

MIT Open Access Articles

Oxidation of ambient biogenic secondary organic aerosol by hydroxyl radicals: Effects on cloud condensation nuclei activity

The MIT Faculty has made this article openly available. *Please share* how this access benefits you. Your story matters.

Citation: Wong, J. P. S. et al. "Oxidation of Ambient Biogenic Secondary Organic Aerosol by Hydroxyl Radicals: Effects on Cloud Condensation Nuclei Activity." Geophysical Research Letters 38.22 (2011). ©2012. American Geophysical Union

As Published: http://dx.doi.org/10.1029/2011gl049351

Publisher: American Geophysical Union (AGU)

Persistent URL: http://hdl.handle.net/1721.1/75747

Version: Final published version: final published article, as it appeared in a journal, conference proceedings, or other formally published context

Terms of Use: Article is made available in accordance with the publisher's policy and may be subject to US copyright law. Please refer to the publisher's site for terms of use.



Oxidation of ambient biogenic secondary organic aerosol by hydroxyl radicals: Effects on cloud condensation nuclei activity

J. P. S. Wong,¹ A. K. Y. Lee,¹ J. G. Slowik,^{1,2} D. J. Cziczo,^{3,4} W. R. Leaitch,⁵ A. Macdonald,⁵ and J. P. D. Abbatt¹

Received 22 August 2011; revised 14 October 2011; accepted 19 October 2011; published 22 November 2011.

[1] Changes in the hygroscopicity of ambient biogenic secondary organic aerosols (SOA) due to controlled OH oxidation were investigated at a remote forested site at Whistler Mountain, British Columbia during July of 2010. Coupled photo-oxidation and cloud condensation nuclei (CCN) experiments were conducted on: i) ambient particles exposed to high levels of gas-phase OH, and ii) the watersoluble fraction of ambient particles oxidized by aqueousphase OH. An Aerodyne Aerosol Mass Spectrometer (AMS) monitored the changes in the chemical composition and degree of oxidation (O:C ratio) of the organic component of ambient aerosol due to OH oxidation. The CCN activity of size-selected particles was measured to determine the hygroscopicity parameter ($\kappa_{\text{org,CCN}}$) for particles of various degrees of oxygenation. In both cases, the CCN activity of the oxidized material was higher than that of the ambient particles. In general, $\kappa_{\text{org,CCN}}$ of the aerosol increases with its O:C ratio, in agreement with previous laboratory measurements. Citation: Wong, J. P. S., A. K. Y. Lee, J. G. Slowik, D. J. Cziczo, W. R. Leaitch, A. Macdonald, and J. P. D. Abbatt (2011), Oxidation of ambient biogenic secondary organic aerosol by hydroxyl radicals: Effects on cloud condensation nuclei activity, Geophys. Res. Lett., 38, L22805, doi:10.1029/2011GL049351.

1. Introduction

[2] One of the largest uncertainties in global radiative forcing assessment is the cooling effect of atmospheric particulates via their ability to act as CCN, which is dependent on their size and composition. Organic compounds constitute a large fraction of ambient aerosol, as demonstrated at various locations in the Northern Hemisphere [*Zhang et al.*, 2007]. Characterization of organic aerosol hygroscopicity is thus crucial for reducing uncertainties in current climate models.

[3] The earliest studies of organic aerosol CCN properties used model compounds, mainly chosen by their solubility properties [*Cruz and Pandis*, 1997; *Corrigan and Novakov*, 1999; *Pradeep Kumar et al.*, 2003; *Shantz et al.*, 2003]. These studies illustrated that highly water soluble compounds

Copyright 2011 by the American Geophysical Union. 0094-8276/11/2011GL049351

give rise to good CCN – indicating that organic constituents could be of comparable importance to cloud droplet formation as highly soluble, inorganic components. CCN closure studies were also performed [e.g., *VanReken et al.*, 2003; *Chang et al.*, 2007]. Recent lab work has demonstrated that SOA, formed by photo-oxidation processes from specific precursors, is quite hygroscopic [*Huff Hartz et al.*, 2005; *VanReken et al.*, 2005; *King et al.*, 2007; *Prenni et al.*, 2007; *Duplissy et al.*, 2008; *Engelhart et al.*, 2008; *Massoli et al.*, 2010; *Lambe et al.*, 2011b].

