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# Plant Nanobionic Sensors for Arsenic Detection

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- 1 Plant Nanobionic Sensors for Arsenic Detection
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- 6 ABSTRACT

7 Arsenic is a highly toxic heavy metal pollutant which poses a significant health risk to humans and other ecosystems. In this work, the natural ability of wild-type plants to pre-8 9 concentrate and extract arsenic from the belowground environment is exploited to engineer plant 10 nanobionic sensors for real-time arsenic detection. Near-infrared (NIR) fluorescent nanosensors were specifically designed for sensitive and selective detection of arsenite. These optical 11 12 nanosensors were embedded in plant tissues to non-destructively access and monitor the internal 13 dynamics of arsenic taken up by the plants via the roots. The integration of optical nanosensors with 14 living plants enabled the conversion of plants into self-powered autosamplers of arsenic from their environment. Arsenite detection was demonstrated with three different plant species as nanobionic 15 sensors. Based on an experimentally-validated kinetic model, the nanobionic sensor could detect 0.6 16 ppb and 0.2 ppb levels of arsenic after 7 and 14 days respectively by exploiting the natural ability of 17 18 Pteris cretica ferns to hyperaccumulate and tolerate exceptionally high level of arsenic. The sensor 19 readout could also be interfaced with portable electronics at a standoff distance, potentially 20 enabling applications in environmental monitoring and agronomic research.

- 21 **Keywords:** plant nanobionic, optical sensor, carbon nanotube, arsenic, nanoparticles, molecular
- 22 recognition

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#### 24 MAIN TEXT

25 The abundance of arsenic compounds in the environment poses a serious threat to human health and ecosystems.<sup>[1,2]</sup> Long-term exposure to arsenic in humans is associated with cardiovascular 26 diseases, birth defects, severe skin lesions and various types of cancer.<sup>[3,4]</sup> Anthropogenic activities 27 28 such as mining, smelting, irrigation with arsenic-contaminated water and the extensive use of 29 arsenic-based pesticides in the past decades have led to significant arsenic accumulation in underground water and agricultural soils.<sup>[5–7]</sup> Elevated levels of arsenic in the soils not only inhibit 30 plant growth and result in substantial losses in crop production, but also lead to higher arsenic 31 uptake by crops and contamination of the food chain.<sup>[7–9]</sup> These concerns over arsenic exposure 32 prompted the World Health Organization (WHO) and Food and Agriculture Organization (FAO) of the 33 United Nations to set the maximum contaminant level of arsenic in drinking and irrigation water to 34 10 ppb and 100 ppb respectively.<sup>[10]</sup> 35

Arsenic exists primarily as arsenite  $(As^{3+})$  and arsenate  $(As^{5+})$  in aqueous environment.<sup>[11]</sup> In 36 anaerobic conditions such as paddy soils, arsenite is the predominant chemical form of arsenic and it 37 can be efficiently taken up by plants via different mechanisms.<sup>[12,13]</sup> However, there is a lack of 38 reliable techniques capable of rapidly assessing the uptake of arsenic in plants or the arsenic content 39 40 within agricultural soil. The conventional method to determine the arsenic level in plants and soil is based on regular field sampling, plant tissue digestion, extraction and analysis using mass 41 spectrometry.<sup>[14–17]</sup> Such sampling procedure requires extensive sample pre-treatment, bulky and 42 expensive instrumentation, and does not allow for real-time monitoring of arsenic contamination in 43 the field.<sup>[18]</sup> Reflectance spectroscopy and hyperspectral imaging have been proposed as alternatives 44 to monitor arsenic level in plants at a remote distance.<sup>[19-21]</sup> However, these methods are non-45 46 specific towards arsenic contamination, and they rely on slow phenotypic changes of stressed plants 47 such as significant reduction in chlorophyll concentration, destruction of leaf cellular structure and appearance of chlorotic symptoms.<sup>[19,22]</sup> Electrochemical and optical arsenic detection using 48 nanoparticles have been demonstrated *in-vitro* and in contaminated water samples, <sup>[23–27]</sup> but their 49 50 application to monitor the arsenic uptake within plants in real time remains unexplored.

51 In this work, we demonstrate the use of living plants, interfaced with specifically designed 52 nanomaterials, to serve as self-powered and naturally occurring detectors of arsenic present in 53 belowground environment. This plant nanobionic approach enables real-time monitoring of arsenite taken up by the roots of wild-type plants at a standoff distance. A pair of SWNT-based NIR 54 55 fluorescent nanosensors was rationally designed to selectively recognize arsenite via modulation of 56 their emission intensity. These nanosensors were embedded in the leaf mesophyll of living plants, 57 enabling the detection of arsenite molecules as they are taken up by the roots, transported along 58 the plant vasculature and pre-concentrated in the leaf lamina. The integration of our nanosensors 59 with Cretan brake fern (Pteris cretica), a fern species capable of hyperaccumulating high levels of arsenic in their tissues, enabled the standoff detection of arsenite at the low ppb level, well below 60 61 the regulatory limit of arsenic in drinking and irrigation water. By harnessing the unique optical 62 properties of nanomaterials and the natural properties of plants to pre-concentrate and hyperaccumulate arsenic, we show the engineering of living plants as autonomous microfluidic 63

- 64 samplers capable of real-time, non-destructive and ultrasensitive detection of arsenite in the
- 65 environment.
- 66

## 67 Nanosensor development and characterization

Herein, we use the corona phase molecular recognition (CoPhMoRe) technique, which we 68 have previously introduced,<sup>[28]</sup> to develop SWNT-based optical nanosensors for selective detection of 69 arsenite. In this technique, an adsorbed heteropolymer phase on the SWNT surface, called the 70 71 corona, provides synthetic molecular recognition sites that can bind or interact with the target 72 analyte. Such interaction translates into modulations in the NIR fluorescence spectrum of SWNT, and enables the detection of a variety of target analytes including small signaling molecules and 73 nitroaromatic compounds in living plants.<sup>[29–31]</sup> SWNT offers unique advantages for long-term 74 75 sensing applications in planta because they fluoresce in the near-infrared region away from the chlorophyll autofluorescence and do not photobleach.<sup>[32,33]</sup> In addition, their surface properties can 76 be engineered to target different plant organs or subcellular organelles.<sup>[34–36]</sup> 77

