, chimney field, north cone;	877	-36.45	177.84	
	82260		NIWA	82260	2012	Kermadec Ridge; Site SM3a, summit of Clark Seamount; TAN	850	-36.45	177.84	
	82342		NIWA	82342	2012	Kermadec Ridge; Site SM3a, Clark Seamount; TAN1206/40	1100	-36.45	177.84	
	84804		NIWA	84804	2005	Kermadec Ridge; Clark Seamount; KOK0506/12	870	-36.45	177.84	
	Pkaupeka3320	GQ293313	NIWA	3320	1989	New Zealand, X152	820	-36.16	176.81	type
	64980		NIWA	64980	2010	Kermadec Ridge; Silent II seamount; TAN1007/120	772	-35.17	178.89	
	Pmaunga28393		NIWA	28393	1999	New Zealand, Z9779	1121	-34.12	174.90	
	Pmaunga3322		NIWA	3322	2002	New Zealand, Wanganella Bank, TRIP, Z10989	1082	-33.89	167.94	type
	Pmaunga3323		NIWA	3323	1999	New Zealand, Z9779	1121	-34.12	174.90	type
	ParagospT629A6		USNM	1122239	200?	USA, California, Rodriguez Seamount	1843.8	33.95	-121.14	
	Pcfdendro98789		USNM	98789	1996	USA, Hawaii, Lanai Island, Keanapapa Point, SW of Point	1018	20.78	-157.15	
	PdendroidT630A5		USNM	1122192	200?	USA, California, Rodriguez Seamount	1031.8	34.00	-121.10	
	Pregalis1027063		USNM	1027063	2003	USA, California, Rodriguez Seamount, West of San Miguel Pa	1840	33.95	-121.14	
	Pregalis1072338		USNM	1072338	2003	USA, Hawaii, Laysan Island, SE of Island	1136	25.67	-171.41	
	Pregalis1072339		USNM	1072339	2003	USA, Hawaii, Pioneer Bank	1743.7	25.57	-173.51	
	Pregalis1072340		USNM	1072340	2004	USA, Hawaii, Pioneer Bank	1744.7	25.57	-173.51	
		GO293307	USNM	1014743	2003	Cross Seamount: Hawaii: USA	452	19.74	-158.30	Sanchez 2005
				-				-		

Paragorgia regalis		JQ241244							
Paragorgia sp	Paragosp1007354		USNM	1007354	2001	Canada, British Columbia, Vancouver Island	1168	48.44	-126.38
Paragorgia sp	Paragosp1011093		USNM	1011093	2002	USA, Bering Sea, Forrester Island	601-800	55.40	-134.83
Paragorgia sp	Paragosp1011094		USNM	1011094	2002	USA, Bering Sea, East of Kodiak Island	601-836	58.22	-148.70
Paragorgia sp	Paragosp1011095		USNM	1011095	2002	USA, Bering Sea, Alaska, Alexander Archipielago, Baranof Isl	601-715	57.19	-136.24
Paragorgia sp	Paragosp1027079		USNM	1027079		USA, California, Rodriguez Seamount, West of San Miguel Pa	1030	34.05	-121.10
Paragorgia sp	Paragosp1071233		USNM	1071233	2004	USA, Hawaii, Cross Seamount	410	18.73	-158.26
Paragorgia sp	Paragosp1071440		USNM	1071440	2004	USA, Hawaii, Cross Seamount	427.51	18.73	-158.26
Paragorgia sp	Paragosp1071441		USNM	1071441	2004	USA, Hawaii, Cross Seamount	427.51	18.73	-158.26
Paragorgia sp	Paragosp1072336		USNM	1072336	2003	USA, Hawaii, Raita Bank	573	25.63	-169.32
Paragorgia sp	Paragosp1075392		USNM	1075392	2004	USA, Gulf of Alaska, Alaska, Giacomini Seamount	733	56.42	-146.37
Paragorgia sp	Paragosp1075750		USNM	1075750	2004	USA, Gulf of Alaska, Alaska, Welker Seamount	718	55.05	-140.31
Paragorgia sp	Paragosp1075751		USNM	1075751	2004	USA, Gulf of Alaska, Alaska, Welker Seamount	782	55.05	-140.31
Paragorgia sp	Paragosp1075752		USNM	1075752	2004	USA, Gulf of Alaska, Alaska, Welker Seamount	1119	55.06	-140.41
Paragorgia sp	Paragosp1075755		USNM	1075755	2004	USA, Gulf of Alaska, Alaska, Welker Seamount	1050	55.06	-140.41
Paragorgia sp	Paragosp1075756		USNM	1075756	2004	USA, Gulf of Alaska, Alaska, Welker Seamount	1050	55.06	-140.41
Paragorgia sp	Paragosp1075757		USNM	1075757	2004	USA, Gulf of Alaska, Alaska, Pratt Seamount	1093	56.30	-142.47
Paragorgia sp	Paragosp1075759		USNM	1075759	2004	USA, Gulf of Alaska, Alaska, Pratt Seamount	1069	56.17	-142.70
Paragorgia sp	Paragosp1075764		USNM	1075764	2004	USA, Gulf of Alaska, Alaska, Pratt Seamount	920	56.17	-142.70
Paragorgia sp	Paragosp1075769		USNM	1075769	2004	USA, Gulf of Alaska, Alaska, Giacomini Seamount	730	56.42	-146.37
Paragorgia sp	Paragosp1082595		USNM	1082595	2002	USA, Gulf of Alaska, Alaska, Murray Seamount	855	53.89	-148.53
Paragorgia sp	Paragosp1082600		USNM	1082600	2002	USA, Gulf of Alaska, Alaska, Murray Seamount	1376	53. 9 9	-148.50
Paragorgia sp	Paragosp1082644		USNM	1082644	2002	USA, Washington, Warwick Seamount	768	48.05	-132.74
Paragorgia sp	Paragosp1200601		Stone-Sai	1200601	2006	Gulf of Alaska	741	53.98	-162.85
Paragorgia sp	Paragosp121	JX128347	USNM	1122305	2004	British Columbia, Canada		51.20	-130.14
Paragorgia sp	Paragosp1310601		Stone-Sai	1310601	2006	Gulf of Alaska	549	53.73	-163.98
Paragorgia sp	Paragosp159		USNM	1122302	2003	Canada, British Columbia			
Paragorgia sp	Paragosp165		USNM	1122303	2003	Canada, British Columbia, Brooks Penninsula			
Paragorgia sp	Paragosp27294		YPM	27294	2000	Bear Seamount		39.92	-67.44
Paragorgia sp	Paragosp28905		YPM	28905	2003	USA, Atlantic Ocean, Muir Seamount	1713	33.85	-62.66
Paragorgia sp	Paragosp28911		YPM	28911	2003	USA, Atlantic Ocean, Manning Seamount	1550	38.22	-60.46
Paragorgia sp	Paragosp28921		YPM	28921	2003	USA, Atlantic Ocean, Manning Seamount	1579	38.21	-60.53
Paragorgia sp	Paragosp28946		YPM	28946	2003	USA, Atlantic Ocean, Manning Seamount	1597	38.20	-60.53
Paragorgia sp	Paragosp291		USNM	1122306	2004	Canada, British Columbia			
Paragorgia sp	Paragosp34786		YPM	34786	2004	Atlantic Ocean, Manning Seamount, on summit	1336	38.22	-60.51
Paragorgia sp	Paragosp35381		YPM	35381	2004	USA, Manning Seamount, on summit	1337	38.22	-60.51
Paragorgia sp	Paragosp36781		YPM	36781	2004	Bear Seamount, South rim of table top peak	1428-165	39.87	-67.36
Paragorgia sp	Paragosp36910C		YPM	36910C	2004	Bear Seamount, South rim of table top peak	1428-165	39.87	-67.36
Paragorgia sp	Paragosp37148		YPM	37148	2003	USA, Manning Seamount	1483	38.26	-60.55
Paragorgia sp	Paragosp38636		YPM	38636	2004	Bear Seamount, South rim of table top peak	1428-16	39.87	-67.36
Paragorgia sp	Paragosp41849		NIWA	41849	2007	New Zealand, TRIP2494/97	1155.0	-47.0100	175.580000000
Paragorgia sp	Paragosp44607		NIWA	44607	2007	New Zealand, TRIP2506/24	858.0	-50.0516	174.7150000000
Paragorgia sp	Paragosp54830		USNM	54830	1974	USA, Florida, Straits of Florida, Off Delray Beach	743-761	26.38	-79.60
Paragorgia sp	Paragosp56615		USNM	56615	1974	USA, Florida, Straits of Florida, Off Mile Beach	770-660	25.71	-79.79
Paragorgia sp	Paragosp56690		USNM	56690		USA, Hawaii, Oahu Island, Makapuu Point	366		