[4] Given that a complete chemical characterization of ambient particles is not yet achievable, their CCN properties cannot be calculated from first-principles. As a result, the particle hygroscopicity is now typically described using the κ -Köhler method, where one parameter (κ) describes the average properties (e.g., molecular weight and solubility) of all compounds present [*Petters and Kreidenweis*, 2007]. This eliminates the need for the specific chemical identities of the constituents found in ambient aerosol and allows for a simple representation of CCN activation in global climate models.

[5] In this context, a highly oxygenated organic particle is more likely to be a better CCN due to the increased polarity and solubility of its constituents. Given that oxidative aging leads to more functionalized solute molecules [George and Abbatt, 2010a], it is reasonable to inquire whether such aging increases the hygroscopicity of these particles. For laboratory data, a simple trend of increasing CCN activity and sub-saturated hygroscopicity with degree of SOA oxidation (approximated by O:C atomic ratios) has been reported [Jimenez et al., 2009; Massoli et al., 2010; Duplissy et al., 2008, 2011]. However, since the chemical composition of laboratory SOA is different than that measured in the atmosphere [Hallquist et al., 2009], it is necessary to determine how oxidation processes affect the CCN activation of ambient SOA. In particular, Chang et al. [2010] conducted a CCN closure study using ambient aerosol measurements, where a direct relationship between the organic aerosol's degree of oxygenation and its hygroscopicity parameter determined from CCN activity ($\kappa_{\text{org,CCN}}$) was proposed. An objective of the current study was to investigate the postulate made by Chang et al. [2010] that there is a simple linear relationship between $\kappa_{\rm org,CCN}$ of an organic aerosol and its degree of oxidation.

[6] Here we report the degree to which OH-initiated gasand aqueous-phase oxidation affect the CCN activity of ambient organic aerosol sampled in a biogenically-rich region. This marks the first deployment of a portable flow tube apparatus that can generate controlled levels of gasphase OH, and it demonstrates the relationship between the CCN activity of ambient particles and their O:C ratios due to

¹Department of Chemistry, University of Toronto, Toronto, Ontario, Canada.

²Paul Scherrer Institut, Villigen, Switzerland.

³Pacific Northwest National Laboratory, Richland, Washington, USA.

⁴Department of Earth, Atmospheric, and Planetary Sciences, Massachusetts Institute of Technology, Cambridge, Massachusetts, USA. ⁵Science and Technology Branch, Environment Canada, Toronto,

Ontario, Canada.



Figure 1. Toronto Photo-Oxidation Tube (TPOT) setup for the study of the gas-phase OH oxidation of ambient organic aerosol and their CCN activity.

OH oxidation. The study of ambient particles, in which their O:C ratios were changed by OH exposure on-site, further supports the observations from earlier laboratory experiments [*Jimenez et al.*, 2009; *Massoli et al.*, 2010; *George and Abbatt*, 2010b; *Duplissy et al.*, 2011; *Lambe et al.*, 2011a] of increasing CCN activity with oxidation, and the relationship of $\kappa_{\text{org,CCN}}$ vs. O:C postulated by *Chang et al.* [2010] and *Lambe et al.* [2011b].

2. Experiment

[7] Experiments were conducted at the Raven's Nest site at Whistler, BC, Canada (1300 m ASL, 50°N 122°W) as part of Environment Canada's Whistler Aerosol and Cloud Study (WACS) 2010. The site is situated in a coniferous forest that produces high levels of biogenic organic aerosol [*Schwartz et al.*, 2010] during warm summertime conditions, with minimal contributions from local pollution and no evidence for biomass burning impact during the periods described below that had daytime temperatures ranging from 20 to 27°C.

2.1. Gas-Phase OH Oxidation

[8] Gas-phase OH oxidation of ambient particles was conducted in the Toronto Photo-Oxidation Tube (TPOT) (Figure 1), which is a modified version of a system [*George et al.*, 2007; *Lambe et al.*, 2011b] that produces high OH exposures $(1.0-1.3 \times 10^{12} \text{ molecules cm}^{-3} \text{ s})$ via the photolysis ($\lambda = 254 \text{ nm}$) of gas-phase O₃ ($1.3-2.7 \times 10^{13} \text{ molecules cm}^{-3}$) in the presence of water vapour (relative humidity = 40%) for a residence time of 2.2 minutes in the oxidation region. Modifications from the technique described in previous publications [*George et al.*, 2007; *Lambe et al.*, 2011b] include the use of a wider diameter, stainless steel flow reactor that was electrochemically coated with amorphous