78 Arsenite is chosen as the target analyte because it is the predominant form of arsenic in 79 anaerobic paddy soils which can be taken up efficiently by crops through silicon transporters in the roots.<sup>[12,37,38]</sup> Previous studies have shown that guanine (G) and thymine (T) nucleotides can form 80 strong hydrogen bonds with the hydroxy (-OH) groups of arsenite.<sup>[39,40]</sup> To exploit the ability of 81 certain DNA bases to interact with arsenite, we first constructed a library of single stranded DNA 82 83 (ssDNA)-wrapped SWNT with oligonucleotides of varying lengths and G-/T- compositions. The optical 84 sensor responses of DNA-wrapped SWNT constructs were recorded following a 15-minute 85 incubation of 100 µM arsenite in 0.1 M NaCl solution buffer. SWNT wrapped with oligonucleotide 86 sequences containing high G-/T- content, such as  $(GT)_N$ -SWNT where N = 5 to 30, exhibit a significant increase in fluorescence intensity  $((I-I_0)/I_0)$  by as much as 650% for the (9,4) SWNT chirality upon the 87 88 addition of arsenite (Figure 1a). Substitution of G-/T- nucleotides with adenine (A) or cytosine (C), such as (GTAA)<sub>7</sub>-SWNT and (GTCC)<sub>7</sub>-SWNT, diminishes the DNA-SWNT sensor response towards 89 90 arsenite. Oligonucleotide sequences which do not contain G-/T- bases, such as  $C_{30}$ -SWNT and (AC)<sub>15</sub>-91 SWNT, remain largely nonresponsive when exposed to arsenite (Figure 1a). The decrease in sensor 92 sensitivity as the G-/T- composition in the corona phase is reduced confirms previous findings which show that G-/T- nucleotides are potential binding sites with arsenite.<sup>[39,40]</sup> In addition, we found that 93 the oligonucleotide length of the (GT)<sub>N</sub>-SWNT construct significantly affected the fluorescence 94 95 intensity modulation from arsenite. Shorter (GT)<sub>N</sub> polymers yield higher intensity change in response 96 to arsenite compared to longer (GT)<sub>N</sub> sequences, with (GT)<sub>5</sub>-SWNT and (GT)<sub>25</sub>-SWNT showing a 97 maximum and minimum response of 650% and 210% respectively. (GT)<sub>N</sub> sequences for N < 5 were not investigated due to the apparent instability of these DNA-SWNT constructs which would hinder 98 their in planta sensing applications.<sup>[41]</sup> Since G and T nucleobases also strongly pi-stack onto the 99 SWNT surface,<sup>[42]</sup> the length dependence suggests a trade-off between arsenite and SWNT surface 100 binding that is optimized at N = 5. 101

102 Nanosensor responses towards arsenate were also evaluated to investigate the selectivity of 103 DNA-SWNT constructs in distinguishing different arsenic species. In comparison to the sensitivity of 104 the nanosensors towards arsenite, the modulation of DNA-SWNT fluorescence intensity was smaller 105 upon the addition of 100 µM arsenate for all DNA-SWNT hybrids tested in this study (Figure 1b). 106 Among (GT),-SWNT, (GT)<sub>5</sub>-SWNT shows the minimal turn-on response of 15% while (GT)<sub>30</sub>-SWNT shows a 110% response towards arsenate. The weaker response of (GT)<sub>N</sub>-SWNT elicited by arsenate 107 108 compared to arsenite may be attributed to the presence of ketone group in arsenate, which can disrupt the formation of hydrogen bonds with the amine groups of G-/T- nucleotides.<sup>[43]</sup> As (GT)<sub>5</sub>-109 SWNT shows the highest sensitivity and selectivity towards arsenite, we further investigated if such 110 111 sensor performance is unique to (GT)<sub>5</sub> sequence or if other short 10-mer sequences also exhibit similar sensitivity. We found that substitution of G-/T- bases with C nucleotides, which have a high 112 affinity to bind onto SWNT sidewall,<sup>[44,45]</sup> decreased the sensor sensitivity towards arsenite (Figure 113 **S1, Supporting Information**). This is similar to the trend observed among longer oligonucleotide 114 115 sequences and suggests that the specific oligonucleotide chemistry is responsible for the sensor sensitivity and selectivity (Figure 1a). 116

117 The large intensity increase of (GT)<sub>5</sub>-SWNT observed upon arsenite interaction can be attributed to the low baseline fluorescence of SWNT chiralities with larger diameter (Figure 1c). The 118 excitation-emission map of (GT)<sub>5</sub>-SWNT also confirmed that the largest intensity modulations 119 induced by arsenite were exhibited by large-diameter SWNT chiralities (Figure S2, Supporting 120 121 Information). Recent studies suggested that short (GT)<sub>6</sub> polymers form highly-ordered ring 122 structures on the SWNT surface, creating a periodic charge distribution which effectively provides an effective doping of SWNT.<sup>[41]</sup> This doping effect suppresses the radiative exciton relaxation and 123 activates nonradiative exciton relaxation mechanisms, giving rise to the quenched baseline 124 125 fluorescence of short (GT)<sub>N</sub>-SWNT constructs. The corona structure of adsorbed DNA on the SWNT surface is influenced by the solution microenvironment such as ionic strength and pH.<sup>[46]</sup> Thus, we 126 further tested the response of (GT)<sub>N</sub>-SWNT nanosensors in MES and TES buffers which are commonly 127 used for plant infiltration.<sup>[31,47]</sup> (GT)<sub>N</sub>-SWNT constructs maintain their turn-on response upon arsenite 128 exposure, with (GT)<sub>5</sub>-SWNT exhibiting the highest sensitivity of 218% and 195% in MES and TES 129 130 buffer respectively (Figure S3, Supporting Information). The responses of (GT)<sub>5</sub>-SWNT against different concentrations of arsenite could be fitted to a kinetic adsorption model to yield a sensor 131 132 dissociation constant (K<sub>d</sub>) of 26  $\mu$ M (Figure 1d).<sup>[48]</sup> The limit of detection of (GT)<sub>5</sub>-SWNT, calculated from the arsenite concentration which resulted in a signal-to-background ratio  $\geq$  3, was estimated to 133 be 122 nM. In addition, the (GT)<sub>5</sub>-SWNT sensor response towards arsenite can be reversed with the 134 135 introduction of ethylenediaminetetraacetic acid (EDTA), a common metal chelating agent (Figure S4, Supporting Information). (GT)<sub>5</sub>-SWNT complexes were also selective towards arsenite over other 136 heavy metal ions which may be present as contaminants in the soil (Figure 1e). Taken together, the 137 138 high sensitivity, selectivity and compatibility in biologically relevant environment motivates the application of  $(GT)_5$ -SWNT to probe arsenite level *in planta*. 139