Paragorgia sp	Paragospn32830		NIWA	32830	2005	New Zealand, Tangaroa Seamount, P629-4B, KOK0507/4	790	-36.3284	178.0359	955
Paragorgia sp	ParagospT665A3		USNM	1122225	200?	USA, California, San Juan Seamount	706.6	33.11	-120.96	
Paragorgia sp	Parborea28423		NIWA	28423	1998	New Zealand, 1152/48, Z9275	660	-34.18	162.65	
Paragorgia sp	Parborea28717		ҮРМ	28717	2003	Manning Seamount	1372	38.22	-61.51	
Paragorgia sp	PcfaoteaT668A3		USNM	1122226	200?	USA, California, Little Joe Seamount	2397.8	31.89	-120.05	
Paragorgia sp	PcfaoteaT669A15		USNM	1122227	200?	USA, California, San Marcos Seamount	2061.7	32.64	-121.51	
Paragorgia sp	PcfstepheT665A4		USNM	1122235	200?	USA, California, San Juan Seamount	720.5	33.11	-120.96	
Paragorgia sp	PcfwahineT627A3		USNM	1122238	200?	USA, California, Pioneer Seamount	1771.5	37.40	-123.44	
Paragorgia sp	PcfwahineT630A6		USNM	1122234	200?	USA, California, Rodriguez Seamount	1031.9	34.00	-121.10	
Paragorgia sp	PcfvutliT663A16J4		USNM	1122231	200?	USA, California, Northeast Bank	546.6	32.32	-119.61	
Paragorgia sp		KC788261								
Paragorgia stephencairnsi	2344		RBCM	010-00234-004	2004	British Columbia; VE13978	1194	53.37	-133.31	
Paragorgia stephencairnsi	101010		USNM	1157074	2008	California; Piggy Bank, southern California; DW-026-02	283	33.92	-119.47	
Paragorgia stephencairnsi	1007316		USNM	1007316	2001	British Columbia; Vancouver Island	1168	48.44	-126.38	
Paragorgia stephencairnsi	1124300		USNM	1124300	2006	British Columbia; Vancouver Island, Ohiat Island; OC 06/952	188	48.83	-125.13	
Paragorgia stephencairnsi	Agam		WHOI	Agam	2012	British Columbia; Agamemnon Channel	32	49.72	-124.05	
Paragorgia stephencairnsi	C02		WHOI	C02	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03	
Paragorgia stephencairnsi	C03		WHOI	C03	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03	
Paragorgia stephencairnsi	C04		WHOI	C04	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03	
Paragorgia stephencairnsi	C05		WHOI	C05	2013	British Columbia; Agamemnon Channel; Dive02	41	49.74	-124.03	
Paragorgia stephencairnsi	C100		WHOI	C100	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	
Paragorgia stephencairnsi	C101		WHOI	C101	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	
Paragorgia stephencairnsi	C102		WHOI	C102	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	
Paragorgia stephencairnsi	C104		WHOI	C104	2013	British Columbia; Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	
Paragorgia stephencairnsi	FOC25		DFO	25	2012	British Columbia; W of Graham Island; 2012-65	204	53.31	-133.03	
Paragorgia stephencairnsi	FOC26		DFO	26	2012	British Columbia; W of Graham Island; 2012-65	221	53.30	-133.04	
Paragorgia stephencairnsi	FOC30		DFO	30	2012	British Columbia; W of Graham Island; 2012-65	318	53.48	-133.07	
Paragorgia stephencairnsi	FOC5		DFO	5	2009	British Columbia; E of Graham Island; 2009-47	201	52.13	-128.90	
Paragorgia stephencairnsi	Paragosp1075741		USNM	1075741	2004	USA, Gulf of Alaska, Alaska, Dickins Seamount	751	54.55	-136.84	
Paragorgia stephencairnsi	Paragosp1092785		USNM	1092785	2005	USA, North Pacific Ocean, Alexander Archipielago, Baranof I	171	56.19	-135.10	
Paragorgia stephencairnsi	Paragosp200601		Stone-Sai	200601106B01	2006	Gulf of Alaska	479	55.40	-134.83	
Paragorgia stephencairnsi	Paragosp77		USNM	1122304	2004	Canada, British Columbia, Brooks Penninsula		49.94	-128.06	
Paragorgia stephencairnsi	Parborea41106		USNM	1123931	2004	USA, Bering Sea, GOA	417	55.37	-134.78	
Paragorgia stephencairnsi	Parborea411081		USNM	1123930	2004	USA, Bering Sea, GOA	427	54.47	-133.97	
Paragorgia stephencairnsi	ParboreaNew100		USNM	1124301	2006	Canada, British Columbia, Vancouver Island, Barkley Sound	270.22	48.26	-125.01	
Paragorgia stephencairnsi	ParboreaNew545		USNM	1124298	2006	Canada, British Columbia, Vancouver Island, Barkley Sound	309.5	48.15	-125.07	
Paragorgia stephencairnsi			CAS	190438	2010	California, Farallon Escarpement	424	37.74	-123.19	
Paragorgia wahine	Pwahine3326	GQ293314	NIWA	3326	2001	New Zealand, Diabolical seamount, TAN0104/113	900	-42.79	179.99	type
Paragorgia whero	Pwhero3436		NIWA	3436	1 998	New Zealand, TRIP1171/12, Z9583	935	-48.03	166.10	type
Paragorgia yutlimux	Pyutlinux1073480	GQ293315	USNM	1073480	2003	Off Vancouver Isl., British Columbia, Canada	846-861	50.23	-128.58	type
Paraminabea aldersladei		JX203767								
Paraminabea aldersladei		KF915662								
Paraminabea aldersladei		KF915665								
Sibogagorgia cauliflora	2036	KC984605	TU-WHO	2036-Oct1	2009	Gulf of Mexico; DC583; J2-454	2440	28.39	-87.39	
Sibogagorgia cauliflora	1122230	GQ293310	USNM	1122230	2006	California; Davidson seamount; dive 945	2502	35.83	-122.61	type