silicon material (SilcoTek Corp.) to reduce gas-phase adsorption and the use of multiple point aerosol injection to accelerate the mixing of particles with the gas flows. The sample (1100 sccm) was introduced through 3 m of electrochemically coated 6 mm o.d. stainless steel tubing (no cyclone) that extended 0.5 m above the roof of the sampling site. The UV lamp was placed in the centre of the flow tube, and was surrounded by a quartz tube cooled with a flow of room air (~20 lpm). The OH concentration was determined by monitoring the decay of methyl-ethyl-ketone by reaction with OH using the Ionicon PTR-MS. The OH exposures were equivalent to 8–10 days (24-h) of ambient OH exposure of 1.5×10^6 molecules cm⁻³.

[9] During normal operation, ambient aerosol was sampled continuously. Aerosol is mixed with a humidified flow (700 sccm). An automated switching valve alternately injects either an ozone-containing flow (200 sccm) or a blank flow of O_2 and N_2 (200 sccm) with a period of 12 min. These flows are allowed to mix and then pass into the reaction tube, where they are irradiated with UV light. If the ambient aerosol was mixed with ozone, OH radicals are produced; otherwise unreacted particles are measured. In addition to reaction with OH, ambient aerosol can potentially react with ozone and/or be volatilized by heat and UV light produced from the lamp. Ozone control experiments were conducted by comparing organic mass concentrations and mass spectra with and without added ozone under dark conditions using a bypass tube. No reaction with ozone was observed. Volatilization control experiments were conducted by comparing ozonefree organic AMS mass spectra and concentrations in light (i.e., UV lamp on) vs. dark conditions. With the lamp on, the temperature increased by approximately 4°C (25°C to 29°C). In the presence of UV light, an 18% decrease in total organic mass was observed, which can be due to both increased temperature and photolysis of aerosol compounds leading to

their volatilization. More details on these control experiments are in the auxiliary material.¹ Ambient particles were subjected to OH and no OH oxidation alternatively for 12 minutes each, for the total time period required to complete a CCN activation curve scan (described below). Three time periods from each of two days of experiments are analyzed here.

2.2. Aqueous-Phase OH Oxidation

[10] A sampling inlet for particle filter samples was situated beside that for the gas-phase OH oxidation studies. The particle samples passed through a cyclone with PM1 cut size (UGR, Model 463) and were collected on 47 mm Teflon filters (2.0 μ m pores) for 24 hour periods, following which water extraction was conducted. The aqueous solutions were oxidized on-site using our custom photoreactor [*Lee et al.*, 2011], with a final H₂O₂ concentration of 70 mM (continuously stirred) and then irradiated by a UV lamp (UVP, 254 nm) inserted into the solution. The solutions were oxidized for 10 minutes, then nebulized (TSI 3076 atomizer) with compressed air for the detection by particle instruments. Solutions without H₂O₂ were also nebulized under dark conditions (i.e., UV lamp off). Four samples are considered for the current analysis.

2.3. Particle Composition and CCN Measurements

[11] The non-refractory particle composition was measured by an Aerodyne Time-of-Flight aerosol mass spectrometer (C-ToF-AMS). For the CCN measurements, particles were first dried using a silica gel diffusion dryer and then sizeselected at a mobility diameter of 75 nm (TPOT output) or 100 nm (aqueous reactor output) by a TSI-3081 DMA before being measured by the CCNC. The CCN active number fraction was determined from the ratio of the number of CCN-active particles measured by one channel of a DMT-200 CCN counter to the total particle number concentration, measured by a TSI-3010 Condensation Particle Counter (CPC).

[12] To obtain a CCN activation curve, multiple supersaturations between 0.07–1.0% were scanned for a period of 24 minutes on each supersaturation for the gas-phase oxidation experiments, and 10 minutes for the aqueous-phase oxidation. CCN activation curves were constructed by calculating the fraction of activated particles, determined from the CCN/CPC ratio at the various supersaturation values. Using the dry mobility diameter (D_i), and the critical supersaturation (*Sc*), defined as the supersaturation in which 50% of the size-selected particles act as CCN, the kappaparameter for the total particles was calculated as described by *Petters and Kreidenweis* [2007],

$$Sc = \frac{D^3 - D_i^3}{D^3 - D_i^3 (1 - \kappa_{total,CCN})} exp\left(\frac{4\sigma_W M_W}{RT\rho_w D}\right)$$
(1)

where D is the droplet wet diameter, M_w , ρ_w and σ_w are the molecular weight, density, and surface tension of water.