## 140 Integration of optical nanosenosrs with living plants

141 Spinach plants (Spinacia oleracea) were turned into an autonomous detector of arsenite by 142 interfacing with a SWNT-based ratiometric sensor platform consisting of a reference and an active 143 sensor. In this platform,  $(GT)_5$ -SWNT served as the active sensor which would exhibit a turn-on 144 response upon arsenite detection, while C10-SWNT was selected as the reference sensor which 145 would remain invariant upon arsenite exposure. The DNA-SWNT constructs were infiltrated into two different regions of a leaf lamina of spinach plants, separated by the midrib, via syringe infiltration at 146 the adaxial side (Figure 2a). Arsenite solution was then introduced to the root environment and as 147 148 transpiration occurs, arsenite would be taken up by the roots and transported to the leaf via the 149 plant vasculature where they would eventually accumulate and come into contact with the 150 embedded manosensors. The NIR fluorescence of both sensor complexes were monitored at a standoff distance of 1 m with a two-dimensional (2D) array InGaAs detector. (GT)<sub>5</sub>-SWNT 151 152 fluorescence intensity started to increase approximately 30 minutes after the introduction of 10 µM 153 arsenite solution to the roots of spinach plants (Figure 2b). After 5 hours, an 11% increase in the 154 average (GT), SWNT fluorescence intensity was observed – this intensity modulation corresponds to approximately 0.3 µM change in leaf arsenite concentration (Figure 2c). In contrast, the fluorescence 155 156 intensity of the control sensor, C<sub>10</sub>-SWNT, remained relatively invariant throughout the experiment 157 as expected. The relative intensity of (GT)<sub>5</sub>-SWNT to C<sub>10</sub>-SWNT (I<sub>G/C-SWNT</sub>) was defined as the readout of the ratiometric sensor approach. When water was introduced to the roots of spinach plants as a 158 159 control, Ig/cswnt remained relatively constant over 5 hours (Figure 2d). The difference in the 160 responses of the nanosensors confirmed that our ratiometric platform enabled the selective detection of arsenite as they were taken up by the roots and transported to the leaf lamina. The 161 embedded nanosensors could tap into plants' internal state and allow the interfacing of such 162 information to electronics, enabling plants to serve as nanobionic devices which can communicate 163 164 the information they receive from the environment to detectors easily interpreted by human.

The nanobionic approach can also be extended to other plant species to convert any wild-165 166 type plants into arsenic detectors. The nanosensor platform was applied to monitor arsenic uptake in rice plants (Oryza sativa). As a staple food for half of the global human population, rice is a major 167 dietary source of arsenic.<sup>[49,50]</sup> Previous reports have shown that rice accumulates arsenite more 168 169 efficiently than other cereal crops such as barley (Hordeum vulgare) or wheat (Triticum aestivum), elevating the concerns of arsenic contamination of the human food chain.<sup>[51,52]</sup> The introduction of 170 10 µM arsenite to the roots of 6-week old rice plants resulted in an average of 15% increase in the 171 172 I<sub>G/C-SWNT</sub> profile after 5 hours (Figure 2e, f). The I<sub>G/C-SWNT</sub> level remained relatively constant in the absence of arsenite. The variance in the sensor response dynamics between rice and spinach plants 173 174 may be due to differences in biological factors such as the vascular structure between 175 monocotyledonous (e.g. rice) and dicotyledonous plants (e.g. spinach), distribution of arsenic uptake 176 channels in the roots, as well as the leaf surface area which affects the transpiration rate. 177 Nonetheless, these results suggest that the nanosensors can be applied to probe the arsenite uptake 178 in both monocotyle donous and dicotyle donous plant species, such as rice and spinach respectively. 179 This provides a unique practical advantage in contrast to genetic engineering methods to produce biosensors for analyte detection in planta, which are only feasible in a limited number of plant 180 species.<sup>[33]</sup> 181

#### 182 Arsenic detection with nanosensors in the visible range

We further demonstrated the versatility of our nanosensor probe for imaging in both the NIR range 183 184 as well as the visible spectra. To enable imaging of the probe in the visible region, we prepared selfassembled nanostructures comprising of SWNT, single-stranded (GT)<sub>5</sub> sequence and TO-PRO-1 (TP), 185 a cyanine dye that intercalates with DNA. Unlike common fluorescent dyes which are typically 186 quenched in the proximity of SWNT, TP switches from a non-fluorescent state to a highly fluorescent 187 state when constrained in a conformationally restrictive environment.<sup>[53]</sup> Upon tip-sonication, the 188 three components readily self-assemble to form TP-(GT)<sub>5</sub>-SWNT nanoconstructs which are 189 190 fluorescent in both the NIR range, enabled by the SWNT backbone, and the visible range, enabled by the TP dye. Successful incorporation of TP into the (GT)<sub>5</sub>-SWNT construct was confirmed the 191 192 appearance of a distinct absorption peak at 515 nm in TP-(GT)<sub>5</sub>-SWNT absorbance spectrum, which 193 corresponds to the absorption maximum of TP dye (Figure 3a). TP-(GT)<sub>5</sub>-SWNT still maintain well-194 defined NIR fluorescence profile and, more importantly, the nanoconstructs show similar intensity modulation towards arsenite with and without TP intercalation (Figure 3b). Additionally, the 195 196 fluorescence of TP-(GT)<sub>5</sub>-SWNT in the visible range decreases in response to arsenite with a 197 comparable sensitivity range as that of (GT)<sub>5</sub>-SWNT in the NIR range (Figure 3c). To determine if this 198 intensity modulation in the visible region is caused by the specific interaction between arsenite and 199  $(GT)_5$  sequences, we also prepared TP-C<sub>10</sub>-SWNT and monitored its response towards arsenite. The fluorescence of TP-C<sub>10</sub>-SWNT in the visible range remained unaffected upon the introduction of 200 201 arsenite at different concentrations (Figure 3c). These findings suggest that the interaction between 202 the (GT)<sub>5</sub> wrapping and arsenite may induce a conformational change in the SWNT corona phase and 203 the bound TP molecules, leading to intensity modulation of the dye-labelled nanoconstructs.