Sibogagorgia cauliflora	Sibogasp1T947A9 GQ293317 USNM	1122229	2006	Davidson Seamount, California, USA	3042.4	35.63	-122.83	
Sibogagorgia cauliflora	GQ293308 USNM	1081143	2004	Derickson Seamount: Alaska: USA	2766	52.98	-161.25	type
Sibogagorgia cauliflora	GQ293309 USNM	54831	1968	Straits of Florida: Havana: Cuba	1638-13	23.55	-82.78	type
Sibogagorgia dennisgordoni	GQ293316 NIWA	3329	1998	New Zealand, 1124/70, Z9228	820	-36.69	176.46	type
Sphaerasclera flammicerebra	JX203765							51

Acronyms as follows: National Museum of Natural History, Smithsonian Institution, USA (USNM); The National Institute of Water and Atmospheric Research, New Zealand (NIWA); Museum of Comparative Zoology, Harvard University, USA (MCZ); Museum National d'Histoire Naturelle, Paris, France (MNHN); Senckenberg Research Institute And Natural History Museum Frankfurt, Germany (SMF); Uppsala University Evolutionsmuseet, Sweden (UUZM); Wakayama Prefectural Museum of Natural History, Japan (WPMNH); Yale Peabody Museum of Natural History, USA (YPM), Department of Fisheries and Oceans Canada (DFO), Woods Hole Oceanographic Institution (WHOI), Temple University (TU), Royal British Columbia Museum (RBCM), California Academy of Sciences (CAS)

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CHAPTER 6

The genomics of adaptation potential of deep-sea corals to environmental changes

ABSTRACT

Species that live in a wide range of environmental conditions constitute natural experiments of biological adaptations, which can help us to understand possible ecological consequences of environmental changes on ecosystems. A few populations of some deep-sea coral species can be found in shallow (< 45 m) highlatitude fjord environments, where they experience significantly different environmental conditions than their deep relatives. Therefore, these shallow-water populations are believed to inhabit the extremes of the species' physiological tolerances and likely have developed adaptations that enable them to colonize these shallow-water environments. Here, we aim to identify genomic regions that may have enabled the successful adaptation to shallow-water in the deep-sea octocoral species Paragorgia stephencairnsi. To characterize the genome-wide genetic diversity of populations of P. stephencairnsi found in shallowwater populations and compare it to the genetic diversity from deep-water populations, we performed high-resolution genome-wide scans of single nucleotide polymorphisms. We find patterns of significant population genetic differentiation among the examined populations of P. stephencairnsi, which are consistent with the hypothesis that larvae from outer deep populations seeded shallow-water inner fjord populations. Furthermore, we find candidate positive-selection markers shared between parallel comparisons of two shallow populations and a deep populations, and thus identify them as likely candidate makers for genomic regions involved in adaptation to the shallow-water fjord environment. This study lays groundwork for describing the impacts of natural selection on deep-sea coral species in the face of environmental changes.

INTRODUCTION

Species that live in a wide range of environmental conditions constitute natural experiments of biological adaptations, which can help us to understand possible ecological consequences of environmental changes on ecosystems (e.g. Emerson *et al.* 2010). Populations of deep-sea coral species typically live hundreds or

thousands of meters below the surface in the relatively uniform environmental conditions characteristic of the deep-sea (Roberts *et al.* 2009). However, a few populations of some deep-sea coral species can be found in shallow (< 45 m) high-latitude fjord environments, where low light penetration and cool temperatures presumably create suitable living conditions similar to those found in the deep-sea. Despite the seeming similarities between deep-sea and shallow fjord environments there are significant differences that make these shallow fjords a novel environment for deep-sea species, namely lower hydrostatic pressure and significantly greater ranges of variability for temperature, pH, salinity, current speeds, and sedimentation rates. Therefore, these shallow populations of predominantly deep-sea coral species are believed to inhabit the extremes of the species' physiological tolerances.

Differences in environmental conditions over space and time can have strong selective effects on natural populations by modifying the survival and reproductive success of individuals, and thus altering genetic composition of the populations and their ability to respond to environmental changes (e.g., Prada & Hellberg 2013). When isolated populations of a species are exposed to similar selective pressures, e.g. deep-sea coral populations in semi-enclosed shallow environments such as inlets and fjords, they tend to develop similar solutions to common challenges - a process known as parallel adaptation (e.g., Chan et al. 2010; Hohenlohe et al. 2010; Jones et al. 2012; Miller et al. 2012). Hohenlohe et al. (2010) present an exemplary case exploring the genome-wide consequences of this evolutionary process. In that study the authors investigate the parallel adaptation to freshwater environments in marine stickleback populations by performing high-resolution genomic scans of single nucleotide polymorphisms (SNPs) from ancestral marine and derived freshwater populations. When genome-mapped data from marine populations are compared to freshwater populations it is possible to detect specific regions in the genome that had more differentiation than what is expected under neutrality. Common regions of differentiation across multiple populations are identified as strong candidates for parallel adaptation of populations of a marine species to freshwater. Hohenlohe et al. (2010) show that several genomic regions identified through this method co-localize with previously identified quantitative trait loci (regions of the genome that account for particular observable characteristics of organisms), thus demonstrating the usefulness of this approach for the identification of ecologically important genes.

Bubblegum corals (Paragorgiidae, Octocorallia) are among the most abundant and widely distributed benthic foundation species in deep-water ecosystems worldwide (Roberts *et al.* 2009; Wating *et al.* 2011). They play an important ecological role, akin to the structural role of large trees in a rainforest, by generating three-dimensional habitats for a great number of micro- and macro-organisms (Buhl-Mortensen & Mortensen 2004; Auster *et al.* 2005; Buhl-Mortensen & Mortensen 2005; DeVogelaere *et*

al. 2005; Nedashkovskaya et al. 2005). Populations of the bubblegum coral species Paragorgia stephencairnsi Sánchez, 2005 (sensu Herrera and Shank (Chapter 5)) are typically found in the ocean at depths greater than 200 meters, attached to hard grounds on the continental shelf and slope, and seamounts along the western coast of North America (Sánchez 2005). A few populations of this species inhabit shallow fjords (as shallow as 30 meters) at the northern boundary of its distribution. Recent phylogenomic evidence shows *P. stephencairnsi* evolved from deep-sea ancestors (Herrera & Shank Chapter 5). Thus, the most parsimonious scenario is that shallow-water populations of *P. stephencairnsi* in high-latitude fjords originated from colonization seeded by deeper populations.