[13] The overall κ of the particles ($\kappa_{total,CCN}$) is calculated as the sum of the volume-fraction weighted average κ_{CCN} 's of the individual components. To arrive at the

kappa-parameter for the organic component of the particles, the following equation was used [*Petters and Kreidenweis*, 2007]:

$$\kappa_{\text{total,CCN}} = \varepsilon_{\text{org}} \kappa_{\text{org,CCN}} + \varepsilon_{\text{inorg}} \kappa_{\text{inorg,CCN}} \tag{2}$$

where $\varepsilon_{\rm org}$ is the volume fraction of the organic component determined from the C-ToF AMS, assuming an initial organic density of 1200 kg m⁻³. Only samples with $\varepsilon_{\rm org} \ge$ 67% (AMS mass fraction, fOrg \geq 0.63) are considered for the current analysis, so that the changes in $\kappa_{\text{org,CCN}}$ due to oxidation are statistically significant and not dominated by the inorganic fraction. For example, the estimated uncertainty in $\kappa_{\rm org,CCN}$ is 25% when $\varepsilon_{\rm org} = 0.8$, with smaller values for larger ε_{org} values. For the gas-phase oxidation experiments, the increase in the organic component density due to oxidation was accounted for, using results obtained from George and Abbatt [2010b] where they determined changes in ρ_{org} of laboratory α -pinene SOA as a function of OH exposure. This amounts to a relative change in the $\kappa_{\rm org,CCN}$ of only 0.5% for the current analysis. For the aqueous oxidation experiments, changes in organic density were not corrected for since the OH exposures used were unknown. The aerosol inorganics measured by the AMS were considered in the calculation as ammonium nitrate ($\kappa = 0.72$, density = 1730 kg m^{-3} kg) and ammonium sulfate ($\kappa = 0.59$, density = 1770 kg m^{-3}) [Windholz, 1983; Clegg et al., 1996, 1998; Wexler and *Clegg*, 2002]. During the periods studied, ammonium sulfate accounted for more than 80% of the measured inorganic mass. At times the particle was not fully neutralized but the sensitivity of calculated $\kappa_{\rm org,CCN}$ to particle acidity was shown to be very low. An assumption made in the analysis was that all particles were internally mixed. Molecular O:C ratios were estimated from the fraction of the total organic signal measured at m/z 44 (f44), using the empirical relationship formulated by Aiken et al. [2008]. The AMS measurements were averaged over the period of time required to complete the corresponding CCN activation scans.

3. Results and Discussion

[14] Sample CCN activation curves from both gas- and aqueous-phase OH oxidation experiments are shown in Figure 2. For all experiments, the critical superstaturation (*Sc*) decreased for samples subjected to oxidation, compared to the control samples. The average $\kappa_{\text{org,CCN}}$ values for the TPOT experiments were smaller than those for the aqueous-phase samples as there may be some larger, less soluble organic species in the ambient particles compared to the water-soluble component. The CCN activation curves for particles subjected to aqueous-phase OH oxidation are steeper compared to those from gas-phase OH oxidation for two reasons. First, the aqueous reactor is expected to produce more uniform OH exposures as the solution was constantly stirred. Second, the extraction and dissolution process is likely to homogenize the samples.

[15] For all the experiments considered, the O:C ratios increased after oxidation (see auxiliary material for detailed results on all experiments, including sample aerosol mass spectra). The mass fraction of m/z 44 (predominantly CO_2^+) to the total organic component (f44), an indicator of highly oxidized species, increased while the mass fraction of m/z 43 (predominantly $C_2H_3O^+$) to the total organic component

¹Auxiliary materials are available in the HTML. doi:10.1029/2011GL049351.



Figure 2. Activated particle fraction (CCN/CN) plotted as a function of the water supersaturation (*SS*) for unoxidized and oxidized ambient particles by gas-phase OH oxidation ($d_m = 75$ nm; July 22, 2010, 00:02–02:50) (red circle) and aqueous-phase OH oxidation ($d_m = 100$ nm; July 8, 2010) (blue square).