204 The nanosensors' visible fluorescence enables the visualization of nanosensor dynamics 205 within plant cells at a subcellular resolution with visible confocal microscopy. TP-(GT)<sub>5</sub>-SWNT 206 complexes are localized along the cell membrane in the spinach mesophyll layer after syringe 207 infiltration to the adaxial side of a spinach leaf (Figure 3d). As shown in Figure 3d, the addition of 10 208 μM arsenite decreases the visible fluorescence intensity of the nanocomplexes with different 209 magnitude at various subcellular locations. Three randomly-selected locations of TP-(GT)<sub>5</sub>-SWNT 210 showed quenching magnitudes between 30 and 60% (Figure 3e). The heterogeneous sensor 211 dynamics between different subcellular locations may be caused by the spatial profile of arsenite 212 transport within plant cells, or the distribution of nanosensors within the leaf mesophyll. This 213 demonstration highlights the facile modification that can be employed to engineer versatile SWNT-214 based probes, allowing application in the NIR range for whole plant imaging as well as in the visible region for subcellular arsenite detection as shown in this work. 215

#### 216 Nanobionic sensors based on arsenic hyperaccumulators

Plants exhibit natural diversity in their adaptive responses to thrive in arsenic-containing
soils. Some species of plants, primarily ferns from the *Pteris* genus, have naturally evolved the
exceptional capability to accumulate and tolerate a high concentration of arsenic in their
aboveground biomass.<sup>[54–56]</sup> For example, the Chinese brake fern *Pteris vittata*, the first known
arsenic hyperaccumulating fern, can concentrate as much as 5,131 ppm arsenic in the fronds when

grown in soil containing 50 ppm arsenic over 2 weeks.<sup>[57]</sup> In this work, we harnessed the 222 223 hyperaccumulating capability of the Cretan brake fern P. cretica, a previously identified arsenic hyperaccumulator [54,58] to enhance the sensitivity of plant nanobionic sensors for arsenic detection. 224 Both (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT were applied on separate sides of the costa, the midrib of the fern 225 leaflet (Figure 4a). Upon exposure to 10 µM arsenite solution at the roots, the NIR fluorescence 226 intensity of (GT)<sub>5</sub>-SWNT showed a steady increase over 7 days relative to the initial value, while that 227 of  $C_{10}$ -SWNT remained relatively invariant (Figure 4b). The ratiometric sensor response ( $\Delta I/I_0$ ) of *P*. 228 cretica plants is consistently higher throughout the 7-day period than that of spinach or rice plants, 229 230 showing a significant increase of 74% relative to the initial level. (Figure 4c). There was no significant 231 difference in the chlorophyll concentration between plants infiltrated with MES buffer and plants 232 treated with DNA-SWNT and 10 μM arsenite over the 7-day period (Figure S5, Supporting 233 Information). Inductively coupled plasma mass spectroscopy (ICP-MS) analysis was also performed 234 on the treated plant samples to construct a calibration curve which allows the translation of sensor 235 intensity modulation to actual changes in the frond arsenic concentration (Figure S6, Supporting 236 Information). The nanosensor intensity modulation can also be captured by a portable Raspberry Pi 237 platform equipped with a charge-coupled device (CCD) camera, a similar technology to a commercial 238 smartphone-based camera (Figure S7a, Supporting Information). Analysis of images collected 239 through the CCD camera showed a similar ratiometric sensor response as that captured with the 240 InGaAs detector with a 64% increase over 7 days (Figure S7b, Supporting Information), 241 demonstrating the feasibility of interfacing the plant nanobionic sensor with inexpensive, portable 242 electronic devices. The plant nanobionic sensor was then treated with lower concentrations of arsenite in the root uptake solution down to 0.1  $\mu$ M. As expected, the arsenic concentration in P. 243 244 cretica fronds, obtained from monitoring the nanosensor fluorescence intensity, decreased with 245 lower arsenite concentrations in the root uptake solution across the 7-day experiment duration 246 (Figure 4d). We note that there may be slight differences in the sensor response dynamics if arsenite 247 is introduced into the soil instead of the root uptake solution, due to factors such as soil porosity, tortuosity, and gravimetric water content. Nonetheless, our work demonstrates that plants can be 248 engineered as living environmental sensors for sensitive arsenite detection from the belowground 249 250 environment.

251 Kinetic model to estimate plant nanobionic sensor detection limit

We described the uptake of arsenite in *P. cretica* with a kinetic model to obtain a theoretical limit of detection of the plant nanobionic sensor. For arsenite molecules to be detected by the nanosensors embedded in the frond, they have to be taken up by transporters in the roots and translocated to the frond via the xylem before coming into contact with the nanosensors. At the frond, arsenite may be sequestered into the vacuole for long-term storage and detoxification.<sup>[59,60]</sup> The exchange of arsenite between these different compartments can be summarized as a series of reactions:

259

$$As_{sol}^{3+} \rightleftharpoons Is_{root}^{3+} \to As_{frond}^{3+} \to As_{seq}^{3+}$$
(1)

where  $As_{sol}^{3+}$ ,  $As_{root}^{3+}$ ,  $As_{frond}^{3+}$  and  $As_{sed}^{3+}$  denote the arsenite species present in the uptake 260 solution, roots, frond and sequestration compartment respectively. The nanosensor fluorescence 261 intensity indicates the level of  $As_{frond}^{3+}$ . In *P. vittata*, another arsenic hyperaccumulating species in 262 the Pteridaceae family, the transporter-mediated influx of As<sup>3+</sup> from the uptake solution into the 263 roots ( $As_{sol}^{3}$ ) has been shown to follow Michaelis-Menten kinetics, with a maximum net 264 influx rate of 8-10 nmol.As g<sup>-1</sup> root fresh weight (FW) h<sup>-1</sup> at saturating conditions.<sup>[61]</sup> The 265 translocation of As<sup>3+</sup> from the roots to the fronds ( $As_{root}^{3+} \rightarrow As_{frond}^{3+}$ ) in *P. vittata* is mainly driven by 266 transpiration, with a mean transpiration rate of 5-7 g.H<sub>2</sub>O  $g^{-1}$  frond FW  $d^{-1}$  under normal 267 conditions.<sup>[62]</sup> Accounting for the plant biomass used in this study (approximately 23 g root FW and 268 15 g frond FW) and the average arsenite concentration in the xylem sap of hyperaccumulators, <sup>[63]</sup> we 269 estimated the arsenite root-to-frond translocation rate and the maximum root influx rate to be 29-270 38 µmol.As d<sup>-1</sup> and 4.3-5.4 µmol.As d<sup>-1</sup> respectively. We further defined a modified Damköhler 271 number (Da) as the ratio between the compartmental exchange rates: 272