Figure 1. a) Geographic location of *P. stephencairnsi* populations off British Columbia, Canada. Left map shows sites where samples have been obtained. Purple stars indicate locations of shallow populations (accesible via scuba diving). Red flags indicate the locations of deep populations. Orange dots indicate observatory nodes from the NorthEast Pacific Time-Series Undersea Networked Experiments (NEPTUNE) monitoring environmental variables

at 23m, 100m, 892m, and 1280m depths. **Right** map shows the bathymetry of the glacially formed Agamemnon Channel, one of the primary shallow-water fjord collection sites. **b**) Temperature records from monitoring nodes at four different depths off the North Western coast of North America. Location of nodes shown in Fig. 3a. Data courtesy of NEPTUNE ocean network observatory (Canada).

In general, fjords –long and narrow coastal sea inlets formed by glaciers – can be thought of as semienclosed marine basins as they are limited at their sides by landmasses. The seafloor in fjords dramatically shoals at their mouths because of the paleoglaciers' terminal moraines (see Agamemnon Channel example in Fig. 1a). The geological characteristics of fjords constrain the circulation and exchange between these semi-enclosed environments and the 'outer' ocean environment. Thus, the amounts of gene flow between 'outer' deep and 'inner' shallow populations of marine organisms, such as *Paragorgia stephencairnsi*, are likely limited by the these same geomorphic features.

The main goal of this study is to identify and characterize potential genomic regions that have enabled the successful colonization of shallow-water environments from the deep-sea by *P. stephencairnsi*. We hypothesize that semi-isolated shallow water populations of *P. stephencairnsi* in different fjord systems have independently evolved adaptations in parallel to cope with distinct highly variable conditions of the surface ocean. As a result the expectation is that, in each population, the environment has independently selected common sets of genetic diversity. This would leave characteristic signatures of parallel differentiation in their genomes. Here, we perform high-resolution genome-wide scans of SNPs to characterize the genome-wide genetic diversity of populations of *P. stephencairnsi* found in shallow-water populations, and compare it to the genetic diversity from deep-water populations. Through these comparisons we identify patterns of differentiation that would be indicative of non-random evolutionary processes of natural selection and adaptation and lay groundwork for describing the impacts of natural selection on deep-sea coral species in the face of environmental changes.

METHODS

Specimens of *P. stephencairnsi* from shallow water populations (less than 45m; Agamemnon Channel and Tahsis Strait) were collected in May 4-17, 2013 in British Columbia, Canada, during 8 decompression SCUBA dives. Specimens from deep populations around Vancouver Island were collected in various oceanographic expeditions by collaborators at Memorial University, the Department of Fisheries and Oceans Canada, the Royal British Columbia Museum, and the National Oceanographic and Atmospheric Administration (Table 1).

ID	Collection	Catalog Number	Date	Locality	Depth (m)	Lat.	Lon.	Population
Agam	WHOI	Agam	2012	Agamemnon Channel	32	49.72	-124.05	Agamemnon
C02	WHOI	C02	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C03	WHOI	C03	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C04	WHOI	C04	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C05	WHOI	C05	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C07	WHOI	C07	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C08	WHOI	C08	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C11	WHOI	C11	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C12	WHOI	C12	2013	Agamemnon Channel; Dive02	41	49.74	-124.03	Agamemnon
C100	WHOI	C100	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C101	WHOI	C101	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C102	WHOI	C102	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C103	WHOI	C103	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C104	WHOI	C104	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C105	WHOI	C105	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C106	WHOI	C106	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
C107	WHOI	C107	2013	Vancouver Island, Tahsis Inlet; Dive07	40	49.86	-126.67	Tahsis
FOC25	DFO	25	2012	W of Graham Island; 2012-65	204	53.31	-133.03	Deep
FOC26	DFO	26	2012	W of Graham Island; 2012-65	221	53.30	-133.04	Deep
FOC30	DFO	30	2012	W of Graham Island; 2012-65	318	53.48	-133.07	Deep
FOC5	DFO	5	2009	E of Graham Island; 2009-47	201	52.13	-128.90	Deep
L139	MU	R1513-L1-0039		S of Texada Island	268	49.50	-124.17	Deep
L219	MU	R1513-L2-0019		S of Texada Island	267	49.50	-124.17	Deep
L341	MU	R1513-L3-0041		S of Texada Island	268	49.50	-124.17	Deep
PR27	MU	R1513-PR-0027		S of Texada Island	270	49.50	-124.17	Deep
1122479	USNM	1122479	2008	LaPush, west of, Washington	269	48.13	-125.10	Deep
1124300	USNM	1124300	2006	Vancouver Island, Ohiat Island	188	48.83	-125.13	Deep

Table 1. Collection information for the specimens used in this study.

Acronyms as follows: National Museum of Natural History, Smithsonian Institution, USA (USNM); Department of Fisheries and Oceans Canada (DFO), Woods Hole Oceanographic Institution (WHOI), Memorial University (MU).

Molecular laboratory methods

To characterize the genome-wide genetic diversity of populations of *P. stephencairnsi*, we performed high-resolution genomic scans and identified single nucleotide polymorphisms (SNPs) from restriction site-associated DNA markers (RAD tags) (Baird *et al.* 2008; Hohenlohe *et al.* 2010). In short, the RAD sequencing method consists of: 1) the digestion of genomic DNA for each individual with a restriction enzyme; 2) ligation of the resulting fragments to sequencing adapters with unique barcodes for each individual; 3) size-selection and enrichment of the fragments successfully ligated to the adapters; and 4) sequencing via a high-throughput platform (Illumina HiSeq 2000). We performed RAD sequencing with the 6-cutter restriction enzyme PstI, which is predicted to cut between 32,000 and 110,000 times in the

genome of an octocoral (Herrera & Shank Chapter 5). This predicted range was obtained using the observed frequency of the PstI recognition sequence, and its probability using a trinucleotide composition model, in the genomes of the cnidarians *Nematostella vectensis, Acropora digitifera, Hydra vulgaris,* and *Alatina moseri* (Herrera *et al.* Chapter 2). Genome size range of 0.3-0.5 pg was used based on observations obtained through flow cytometry in gorgoniid octocorals by Luisa Dueñas at the Universidad de los Andes, Bogotá, Colombia (personal communication).

Total genomic DNA was purified from specimens as in Herrera and Shank (Chapter 5) by: (1) digesting the tissue in 2% CTAB buffer (Teknova) with proteinase K and RNAse A/T1 (Fermentas) for 1 hour, (2) separating nucleic acids with chloroform: isoamyl alcohol (24:1) (Fermentas) and phenol: chloroform: isoamyl alcohol (25:24:1, Tris buffered at pH 8.0) (Fermentas), (3) precipitating nucleic acids with 100% ethanol (1:1 volume ratio), and (4) washing the precipitate twice with 70% ethanol. Concentration-normalized genomic DNA was submitted to Floragenex Inc. (Eugene, OR) for library preparation and RAD sequencing. Libraries were sequenced by 48-multiplex, using 10-base pairs long barcodes, on a single lane of an Illumina Hi-Seq 2000 sequencer.