(f43), an indicator of less oxidized species, decreased slightly [*Ng et al.*, 2011]. For the gas-phase OH oxidation experiments, the changes in hygroscopicity can arise from conversion of insoluble organic species to more soluble species, changes in the properties of dissolved constituents, the loss of volatile components that are less hygroscopic, and the condensation of oxidation products onto the aerosol. While it is possible there was some condensation of oxidation products, an average decrease in the total organic mass of 10% was observed and so changes in hygroscopicity due to this process were likely to be minor. Note that the small increase in the mass fraction of sulphate to total particle mass (fSO_4^2) due to oxidation are not large enough to explain the overall increase in $\kappa_{total,CCN}$ observed.

[16] Figure 3 shows the relationship between O:C to $\kappa_{\rm org,CCN}$. Literature values of other oxidation-CCN measurements, mostly from lab studies, are included for comparison. We note that other studies have examined this relationship through the hygroscopic growth factor (HGF_{95%}) [*Jimenez et al.*, 2009; *Massoli et al.*, 2010; *Duplissy et al.*, 2011]. The $\kappa_{\rm org,CCN}$ and O:C values from a laboratory oxidative study of O₃- α -pinene SOA are included; specifically, the "lamp" and "OH" flow tube conditions from Figure 9a of



Figure 3. Relationship between O:C and $\kappa_{\text{org,CCN}}$ from the current study and from several laboratory studies of model POA and biogenic SOA [*George et al.*, 2009; *George and Abbatt*, 2010b; *Massoli et al.*, 2010; *Lambe et al.*, 2011a], and a field CCN closure study [*Chang et al.*, 2010] for comparison. The *Lambe et al.* [2011a] solid line represents a fit for SOA from a number of precursors. For each pair of points, the samples of the lowest and highest O:C values are shown from each study. Dashed lines are included only to indicate the relationship between these two points. The following were the OH exposures used for the laboratory studies of: BES (3.0×10^{12} molecules cm⁻³ s) [*George et al.*, 2009], O₃- α -pinene SOA (1.3×10^{12} molecules cm⁻³ s) [*George et al.*, 2009], O₃- α -pinene SOA (1.3×10^{12} molecules cm⁻³ s) [*George and Abbatt*, 2010b], OH- α -pinene SOA (1.2×10^{12} molecules cm⁻³ s) [*Massoli et al.*, 2010] and OH- α -pinene SOA (1.7×10^{12} molecules cm⁻³ s) [*Lambe et al.*, 2011a]. The shaded region illustrates the uncertainty of the $\kappa_{\text{org,CCN}}$ and O:C relationships derived from the CCN closure study [*Chang et al.*, 2010] and laboratory generated organic aerosol [*Lambe et al.*, 2011a]. The uncertainties in the calculated values of O:C and $\kappa_{\text{org,CCN}}$ are mainly influenced by the systematic errors associated with each measurement, as discussed in the main text. Also, daily averages are shown for the gas-phase OH oxidation experiments (results from individual experiments included in the auxiliary material).

George and Abbatt [2010b] provided the best comparison of experimental conditions to the no OH and OH oxidation conditions employed in this study. The $\kappa_{\rm org,CCN}$ values from oxidation of bis-2-ethylhexyl sebacate (BES) were also included as a model of primary organic aerosol [George et al., 2009], although we recognize that this is a different starting material from the monoterpene. The corresponding O:C ratios for BES were calculated from the measured f44 values, using the relationship in Figure A1 from *Lambe et al.* [2011b]. Results from other lab CCN studies of α -pinene SOA are included to further evaluate the generality of the relationship of $\kappa_{\text{org,CCN}}$ and O:C [Massoli et al., 2010; Lambe et al., 2011a]. Also included in the plot are two relationships between $\kappa_{\text{org,CCN}}$ and O:C, i.e., from Chang et al. [2010], which was derived from CCN closure at a rural field site under the postulate of a direct linear relationship between the two quantities (solid green line), and from *Lambe et al.* [2011b], which was derived from various laboratory generated organic aerosol (solid blue line).