273 274  $Da = \frac{\text{root influx rate}}{\text{root-to-frond translocation rate}}$ (2)

275 When Da >> 1, the temporal changes of  $As_{frond}^{3+}$  are controlled primarily by the root-to-frond 276 translocation rate, while Da << 1 indicates that the root influx rate is the rate-determining step. The 277 Da for our plant nanobionic system is approximately 0.11 - 0.18, which indicates that the arsenite 278 influx from the uptake solution into the roots is the rate-determining step. As such, assuming the 279 root uptake follows a Michaelis-Menten kinetic model and the sequestration process (

280  $As_{frond}^{3+} \rightarrow As_{seq}^{3+}$ ) follows a first-order reaction, the mass balances of  $As_{frond}^{3+}$  and  $As_{sol}^{3+}$  can be 281 described with the following ordinary differential equations:

282  $\frac{dn_{frond}}{dt} = \frac{FW_{root}I_{max}C_{sol}}{K_m + C_{sol}} - k_d n_{frond}$ (3)

$$\frac{dC_{sol}}{dt} = -\frac{FW_{root}I_{max}C_{sol}}{V_{sol}(K_m + C_{sol})}$$
(4)

where  $n_{frond}$  denotes the amount of arsenite in the frond,  $FW_{root}$  is the root fresh weight,  $I_{max}$  is the maximum net influx rate of arsenite into the roots,  $C_{sol}$  is the arsenite concentration in the uptake solution,  $K_m$  is the Michaelis-Menten constant which is an inverse measure of the root transporters' affinity towards arsenite, and  $k_d$  is the first-order sequestration rate constant of

288 arsenite in the frond, and  $V_{sal}$  is the uptake volume solution. The proposed kinetic model can 289 describe changes in P. cretica frond arsenite concentration, obtained from the nanosensor intensity 290 profile, upon exposure to 10, 5, 1 and 0.1  $\mu$ M arsenite at the roots with high fidelity (Figure 4d). The fitting process yields three kinetic parameters for the *P. cretica* nanobionic system:  $K_m$  of 5.84 ± 291 1.63  $\mu$ M,  $I_{max}$  of 3.65 ± 1.02 nmol g<sup>-1</sup> root FW h<sup>-1</sup>, and  $k_{d}$  of 0.0012 ± 0.0004 h<sup>-1</sup>. The  $K_{m}$  value 292 estimated from our kinetic data is similar to those obtained for P. vittata plants previously reported 293 by other groups (**Table 1**). It is approximately 30 times lower than the  $K_m$  of arsenite transporters 294 in rice roots (180 µM),<sup>[64]</sup> indicating a higher affinity of *P. cretica* roots than rice roots towards 295 arsenite. The lower value of  $I_{max}$  to that of *P. vittata* indicates a slightly slower arsenite net uptake 296 297 rate into the roots of *P. cretica* than *P. vittata*. The low value of  $k_d$  suggests that while the 298 nanosensor detection mechanism is reversible, P. cretica hyperaccumulator takes up arsenite almost 299 irreversibly and the arsenite detection by plant nanobionic can therefore be considered irreversible. 300 The proposed kinetic model can be used to predict the arsenite concentration in the frond as 301 a function of plant root biomass, uptake solution volume and uptake duration (Figure S8, Supporting Information). We further utilized the model to estimate the theoretical detection limit of P. cretica-302 303 based nanobionic arsenite sensor, defined as the minimum arsenite concentration in the root uptake solution that can be detected by the plant nanobionic sensor. The minimum frond arsenite 304 305 concentration that gave a signal-to-background ratio  $\geq$  3 was determined to be 110 nmol g<sup>-1</sup> frond dry weight (DW). The kinetic model was then utilized to compute the limit of detection in the uptake 306 307 solution that results in this level of arsenite in the frond under different experimental conditions and 308 root biomass. Considering an uptake period of 7 days, a limit of detection of 4.7 nM (0.6 ppb) could be achieved with roots of 30 g FW and uptake solution volume of 5 L (Figure 4e). This detection limit 309 310 suggests that the plant nanobionic sensor can be used to monitor arsenite levels well below the regulatory limit of arsenic in drinking water (10 ppb) and in irrigation water (100 ppb). This figure of 311 merit is also lower than the detection limit of the G-SWNT nanosensor alone (122 nM; 15.8 ppb), 312 highlighting the ability of *P. cretica* to pre-concentrate and hyperaccumulate arsenite to increase the 313 detection sensitivity of a nanobionic sensor. A lower detection limit can be achieved at longer 314 uptake duration with a larger root biomass and a higher uptake solution volume (Figure S9, 315 316 Supporting Information). When the uptake period is extended to 14 days, the detection limit of the 317 plant nanobionic sensor can be reduced to 1.6 nM (0.2 ppb) with roots of 30 g FW and uptake solution volume of 5 L (Figure 4f). While this limit of detection may not apply to rice or spinach 318 319 plants tested earlier, these species may still constitute useful plant-based sensors to monitor arsenic 320 accumulation in heavily-contaminated areas, as well as in edible plants for food safety evaluation and plant science studies. 321

#### 322 Competitive inhibition of silicon on arsenite uptake in Pteris cretica

The ability of ferns in the *Pteridaceae* family to tolerate and hyperaccumulate exceptionally high levels of arsenic appears to result from the presence of certain genes and proteins recently