Data filtering

Sequence reads were de-multiplexed and quality filtered with the *process_radtags* program from the package Stacks v1.19 (Catchen *et al.* 2013b). Barcodes and Illumina adapters were excluded from each read and length was truncated to 91bp (-t 91) Reads with ambiguous bases were discarded (-c). Reads with an average quality score below 10 (-s 10) within a sliding window of 15% of the read length (-w 0.15) were discarded (-r). The rescue barcodes and RAD-tags algorithm was enabled (-r).

De novo loci assembly

We performed *de novo* assemblies of RAD loci using the *denovo_map* pipeline in Stacks. A minimum depth of three reads per stack was enforced (-m 3). Significantly high-repetitive stacks were discarded by implementing the deleveraging algorithm (-t), as these likely represent sequencing errors, duplications or repetitive regions. No mismatches among loci were allowed when creating the catalog of all the loci identified among the sampled individuals (-n 0). A maximum number of two mismatches was allowed among loci within each individual (-M 2). The maximum number of stacks at a single locus was set to three (--max locus stacks 3).

Demographic Inferences

We estimated population genetic descriptive statistics per SNP (nucleotide diversity π , proportion of polymorphic loci, observed heterozygosity, minor allele frequency, number of private alleles, inbreeding index F_{IS} and population differentiation index F_{ST}) using the program *populations* of Stacks. We only analyzed loci that were present in all populations of each species (-p) and in all individuals in each population (-r). We calculated population F_{ST} values utilizing a *p_value* filter (-f) to keep only significant estimates (α =0.05). We exported SNP data in *genpop* format, keeping only one SNP per locus to avoid violating the assumption of independence among markers in downstream analyses.

To summarize the variation in the SNP data among individuals and populations we a performed principal component analyses (PCA) as in (Reitzel *et al.* 2013), using the program *smartpca* from the package Eigensoft v5.0 (Patterson *et al.* 2006; Price *et al.* 2006). We evaluated the significance of the identified principal components through Tracy-Widom statistics (Tracy & Widom 1994; Johnstone 2001) and evaluated the statistical significance of the differences between populations with a chi-square test.

Candidate adaptation markers

We identified markers linked to candidate genomic regions involved in adaptation to shallow water environments from the deep-sea by detecting F_{ST} outliers (i.e., allelic frequencies between deep and shallow populations that show greater differentiation than expected under a neutral model of evolution, characterized by the accumulation of random mutations that do not affect survival and reproduction of organisms) with the program LOSITAN (Beaumont & Nichols 1996; Antao *et al.* 2008). Population genomics theory predicts that these outlier variants will be indicative of genomic regions containing genes or regulatory elements that have been subject to natural selection (Lewontin & Krakauer 1973; Beaumont & Balding 2004). SNP positions with outlier F_{ST} values (those above the 97.5 percentile of the neutral distribution of F_{ST}) were considered indicative of loci subject to natural selection. We considered outliers shared among shallow populations as indicative of parallel adaptations to the shallow-water environment, whereas outliers unique to a particular population were considered as likely indicators of local adaptations (see Hohenlohe *et al.* 2010).

Genomic sequences of markers under potential positive selection were scanned for functionality by querying against annotated databases of gene models from cnidarian genomes (*Nematostella vectensis* and *Hydra magnipapillata*) using BLAST searches at the U.S. National Center for Biotechnology Information (NCBI) databases. Gene Ontology (GO) and Kyoto Encyclopedia of Genes and Genomes

(KEGG) pathway assignments were attempted for each gene-model match using the program Blast2GO (Conesa *et al.* 2005).

Population	RAD-seq data file ID	Total sequenced reads	Retained reads	Percentage of retained reads after filtering	Number of Stacks	Mean coverage depth	S.D. of coverage depth
Agamemnon	PAR_Agam_BC	4,322,564	4,062,469	94.0	142,874	25.3	151.6
Agamemnon	PAR_C02_BCS	2,149,938	1,594,818	74.2	110,403	10.8	35.7
Agamemnon	PAR_C03_BCS	2,815,327	2,264,951	80.5	126,442	13.7	73.8
Agamemnon	PAR_C04_BCS	3,240,713	2,754,373	85.0	133,631	15.9	87.3
Agamemnon	PAR_C05_BCS	3,465,396	2,972,548	85.8	136,642	16.9	104.3
Agamemnon	PAR_C07_BCS	2,827,394	2,446,736	86.5	129,688	14.5	59.6
Agamemnon	PAR_C08_BCS	2,552,442	2,104,566	82.5	125,331	12.5	50.9
Agamemnon	PAR_C11_BCS	2,763,403	2,286,010	82.7	128,412	13.6	58.6
Agamemnon	PAR_C12_BCS	2,390,479	2,062,541	86.3	124,553	12.5	53.5
Deep	PAR_1122479_BC	4,607,551	4,331,474	94.0	142,445	27.2	108.6
Deep	PAR_1124300_WA	3,843,250	3,627,391	94.4	136,241	23.7	152.2
Deep	PAR_FOC25_BCD	5,479,109	4,693,804	85.7	152,378	24.2	102.9
Deep	PAR_FOC26_BCD	5,242,949	4,528,629	86.4	210,583	14.9	76.2
Deep	PAR_FOC30_BCD	4,417,520	3,781,973	85.6	174,618	15.8	116.6
Deep	PAR_FOC5_BCD	2,834,377	2,455,041	86.6	135,156	13.3	129.2
Deep	PAR_L139_BCD	2,928,518	2,392,739	81.7	129,058	14.1	62.6
Deep	PAR_L219_BCD	1,752,674	1,289,295	73.6	96,904	9.5	51.0
Deep	PAR_L341_BCD	3,253,979	2,761,200	84.9	133,647	16.1	75.8
Deep	PAR_PR27_BCD	3,536,670	2,942,112	83.2	135,368	17.0	72.3
Tahsis	PAR_C100_BCS	5,998,914	4,984,766	83.1	149,979	26.1	90.1
Tahsis	PAR_C101_BCS	5,332,619	4,652,535	87.2	150,525	24.2	142.6
Tahsis	PAR_C102_BCS	4,347,757	3,710,381	85.3	141,586	20.5	73.7
Tahsis	PAR_C103_BCS	2,272,200	1,706,852	75.1	114,196	11.0	38.5
Tahsis	PAR_C104_BCS	2,720,994	2,203,568	81.0	124,966	13.5	56.5
Tahsis	PAR_C105_BCS	2,778,573	2,003,812	72.1	123,158	12.1	45.9
Tahsis	PAR_C106_BCS	2,565,150	2,116,517	82.5	123,271	13.1	59.1
Tahsis	PAR_C107_BCS	2,697,054	2,189,475	81.2	125,420	13.3	45.6
	AVERAGE	3,449,538	2,922,984	83.7	135,462	16.5	80.5
	S.D.	1,136,671	1,066,781	5.6	21,081	5.2	34.5

Table 2. RAD sequencing results, filtering and de novo assebly statistics.