[17] The uncertainties in the calculated values of O:C and $\kappa_{\rm org,CCN}$ are mainly influenced by the systematic errors associated with each measurement, including the aerosol's chemical composition (AMS) [Bahreini et al., 2009], instrument supersaturation (CCNC), particle size selection (DMA), and aerosol number (CPC). The uncertainties reported for the empirical relationship formulated by Aiken et al. [2008] to calculate O;C ratios from f44 were also considered. For this study, the average uncertainty was estimated to be 15% for $\kappa_{\text{org,CCN}}$ and 10% for O:C ratios. The results obtained from the gas- and aqueous-phase OH oxidation of ambient organic aerosol follow a direct positive relationship between O:C and $\kappa_{\rm org,CCN}$, as observed in previous lab studies. Also, the CCN closure result from Chang et al. [2010] is consistent with the other measurements. Given the range of potential organic aerosol precursors in the field measurements, it is not surprising that there is some scatter in this plot that combines both lab and field data. Also, to match best with the field data, we note that only the data from Massoli et al. [2010] and Lambe et al. [2011a] that arise from the α -pinene precursor are included (as single points). Nevertheless, correspondence between the two quantities appears clear, extending down to low values of O:C. Thus, this study validates the postulate made earlier by Chang et al. [2010] that there is a connection between the two quantities. Indeed, the correspondence to the results from Whistler to those from Chang et al. [2010] may arise because biogenic SOA from monoterpene precursors was present at the CCN closure field site [Slowik et al., 2010]. It is important to note, as mentioned by Jimenez et al. [2009], the $\kappa_{\rm org,CCN}$ and overall composition from a variety of precursors may become increasingly similar with increasing O:C. Also, a similar molecular weight for many of the SOA precursors in Figure 3 will tend to lead to oxidation products of similar molecular weight, and thus similar hygroscopicities for the same O:C values.

4. Summary

[18] This study is the first direct measurement of the relationship between the degree of oxygenation and the CCN hygroscopicity of the organic component of ambient biogenic aerosol ($\kappa_{\text{org,CCN}}$) subject to OH oxidation, using novel methods for exposure to both gas- and aqueous-phase OH.

When combined with lab measurements, the results support a general linear relationship between the $\kappa_{\text{org,CCN}}$ of biogenic organic aerosols and their degree of oxygenation, as postulated by Chang et al. [2010], and shown previously to prevail for hygroscopic growth factors [Jimenez et al., 2009]. The results emphasize that OH is an important aerosol oxidant. Additional field data with onsite OH processing should provide an indication of whether OH oxidation is the main organic aerosol aging process that drives the increases in hygroscopicity. While empirical in nature, this relationship provides a simple method for relating organic aerosol hygroscopicity to its composition, one that could potentially be included in atmospheric models. This trend needs to be further explored with ambient aerosol arising from a variety of sources, especially in urban and biomass burning regions, given that not all lab SOA precursors follow this relationship as closely [Massoli et al., 2010; Lambe et al., 2011a].

[19] Acknowledgments. The authors would like to thank Environment Canada, NSERC and CFCAS-CAFC for funding.

[20] The Editor thanks two anonymous reviewers for their assistance in evaluating this paper.