325 identified in P. vittata. Arsenite antiporter gene ACR3 was found to be necessary for arsenic 326 tolerance in *P. vittata* gametophytes by mediating the vacuolar sequestration of arsenite.<sup>[59]</sup> Similarly, the GAPC1, GSTF1 and OCT4 proteins are required for the import and reduction of 327 arsenate inside the cells.<sup>[65]</sup> However, the uptake pathway of arsenite in *Pteridaceae* ferns has not 328 been fully elucidated, partly due to the difficulty in generating transgenic ferns.<sup>[66]</sup> In rice, the uptake 329 of arsenite into rice roots is primarily facilitated by OsNIP2;1 (Ls1), a member of the nodulin-26 like 330 intrinsic proteins (NIPs) that is also responsible for the uptake of silicon (Si).<sup>[12,67]</sup> In *P. vittata*, the 331 aquaporin tonoplast intrinsic protein 4 (TIP4) is the only channel to date that has been shown to 332 mediate arsenite uptake.<sup>[68]</sup> It is unknown if arsenite uptake in *Pteridaceae* ferns share the same 333 334 pathways or transporters as those responsible for Si influx into the roots. In this study, we used the 335 optical nanosensors to investigate the effect of Si on arsenite uptake in *P. cretica*. As expected, the fluorescence intensity of (GT)<sub>5</sub>-SWNT showed a steady increase upon exposure to 10  $\mu$ M arsenite for 336 337 5 hours, while that of  $C_{10}$ -SWNT remained invariant (Figure 5a). The addition of silicic acid to the medium suppressed arsenite uptake by *P. cretica*, as shown by the slower and insignificant change in 338 embedded (GT),-SWNT intensity after 5 hours of treatment (Figure 5a). Image analysis showed that 339 340 while the mean relative intensity of  $(GT)_5$ -SWNT to  $C_{10}$ -SWNT increased by approximately 15% after 341 5-hour exposure to arsenite, the presence of silicic acid in the arsenite uptake medium led to negligible arsenite accumulation in P. cretica. To ascertain that this competitive inhibition effect can 342 343 be attributed uniquely to silicic acid, the fern roots were also subjected to an uptake medium containing both arsenite and phosphate. Extensive physiological data across plant species have 344 shown that phosphate and arsenate uptake are mediated by the same transporters.<sup>[61,69–72]</sup> The 345 presence of phosphate did not inhibit arsenite uptake in P. cretica as monitored by our nanosensors 346 347 (Figure 5a). The sensor response was similar in terms of magnitude and temporal profile to the case where only arsenite was present in the medium. The average relative intensity of  $(GT)_{5}$ -SWNT to  $C_{10}$ -348 349 SWNT increased by 13% after 5-hour exposure to the root uptake medium containing both 350 phosphate and arsenite (Figure 5b). As an example of the novel utility of the sensors introduced in this work, these results indicate that the arsenite and Si uptake in *P. cretica* may share the same 351 transport systems previously identified in rice.<sup>[12,73]</sup> The application of our nanosensors to investigate 352 mechanisms of arsenite uptake in *P. cretica* further illustrates the versatility of our plant nanobionic 353 approach, which can be utilized for the creation of a new class of sensors as well as to aid botany 354 research. 355

## 356 CONCLUSIONS

357 In this work, we demonstrate the integration of nanoparticles with living plants to engineer plant nanobionic sensors capable of real-time detection of arsenite in the belowground 358 359 environment. DNA-wrapped SWNT nanosensors were rationally designed using the CoPhMoRe technique for selective and sensitive arsenite detection. These nanoconstructs can be incorporated 360 361 into the tissues of wild-type plants and remained sensitive in vivo, enabling the conversion of living 362 plants into microfluidic arsenite detectors capable of autosampling their surroundings through 363 natural transpiration. Surface modification of DNA-SWNT constructs allows the versatile use of these 364 sensors in both the NIR and visible region for whole plant and subcellular imaging. We also showed 365 that the sensitivity of plant nanobionic sensors can be significantly enhanced by exploiting the

366 hyperaccumulating capability of select species such as P. cretica. Such plants exhibit high capacity in 367 arsenite extraction from the belowground environment and its translocation to their fronds. In 368 addition, hyperaccumulators can tolerate high concentrations of arsenite, promoting their use as 369 sensitive sensing devices in their natural environment. The increased sensitivity of plant nanobionic sensors compared to optical nanosensors alone illustrates the synergistic properties of plant 370 nanobionic devices by actively pre-concentrating specific analytes in vivo and enabling the 371 communication of this analyte through an optical signal easily intercepted by electronic devices. We 372 envision that the ability of select plants to pre-concentrate and hyperaccumulate specific analytes, 373 374 resulting in a much higher internal concentration without showing any signs of toxicity, can be 375 extended to engineer other plant nanobionic sensors for environmental monitoring applications. Hyperaccumulators of other metalloids or trace elements, an extensive list of which has been 376 377 compiled,<sup>[74]</sup> can be potentially converted into ultrasensitive detectors of their environment with the proposed plant nanobionic approach. The versatility of plant nanobionics was also shown through 378 379 sensor application in plant science research to investigate the uptake pathways of arsenite in P. 380 cretica. This new class of nanobionic sensors should find immediate utility in environmental 381 monitoring and agronomic studies.

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#### 383 EXPERIMENTAL METHODS

- 384 Materials
- 385 All reagents were purchased from Sigma Aldrich unless otherwise stated.

## 386 Preparation of DNA-SWNT nanoconstructs

Raw HiPCO SWNTs were obtained from NanoIntegris (Lot #HR27-104). Single stranded DNA
oligonucleotides were purchased from Integrated DNA Technologies. 1 mg of SWNT was mixed with
0.25 mg of ssDNA in 1 mL of 0.1 M NaCl. The mixture was sonicated with 3 mm probe tip (ColeParmer) at 40% amplitude for 20 min in an ice bath. The sample was then centrifuged at 30,000 g for
90 minutes to remove unsuspended SWNT aggregates. The collected supernatant was dialyzed
against 0.1 M NaCl with a 20 kDa MWCO dialysis bag (Spectra-Por) for three days to remove excess
ssDNA.