RESULTS

RAD-seq produced high-quality sequence data

We generated restriction site associated DNA sequence (RAD-seq) data for 27 individuals of *P*. *stephencairnsi* collected in the British Columbia region (Table 2). We obtained approximately 3.5 ± 1.1

(mean \pm standard deviation) million sequence reads per individual (100bp length), with individual values ranging from 1.75 to 6.0 million reads. Approximately 83.7 \pm 5.6% of these were retained after quality filters. *De novo* loci assemblies produced approximately 135 \pm 21 thousand unique sequence stacks per individual, with a mean coverage depth of 16.5 \pm 5.2X.

Table 3. Population genetic statistics calculated from only variant positions, and from both variant and fixed positions. Values indicate means \pm standard deviation

Variant positions

Population	Private alleles	Variant sites	% polym. sites	Major allele frequency	Observed heterozygosity	Nucleotide diversity (π)	Fis
Agamemnon	2,713	17,074	65.23	0.8737±0.1517	0.2149±0.2604	0.1858 ± 0.1884	-0.0612±0.2404
Tahsis	2,000	17,071	55.54	0.8793±0.1578	0.2000±0.2737	0.1743±0.1975	-0.0468±0.2717
Deep	3,046	17,062	69.35	0.8697±0.1523	0.2102±0.2565	0.1904±0.1865	-0.0340±0.2851

All positions (variant and fixed)

Population	Private alleles	Sites	% polym. sites	Major allele frequency	Observed heterozygosity	Nucleotide diversity (π)	Fis
Agamemnon	2,713	1,873,995	0.59	0.9988 ± 0.0200	0.0020±0.0316	0.0017±0.0245	-0.0006±0.0245
Tahsis	2,000	1,873,999	0.51	0.9989 ± 0.0200	0.0018±0.0316	0.0016±0.0245	-0.0004±0.0265
Deep	3,046	1,873,984	0.63	0.9988 ± 0.0200	0.0019±0.0316	0.0017±0.0245	-0.0003±0.0283



Figure 2. Principal component analysis of genetic variation from SNP in *Paragorgia stephencairnsi* from British Columbia. Each dot represents an individual. Colors indicate the source population: Agamemnon (red), Tahsis (yellow), and Deep (blue). The three principal axes of variation (eigenvectors) are shown.

Significant population differentiation among all populations

There were 10,920 loci shared among all individuals in all populations. These loci contained over 17 thousand SNPs, for an average of 1.56 SNP per locus (Table 3). The three largest axes of variation (eigenvectors) identified from the principal components analysis of SNP data revealed clear separation between the two shallow-water populations Tahsis and Agamemnon, but evident overlap between the individuals from the Deep population and both of the shallow (Fig. 3). All eigenvectors were statistically significant (*P*<0.001, *P*=0.002, and *P*=0.031 for eigenvectors 1, 2 and 3, respectively; α =0.05). All differences among populations were also statistically significant (Agamemnon vs. Tahsis: χ^2 =40.1, *P*<0.001; Agamemnon vs. Deep: χ^2 =16.4, *P*=0.003; Tahsis vs. Deep: χ^2 =30.3, *P*<0.001; α =0.05). Mean pairwise F_{ST} values indicate that population differentiation is significantly greater between the two shallow-water populations Tahsis and Agamemnon (F_{ST}=0.0519, *P*<0.05), than between either shallow-water population and the Deep population (Agamemnon vs. Deep: F_{ST}=0.0354, *P*<0.05; Tahsis vs. Deep: F_{ST}=0.0430, *P*<0.05).

Differences in nucleotide diversity between shallow and deep-populations

Summary population genetic statistics (private alleles, percentage of polymorphic sites, and nucleotide diversity) calculated from variant positions revealed higher diversity in the Deep population than in either of the shallow-water populations. These same metrics indicated that the Tahsis shallow-water population has the lowest genetic diversity of all three examined populations. Differences in population genetic diversity were not evident when summary population genetic statistics were calculated from both fixed and variant positions.

Differences in allele frequency distributions between shallow and deep populations

Minor allele frequency spectra show that a majority of the alleles in each population have low frequencies (Fig. 2), as expected for population near mutation-drift equilibrium. Shallow-water populations show a small modal shift towards higher allelic frequencies (particularly in Tahsis) and a noticeable increase of intermediate frequency alleles, compared to the deep population. F_{IS} distributions for all populations were centered on zero, with a tendency to negative values, indicating random mating in populations and a slight excess of heterozygotes.

Candidate adaptation markers

Outlier analyses between shallow-water populations and the Deep population found 733 SNPs candidate positive-selection markers when comparing Agamemnon vs. Deep, and 261 when comparing Tahsis vs.

Deep. Of these, 63 SNPs candidate positive selection markers were shared in both comparisons, and thus are considered candidate makers for adaptive genomic regions.



Figure 3. Frequency distributions of population genetic summary statistics for populations of *Paragorgia stephencairnsi* from British Columbia. **a)** Minor allele frequency spectra for SNPs loci. **b)** Frequency spectra of Wright's inbreeding index F_{IS} values calculated for each SNP locus. Colors indicate the source population: Agamemnon (red), Tahsis (yellow), and Deep (blue).

From the markers identified as potentially adaptive, 16 produced Blast matches to the non-redundant nucleotide NCBI database (Table 4). None produced matches to available protein or gene ontology databases. Many of the matches corresponded to coding DNA sequences of unknown function and uncharacterized mRNAs. The top three marker matches, in terms of Blast E-values (the number of expected false-positive matches in a database of given size), correspond to mRNAs of *Hydra magnipapillata:* a mitogen-activated protein kinase 7-like mRNA (RAD locus catalog 17484) potentially involved in cellular cycles and cell differentiation; a phosphoribosylformylglycinamidine synthase partial mRNA (RAD locus catalog 23843) potentially involved in purine metabolism; and a zinc finger protein

41-like mRNA (RAD locus catalog 83962) potentially a transcription factor associated with meiosis in spermatogenesis.

Table 4. Blast results for the candidate positive selection markers shared between comparisons of shallow-water populations vs. Deep population. Only results that produced hits in the non-redundant nucleotide NCBI database are shown.