References

- Aiken, A. C., et al. (2008), O/C and OM/OC ratios of primary, secondary, and ambient organic aerosols with high-resolution time-of-flight aerosol mass spectrometry, *Environ. Sci. Technol.*, 42, 4478–4485, doi:10.1021/ es703009q.
- Bahreini, R., et al. (2009), Organic aerosol formation in urban and industrial plumes near Houston and Dallas, TX, J. Geophys. Res., 114, D00F16, doi:10.1029/2008JD011493.
- Chang, R. Y.-W., et al. (2007), Comparison between measured and predicted CCN concentrations at Egbert, Ontario: Focus on organic aerosol fraction at a semi-rural site, *Atmos. Environ.*, 41, 8172–8182, doi:10.1016/j. atmosenv.2007.06.039.
- Chang, R. Y.-W., et al. (2010), The hygroscopicity parameter (κ) of ambient organic aerosol at a field site subject to biogenic and anthropogenic influences: Relation to degree of aerosol oxidation, *Atmos. Chem. Phys.*, 10, 5047–5064, doi:10.5194/acp-10-5047-2010.
- Clegg, S., S. Milioto, and D. Palmer (1996), Osmotic and activity coefficients of aqueous (NH₄)₂SO₄ as a function of temperature, and aqueous (NH₄)₂SO₄-H₂SO₄ mixtures at 298.15 K and 323.15 K, *J. Chem. Eng. Data*, 41, 455–467, doi:10.1021/je950289d.
- Clegg, S., P. Brimblecombe, and A. Wexler (1998), A thermodynamic model for the system H⁺-NH₄⁺-SO₄²-NO₃⁻Cl⁻H₂O at 298.15 K, *J. Phys. Chem. A*, *102*, 19,395–19,409.
- Corrigan, C. E., and T. Novakov (1999), Cloud condensation nucleus activity of organic compounds: A laboratory study, *Atmos. Environ.*, 33, 2661–2668, doi:10.1016/S1352-2310(98)00310-0.
- Cruz, C. N., and S. N. Pandis (1997), A study of the ability of pure secondary organic aerosol to act as cloud condensation nuclei, *Atmos. Environ.*, 31, 2205–2214, doi:10.1016/S1352-2310(97)00054-X.
- Duplissy, J., et al. (2008), Cloud forming potential of secondary organic aerosol under near atmospheric conditions, *Geophys. Res. Lett.*, 35, L03818, doi:10.1029/2007GL031075.
- Duplissy, J., et al. (2011), Relating hygroscopicity and composition of organic aerosol particulate matter, *Atmos. Chem. Phys.*, 11, 1155–1165, doi:10.5194/acp-11-1155-2011.
- Engelhart, G. J., et al. (2008), CCN activity and droplet growth kinetics of fresh and aged monoterpene secondary organic aerosol, *Atmos. Chem. Phys.*, 8, 3937–3949, doi:10.5194/acp-8-3937-2008.
- George, I. J., and J. P. D. Abbatt (2010a), Heterogeneous oxidation of atmospheric aerosol particles by gas-phase radicals, *Nat. Chem.*, 2, 713–722, doi:10.1038/nchem.806.
- George, I. J., and J. P. D. Abbatt (2010b), Chemical evolution of secondary organic aerosol from OH-initiated heterogeneous oxidation, *Atmos. Chem. Phys.*, 10, 5551–5563, doi:10.5194/acp-10-5551-2010.
- George, I. J., et al. (2007), Heterogeneous oxidation of saturated organic aerosols by hydroxyl radicals: Uptake kinetics, condensed phase products, and particle size change, *Atmos. Chem. Phys.*, 7, 4187–4201, doi:10.5194/acp-7-4187-2007.

- George, I. J., et al. (2009), Modification of cloud condensation nucleus activity of organic aerosols by hydroxyl radical heterogeneous oxidation, *Atmos. Environ.*, 43, 5038–5045, doi:10.1016/j.atmosenv.2009.06.043.
- Hallquist, M., et al. (2009), The formation, properties and impact of secondary organic aerosols: Current and emerging issues, *Atmos. Chem. Phys.*, 9, 5155–5236, doi:10.5194/acp-9-5155-2009.
- Huff Hartz, K. E., T. Rosenørn, S. R. Ferchak, T. M. Raymond, M. Bilde, N. M. Donahue, and S. N. Pandis (2005), Cloud condensation nuclei activation of monoterpene and sesquiterpene secondary organic aerosol, J. Geophys. Res., 110, D14208, doi:10.1029/2004JD005754.
- Jimenez, J. L., et al. (2009), Evolution of organic aerosols in the atmosphere, *Science*, 326, 1525–1529, doi:10.1126/science.1180353.
- King, S. M., et al. (2007), Cloud condensation nucleus activity of secondary organic aerosol particles mixed with sulfate, *Geophys. Res. Lett.*, 34, L24806, doi:10.1029/2007GL030390.
- Lambe, A. T., et al. (2011a), Laboratory studies of the chemical composition and cloud condensation nuclei (CCN) activity of secondary organic aerosol (SOA) and oxidized primary organic aerosol (OPOA), *Atmos. Chem. Phys.*, 11, 8913–8928, doi:10.5194/acp-11-8913-2011.
- Lambe, A. T., et al. (2011b), Characterization of aerosol photooxidation flow reactors: heterogeneous oxidation, secondary organic aerosol formation and cloud condensation nuclei activity measurements, *Atmos. Meas. Tech.*, 4, 445–461, doi:10.5194/amt-4-445-2011.
- Lee, A. K. Y., P. Herckes, W. R. Leaitch, A. M. Macdonald, and J. P. D. Abbatt (2011), Aqueous OH oxidation of ambient organic aerosol and cloud water organics: Formation of highly oxidized products, *Geophys. Res. Lett.*, 38, L11805, doi:10.1029/2011GL047439.
- Massoli, P., et al. (2010), Relationship between aerosol oxidation level and hygroscopic properties of laboratory generated secondary organic aerosol (SOA) particles, *Geophys. Res. Lett.*, 37, L24801, doi:10.1029/ 2010GL045258.
- Ng, N. L., et al. (2011), Changes in organic aerosol composition with aging inferred from aerosol mass spectra, *Atmos. Chem. Phys.*, 11, 6465–6474, doi:10.5194/acp-11-6465-2011.
- Petters, M. D., and S. M. Kreidenweis (2007), A single parameter representation of hygroscopic growth and cloud condensation nucleus activity, *Atmos. Chem. Phys.*, 7, 1961–1971, doi:10.5194/acp-7-1961-2007.
- Pradeep Kumar, P., K. Broekhuizen, and J. P. D. Abbatt (2003), Organic acid as cloud condensation nuclei: Laboratory studies of highly soluble