TP-(GT)<sub>5</sub>-SWNT and TP-C<sub>10</sub>-SWNT prepared according to previously published method with slight modification.<sup>[75]</sup> Briefly, 1 mg of SWNT was mixed with 0.25 mg of ssDNA and TP solution at a dye:ssDNA ratio of 1:4 in 1 mL of deionized water. Tip-sonication and centrifugation were carried as described above in DNA-SWNT nanosconstruct preparation. The collected supernatant was dialyzed

against deionized water with a 20 kDa MWCO dialysis bag (Spectra-Por) for three days.

## 399 Absorption spectra measurement

The UV-VIS absorption spectra of DNA-SWNT and TP-labelled DNA-SWNT were collected using a quartz cuvette (Starna) with 1-cm path length in Shimadzu UV-3101PC spectrophotometer. All absorption spectra were background-subtracted using reference solutions. The concentration of the DNA-SWNT nanosensors was determined using its absorbance at 632 nm and extinction coefficient of 0.036 L mg<sup>-1</sup> cm<sup>-1</sup>.

## 405 Plant growth

Seeds of carmel spinach (Spinacia oleracea) were purchased from David's Garden Seeds. Seeds of 406 indica rice cultivar (Oryza sativa; IR24) were kindly donated by Professor Bing Yang laboratory 407 408 (Donald Dapforth Plant Science Center, St. Louis, MO). Cretan brake fern (Pteris cretica) plants were obtained from Josh's Frogs. Plant seeds were grown in Fafard Professional all-purpose blend potting 409 soil in a Conviron Adaptis 1000 growth chamber. Spinach plants and Cretan brake ferns were grown 410 with a 14-h-light/10-h-dark photoperiod at 100  $\mu$ mol s<sup>-1</sup> m<sup>-2</sup>, 60% relative humidity, and day/night 411 temperatures of 22 and 18°C respectively. Rice seeds were first germinated at 37°C for 4 days. Rice 412 seedlings were then washed carefully with water and transplanted to potting soils in growth 413 414 chamber with 12-h-light/12-h-dark photoperiod at 100  $\mu$ mol s<sup>-1</sup> m<sup>-2</sup>, 60% relative humidity, and day/night temperature of 28 and 25°C respectively. Fertilizer (N:P:K = 15:9:12) was applied to the 415 potting soil every two weeks. 416

#### 417 Nanosensor screening and selectivity test

- 418 DNA-SWNT nanoconstructs were diluted with 0.1 M NaCl, MES buffer (10 mM MES, 10 mM MgCl2,
- pH 5.7), or TES buffer (10 mM TES, 10 mM MgCl<sub>2</sub>, pH 7.5) to a concentration of 2 mg  $L^{-1}$ . Aliquots of
- 420 SWNT suspensions were added to a 96-well plate for high-throughput screening. The fluorescence
- 421 spectra of DNA-SWNT complexes were recorded with a custom-made NIR microscope array before
- and after a 30-minute incubation of SWNT aliquots with 100  $\mu$ M heavy metal cations. Briefly, the 96-
- 423 well plate was mounted on a motorized stage of a Zeiss AxioVision inverted microscope connected
- to a 1D InGaAs detector (Princeton instruments) with a PI Acton SP2500 spectrometer. The samples
   were excited with a 785-nm photodiode laser (Invictus) at the sample plane with x20 objective for a
- 426 10-s exposure time. The fluorescence intensity at 1128 nm wavelength, corresponding to the (9,4)
- 427 chirality, was used to compare the sensor selectivity and response  $((I-I_0)/I_0)$ , where  $I_0$  is the initial
- 428 fluorescence intensity before analyte addition and I is the fluorescence intensity after analyte
- 429 addition. The sensor responses of TP-(GT)<sub>5</sub>-SWNT and TP- $C_{10}$ -SWNT was measured in the NIR range
- 430 using the method described above, and in the visible range using a Varioskan Flash microplate
- 431 reader (Thermo Scientific). The fluorescence intensity corresponding to the maximum fluorescence
- 432 peak at 540 nm was used to obtain the visible sensor response.

### 433 Nanosensor infiltration and standoff imaging of arsenite uptake

434 Spinach, rice and fern plants were infiltrated with both (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT, which were prepared at 5 mg L<sup>1</sup> concentration in MES buffer. Gentle pressure was applied to the abaxial side of 435 the leaf to ensure no damage was inflicted during the needleless syringe infiltration of the 436 nanosensors. For rice, a small puncture was introduced to the leaf surface using a pipette tip, after 437 438 which DNA-SWNT was infiltrated through the puncture with a gentle pressure during syringe 439 infiltration. The nanosensors were infiltrated to opposite sides of the leaf midrib. The plant roots 440 were then washed with deionized water carefully to remove the soils and transferred to a 441 pretreatment solution containing 10 mM KCl and 5 mM MES of pH 6.0. A 785 nm laser was used to 442 excite the embedded DNA-SWNT complexes. The NIR fluorescence intensity of both sensors were 443 spatiotemporally monitored at a standoff distance of 1 m with a 2D InGaAs array (Princeton 444 Instruments OMA V) equipped with a Nikon AF Micro-Nikkor 60 mm f/2.8D lens. Images were 445 collected at a 2-second exposure time unless otherwise stated. A 900 nm long-pass filter was placed 446 in front of the camera lens to eliminate chlorophyll autofluorescence and the reflected excitation 447 beam. After 20 minutes, the pretreatment solution incubating the roots was then replaced with a 448 solution containing 10  $\mu$ M or 0.1  $\mu$ M of sodium meta-arsenite (NaAsO<sub>2</sub>) unless otherwise stated.

For 7-day experiments, 10 images were taken daily at a specific time during the day and the fluorescence intensity was averaged from the 10 images for a daily profile. The excitation laser was turned off when images were not collected. Roots of treated plants were weighed to obtain their fresh weights. To calibrate the nanosensor fluorescence intensities in *Pteris cretica*, the frond arsenic concentration was analyzed with ICP-MS by Galbraith Laboratories Inc. (Knoxville, TN). Briefly, frond samples were cut at specific time points after arsenite treatment, rinsed with deionized water and dried at 60°C for 48 hours. They were then weighed, ground to fine powder and analyzed by ICP-MS.