RAD locus catalog name	Blast hit species	Blast hit sequence description	Blast E- value	Blast hit accession	Blast hit % similarity
17484	Hydra magnipapillata	mitogen-activated protein kinase 7-like partial mrna	1.00E-08	XM004210406	75
23843	Hydra magnipapillata	phosphoribosylformylglycinamidine synthase partial mrna	3.60E-08	XM001641610	77
83962	Hydra magnipapillata	zinc finger protein 41-like mrna	2.80E-03	XM002162877	80
57130	Hydra magnipapillata	uncharacterized loc100207636 partial mrna	9.60E-03	XM002158158	79
26651	Nematostella vectensis	protein partial mrna	9.60E-03	XM001627193	83
843	Nematostella vectensis	gene for complete cds	1.20E-01	BR000671	79
36131	Nematostella vectensis	protein partial mrna	4.10E-01	XM001629268	81
86101	Hydra magnipapillata	uncharacterized loc101234456 mrna	1.40E+00	XM004211848	78
101291	Hydra magnipapillata	uncharacterized loc101240885 mrna	1.40E+00	XM004206580	80
22838	Hydra magnipapillata	lysine-specific demethylase 6a-like mma	1.40E+00	XM002167270	83
67285	Hydra magnipapillata	uncharacterized loc100199733 mrna	1.40E+00	XM002159843	80
76378	Hydra magnipapillata	uncharacterized loc100202739 mrna	5.00E+00	XM002162521	77
103847	Hydra magnipapillata	uncharacterized loc100207904 mma	5.00E+00	XM002160483	75
71485	Hydra magnipapillata	uncharacterized loc100209560 mma	5.00E+00	XM002160476	78
8691	Nematostella vectensis	protein partial mrna	5.00E+00	XM001627130	80
27970	Nematostella vectensis	protein complete cds	5.00E+00	XM001639122	79

DISCUSSION

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Fjord environments effectively isolate marine populations

In this study, we find significant population genetic differentiation among all examined populations of *Paragorgia stephencairnsi*. This genetic differentiation is particularly marked between the shallow-water fjord populations as indicated by the lack of overlap among groups in the principal components analysis (PCA) and the elevated population differentiation F_{ST} value. Our results indicate that fjords act as semienclosed basins effectively isolating coral populations living within them. The evidence we present conforms to previous results from population genetic studies in pelagic and benthic organisms such as calanoid copepods (Bucklin *et al.* 2000), glacier lanternfish (Suneetha & Salvanes 2001), sea stars (Perrin *et al.* 2004), Pacific cod (Cunningham *et al.* 2009), sprat (Glover *et al.* 2011), and Pacific herring (Wildes *et al.* 2011), which show significant population genetic structuring in small spatial scales among fjord populations, and with respect to open ocean populations.

Shallow-water fjord populations originated from the deep

Taken together, our results are consistent with the hypothesis that larvae from outer deep populations seeded shallow-water inner fjord populations. Smaller genetic diversity found in the shallow-water fjord populations relative to those from the deep population, indicated by smaller spread in the PCA and smaller nucleotide diversity values (π), is suggestive of larger effective population size (N_e) in the deep population (given the relationship $\pi = 4N_e\mu$, assuming equal mutation rates μ and mutation-drift equilibrium). The evident, although small, overlap in the PCA between the large deep population and each one of the smaller shallow-water fjord populations suggests limited gene flow from the deep. This colonization from the deep must have occurred relatively recently, given that present-day North American fjords were predominantly occupied by massive glaciers or exposed to the atmosphere due to sea-level change during the last glacial maximum 33,000-14,500 years ago (Clague & James 2002; Clark & Mix 2002; Clark et al. 2009). However, we suggest enough time may have passed since colonization because allele frequency distributions are not strongly skewed towards fixation, as expected for young nonequilibrium populations founded by few individuals (e.g., Catchen et al. 2013a). The dynamics of population structuring between shallow-water inner fjord and outer deep populations of P. stephencairnsi over time are unknown. However, compelling evidence from other species suggests that these patterns of differentiation between fjord and open water populations can be stable over at least hundreds of years (Harnstrom et al. 2011).

Natural selection in shallow-water fjord environments

The differences in conditions between open water and shallow fjord environments (e.g. Fig. 1b) can also act as barriers for gene flow, further limiting the amount of gene flow between shallow and deep-sea populations. Compared to deep populations, the shallow-water fjord populations are exposed to lower hydrostatic pressure and significantly greater variability ranges of temperature, pH, salinity, current speeds, and sedimentation rates influenced by marked seasonality (observations from the NEPTUNE and VENUS time series). These conditions may not only act as barriers for gene flow but also as selective forces. The observed F_{1S} distributions with a tendency to negative values and the higher percentage of loci of intermediate allelic frequencies in shallow-water populations, relative to the deep population, may be the result of natural selection (e.g. heterozygote advantage or balancing selection). Alternatively, demographic processes, such as bottlenecks, may also produce similar patterns (Luikart *et al.* 1998).

The comparative approach between shallow-water populations and the deep ancestral population of P. stephencairnsi, allowed us to identify, for the first time, potential markers of parallel adaptation to the shallow-water environment in a deep-sea organism. Although only three markers were mapped to known cnidarian functional regions, there is the potential to identify more genomic regions and link them to potential functions and mechanisms as more genomic resources become available (Hohenlohe et al. 2010; Reitzel et al. 2013). Significant environmental changes are occurring due to anthropogenic CO₂ emissions (Hoegh-Guldberg & Bruno 2010). Despite efforts to understand the effect of these changes on marine species, little is known about the adaptive mechanisms that would allow them to survive over ecological and evolutionary time scales. Shallow-water populations of deep-sea coral species have already adapted to deal with the environmental extremes of the surface ocean, thus they could constitute pre-adapted populations that could expand their range to deeper water in the case of significant environmental changes at depth, thereby seeding future deep-sea ecosystems. On the other hand, shallow-water populations of deep-sea corals may live at the tolerance boundary of their species, and thus could face habitat shifts and local extinction in the near future. A better understanding of the adaptive potential of these corals will allow us to assess the possible impacts of climate change on the diverse but vulnerable ecosystems supported by these habitat-forming corals.

Future work

To increase the confidence in our demographic and natural selection inferences we plan to perform RAD sequencing on additional available individuals from deep and shallow populations. Additionally, the potential markers of parallel adaptation to the shallow-water environment will be mapped to the draft genome sequence of an *P. stephencairnsi* individual, which will be generated using high-throughput sequencing and routine algorithms developed to assemble full genomes from short sequence-reads (following Gnerre *et al.* 2011). High-resolution genomic scans generated by RAD-seq provide genotypic data for tens of thousands of SNPs, thus allowing the creation of genome-wide distributions of F_{ST} and other population genetic summary statistics, which allow further identification of candidate potential genomic regions and elements involved in parallel adaptation to the shallow-water environment (see Hohenlohe *et al.* 2010).

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CHAPTER 7

Summary and conclusions

Deep-sea hydrothermal vents and deep-sea coral ecosystems are some of the most conspicuous biological hotspots in the deep-sea. These ecosystems face increasing threats caused by human activities, such as bottom trawling and deep-sea mining. Knowledge of conservation targets is fundamental for the implementation of efficient conservation strategies that help mitigate these threats. Such knowledge must include well-founded taxonomic inventories that allow us to identify species and ecosystems at risk, as well as an understanding of their relatedness, genetic variance, distribution, connectivity patterns, and adaptation potential. Nonetheless, gaining this knowledge in deep-sea ecosystems is difficult due to the extreme challenges of working in this environment, combined with the paucity of genetic resources for deep-sea organisms. Here I provide fundamental high-priority knowledge in taxonomic, evolutionary, and ecological aspects of deep-sea coral and vent species, by harnessing the power of novel genomic tools and overcoming long-standing methodological barriers.