and insoluble species, Atmos. Chem. Phys., 3, 509-520, doi:10.5194/acp-3-509-2003.

- Prenni, A. J., M. D. Petters, S. M. Kreidenweis, P. J. DeMott, and P. J. Ziemann (2007), Cloud droplet activation of secondary organic aerosol, *J. Geophys. Res.*, 112, D10223, doi:10.1029/2006JD007963.
- Schwartz, R. E., et al. (2010), Biogenic oxidized organic functional groups in aerosol particles from a mountain forest site and their similarities to laboratory chamber products, *Atmos. Chem. Phys.*, 10, 5075–5088, doi:10.5194/acp-10-5075-2010.
- Shantz, N. C., W. R. Leaitch, and P. F. Caffrey (2003), Effects of organics of low solubility on the growth rate of cloud droplets, *J. Geophys. Res.*, 108(D5), 4168, doi:10.1029/2002JD002540.
- Slowik, J. G., et al. (2010), Characterization of a large biogenic secondary organic aerosol event from eastern Canadian forests, *Atmos. Chem. Phys.*, 10, 2825–2845, doi:10.5194/acp-10-2825-2010.
- VanReken, T. M., T. A. Rissman, G. C. Roberts, V. Varutbangkul, H. H. Jonsson, R. C. Flagan, and J. H. Seinfeld (2003), Toward aerosol/cloud condensation nuclei (CCN) closure during CRYSAL-FACE, *J. Geophys. Res.*, 108(D20), 4633, doi:10.1029/2003JD003582.
- VanReken, T. M., N. L. Ng, R. C. Flagan, and J. H. Seinfeld (2005), Cloud condensation nucleus activation properties of biogenic secondary organic aerosol, J. Geophys. Res., 110, D07206, doi:10.1029/2004JD005465.
- Wexler, A. S., and S. L. Clegg (2002), Atmospheric aerosol models for systems including the ions H^+ , NH_4^+ , Na^+ , SO_4^{2-} , NO_3^- , Cl^- , Br^- and H_2O , J. Geophys. Res., 107(D14), 4207, doi:10.1029/2001JD000451.
- Windholz, M. (Ed.) (1983), *The Merck Index*, 10th ed., Merck, Rahway, N. J.
- Zhang, Q., et al. (2007), Ubiquity and dominance of oxygenated species in organic aerosols in anthropogenically-influenced Northern Hemisphere midlatitudes, *Geophys. Res. Lett.*, 34, L13801, doi:10.1029/2007GL029979.
- J. P. D. Abbatt, A. K. Y. Lee, and J. P. S. Wong, Department of Chemistry, University of Toronto, 80 St. George St., Toronto, ON M5S 3H6, Canada. (jpswong@chem.utoronto.ca)
- D. J. Cziczo, Department of Earth, Atmospheric, and Planetary Sciences, Massachusetts Institute of Technology, Cambridge, MA 02139, USA. W. R. Leaitch and A. Macdonald, Science and Technology Branch,
- W. R. Leattch and A. Macdonald, Science and Technology Branch, Environment Canada, 4905 Dufferin St., Toronto, ON M3H 5T4, Canada. J. G. Slowik, Paul Scherrer Institut, CH-5232 Villigen PSI, Switzerland.