In Chapter 2, I developed bioinformatic tools that help guide the design of studies aiming to characterize eukaryotic genome diversity using restriction-site associated DNA sequencing. With these tools I performed *in silico* genome-wide surveys thought the eukaryotic tree of life. I tested the hypothesis that genome composition, in terms of GC content, and mono-, di- and trinucleotide composition, can be used to predict the number of restriction sites for a given combination of restriction enzyme and genome across the eukaryotic tree of life. In most cases the trinucleotide genome composition model was the best predictor of the expected number of restriction sites in a eukaryotic genome, and the GC content and mononucleotide models the worst. I conclude that the predictability of restriction site frequencies in eukaryotic genomes needs to be treated on a case-specific basis, whereby the phylogenetic position of the taxon of interest and the specific recognition sequence of the selected restriction enzyme are the chief foci among the most determinant factors. The knowledge gained in this chapter, and the bioinformatic tools developed, was applied in all other subsequent chapters.

In Chapter 3, I tested global-scale historical biogeographic hypothesis of vent fauna using barnacles as model. I characterized the global genetic diversity of vent barnacles to infer their time and place of origin, mode of dispersal, and diversification throughout the world's vents. The approach was to target a suite of

multiple loci in samples representing seven out of the eight described genera. I also performed restrictionsite associated DNA sequencing (RAD-seq) on individuals from each species. Phylogenetic inferences indicated that vent barnacles have colonized deep-sea hydrothermal vents at least twice in history. The late Mesozoic/Cenozoic was the time of colonization and radiation of barnacles in vent ecosystems. Further analyses suggested that the western Pacific was the place of origin of the major vent barnacle lineage, followed by circumglobal colonization eastward along the southern hemisphere during the Neogene. The inferred time of origin rejects previous hypotheses of antiquity of vent taxa. The timing and the mode of origin, radiation and dispersal are consistent with the inferences made for other deep-sea taxa, including non-vent species, and are correlated with the occurrence of major geological events and mass extinctions. Thus, I suggest that the geological processes and dispersal mechanisms discussed here can explain current distribution patterns of many other marine taxa and have played an important role shaping deep-sea faunal diversity.

In Chapter 4, I examined genetic diversity patterns in vent barnacles at a regional scale. To test the hypothesis that seamounts behave as isolated island-like systems, where population connectivity is limited and endemicity is promoted, I examined genome-wide RAD-seq data from three hydrothermal vent barnacle species. I compared the genetic diversity and population structuring patterns of barnacle populations from seamount and spreading ridges. Among the study populations I found patterns of population genetic structuring that do not conform to the predictions from the seamount endemicity hypothesis. The patterns of genetic variation among individuals collected from seamount and spreading ridges, separated horizontally by hundreds of kilometers and vertically by hundreds of meters, did not reject the null hypothesis of panmixia within each species. I found that these inferences are largely insensitive to the *de novo* assembly parameters used to identify loci from sequence reads. In conclusion, I suggest that the seamount endemicity hypothesis warrants further testing using high-resolution genetic markers in other vent organisms with differing life history strategies (e.g. brooders) that may limit their dispersal potential, as well as in non-vent organisms, which are not exposed to evolutionary pressures imposed by the dynamic nature of hydrothermal vent systems.

I then moved on to resolve long-standing questions regarding species definitions and relationships in deep-sea corals. In Chapter 5, I demonstrated the empirical utility of RAD-seq by unambiguously resolving phylogenetic relationships among recalcitrant octocoral taxa with divergences greater than 80 million years. I objectively inferred robust species boundaries in the genus *Paragorgia*, which contains some of the most important ecosystem engineers in the deep-sea, by testing alternative taxonomy-guided or unguided species delimitation hypotheses using the Bayes factors delimitation method (BFD*) with

genome-wide SNP data. I presented conclusive evidence rejecting the current morphological species delimitation model for the genus *Paragorgia* and indicating the presence of cryptic species boundaries associated with environmental variables. I argue that the suitability limits of RAD-seq for phylogenetic inferences in divergent taxa cannot be assessed in terms absolute time, but depend on taxon-specific factors such as mutation rate, generation time and effective population size. Classic morphological taxonomy can greatly benefit from integrative approaches that provide objective tests to species delimitation hypothesis.

Finally, in Chapter 6, I explored the adaptation potential of deep-sea coral species to environmental changes by examining a case of adaptation to shallow water from the deep-sea. Few populations of some deep-sea coral species can be found in shallow (< 45 m) high-latitude fjord environments where they experience significantly different environmental conditions than their deep relatives. Therefore, these shallow-water populations are believed to inhabit the extremes of the species' physiological tolerances and likely have developed adaptations that enable them to colonize these shallow-water environments. I aimed to identify potential genomic regions that have enabled the successful adaptation to shallow-water in the deep-sea octocoral species Paragorgia stephencairnsi. To characterize the genome-wide genetic diversity of populations of P. stephencairnsi found in shallow-water populations and compare it to the genetic diversity from deep-water populations, I performed high-resolution genome-wide scans of single nucleotide polymorphisms through RAD-seq. I found patterns of significant population genetic differentiation among the examined populations of P. stephencairnsi, which are consistent with the hypothesis that larvae from outer deep populations seeded shallow-water inner fjord populations. Furthermore, I find candidate positive-selection markers shared between parallel comparisons of shallow and deep populations, and thus identify them as likely candidate makers for genomic regions involved in adaptation to the shallow-water fjord environment.

Overall, the results from this thesis constitute critical baseline data with which to assess potential effects of anthropogenic disturbances on deep-sea ecosystems. The species delimitation frameworks here developed will enable rapid species assignments as deep-sea specimens from newly explored geographical regions become available. This thesis lays groundwork for describing the impacts of natural selection on deep-sea coral species in the face of environmental changes. The software here developed, and the resulting databases constitute a valuable reference resource that will help guide the choice of restriction enzyme for any study using RAD-seq or related methods.

I anticipate that the use of novel genomic tools to study deep-sea organisms will accelerate the pace of knowledge acquisition, and thus greatly enhance our understanding of deep-sea ecosystems, their evolution, and their role in the global ecosystems network. However, the speed of genomic data generation has now outpaced the development of analytical tools, and thus there is a great need of developing novel ways to make full use of the information contained in large genomic datasets. 'Omics' techniques promise a fast and direct route to move from descriptive studies in deep-sea organisms to process-oriented studies, which will allow use to understand the mechanisms that have allowed life to thrive in this extreme environment. Understanding these mechanisms can also lead to the development of a myriad of applications that can directly improve human's lives.