### 456 Image and data analysis

457 Image and data analysis were performed with ImageJ and Matlab R2018a. The sensor response was 458 obtained from imaging experiments by normalizing the SWNT fluorescence with the corresponding initial value\_prior to the introduction of arsenite. 100 brightest pixels in a sensor spot were averaged 459 to obtain a mean fluorescence intensity value. The normalized (GT)<sub>5</sub>-SWNT intensity profile was 460 divided with that of C<sub>10</sub>-SWNT to yield a ratiometric sensor profile (I<sub>G/C-SWNT</sub>). In Figure 4, the change 461 in ratiometric sensor profile relative to the initial value is denoted as  $\Delta I/I_0$ . Snapshots of false-colored 462 images were generated by subtracting the first image collected at the time point of arsenite 463 introduction from subsequent images. Numerical simulation of the kinetic model to fit arsenite 464 concentration profile in Pteris cretica was performed with ode45 solver in Matlab R2018a. The 465 466 sensor limit of detection was also simulated using Matlab R2018a.

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#### 467 Confocal microscopy

- 468 Confocal images were obtained using a Zeiss LSM 710 microscope. 20 μL of TP-(GT)<sub>5</sub>-SWNT in MES
- 469 buffer was infiltrated into spinach leaves as attached to living plants. The leaf was excised 1 hour
- 470 after infiltration and a leaf disc (5 mm) was prepared using a cork borer. The leaf disc was then
- 471 transferred to a glass slide with a polydimethylsiloxane (PDMS) chamber filled with water. The
- 472 chamber was sealed with a cover slip and imaged with a 40x water-immersion objective. A final
- 473 concentration of 10  $\mu$ M arsenite solution was introduced into the chamber medium from the side of
- 474 the cover slip with a pipette. TP-(GT)<sub>5</sub>-SWNT nanosensors were excited with a 514 nm laser with
- emission channel from 530 to 570 nm. Chlorophyll autofluorescence imaging was obtained by
- 476 excitation at 633 nm with emission channel between 660 and 750 nm. Confocal images of TP-(GT)<sub>5</sub>-
- 477 SWNT and chloroplast autofluorescence were captured every 5 minutes.

### 478 Silicon, phosphate and arsenite interaction experiments

- 479 *Pteris cretica plants* were used in these experiments. The roots of the plants were gently removed
- 480 from the soil and carefully washed with water. The plants were then transferred to a beaker
- 481 containing a pretreatment solution of 10 mM KCl and 5 mM MES at pH 6.0. After nanosensor
- 482 infiltration, plants were incubated in the growth chamber for 1 hour before imaging. At the start of
- 483 the experiments, the pretreatment solution was replaced with solutions containing 10 μM arsenite,
- 484 10 μM arsenite and 100 μM phosphate (supplied as KH<sub>2</sub>PO<sub>4</sub>), or 10 μM arsenite and 100 μM silicic
- 485 acid which was prepared from  $SiO_2$  according to a previously published method.<sup>[76]</sup>



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- 501 T.T.S.L. and M.S.S conceived the project, designed the study and wrote the manuscript. M.P. and
- 502 J.C. assisted with data analysis and nanoparticle characterization.
- 503 Competing interests
- 504 The authors declare no competing interest.

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#### 505 References

- 506 [1] A. H. Smith, P. A. Lopipero, M. N. Bates, C. M. Steinmaus, *Science (80-. ).* 2002, 296, 2145.
- 507 [2] H. Brammer, P. Ravenscroft, *Environ. Int.* **2009**, *35*, 647.
- 508 [3] M. Argos, T. Kalra, P. J. Rathouz, Y. Chen, B. Pierce, F. Parvez, T. Islam, A. Ahmed, M. Rakibuz509 Zaman, R. Hasan, G. Sarwar, V. Slavkovich, A. Van Geen, J. Graziano, H. Ahsan, *Lancet* 2010,
  510 376, 252.
- 511[4]K. S. Mohammed Abdul, S. S. Jayasinghe, E. P. S. Chandana, C. Jayasumana, P. M. C. S. De512Silva, Environ. Toxicol. Pharmacol. 2015, 40, 828.
- [5] R. Singh, S. Singh, P. Parihar, V. P. Singh, S. M. Prasad, *Ecotoxicol. Environ. Saf.* 2015, *112*, 247.
  [6] S. Quazi, D. Sarkar, R. Datta, *J. Hazard. Mater.* 2013, *262*, 1031.
- 515 [7] R. Q. Huang, S. F. Gao, W. L. Wang, S. Staunton, G. Wang, *Sci. Total Environ.* **2006**, *368*, 531.
- 516 [8] E. M. Muehe, T. Wang, C. F. Kerl, B. Planer-Friedrich, S. Fendorf, *Nat. Commun.* **2019**, *10*, 1.
- 517 [9] A. L. Seyfferth, S. McCurdy, M. V. Schaefer, S. Fendorf, *Environ. Sci. Technol.* 2014, 48, 4699.
- 518[10]N. US EPA, OCSPP, OPPT, "Chemical Contaminant Rules \_ Drinking Water Requirements for519States and Public Water Systems \_ US EPA," can be found under
- 520 https://www.epa.gov/dwreginfo/chemical-contaminant-rules, **2016**.
- 521 [11] D. E. Mays, A. Hussam, Anal. Chim. Acta 2009, 646, 6.
- 522 [12] F. M. Jian, N. Yamaji, N. Mitani, X. Y. Xu, Y. H. Su, S. P. McGrath, F. J. Zhao, *Proc. Natl. Acad.* 523 *Sci. U. S. A.* 2008, *105*, 9931.
- [13] W. Y. Song, T. Yamaki, N. Yamaji, D. Ko, K. H. Jung, M. Fujii-Kashino, G. An, E. Martinoia, Y.
   Lee J. Feng, Proc. Natl. Acad. Sci. U. S. A. 2014, 111, 15699.
- 526 [14] A. H. Ackerman, P. A. Creed, A. N. Parks, M. W. Fricke, C. A. Schwegel, J. T. Creed, D. T.
- 527 Heitkemper, N. P. Vela, *Environ. Sci. Technol.* **2005**, *39*, 5241.
- 528 [15] D. Melamed, Anal. Chim. Acta **2005**, 532, 1.
- 529 [16] E. Durduran, H. Altundag, M. Imamoglu, S. Z. Yildiz, M. Tuzen, *J. Ind. Eng. Chem.* 2015, 27,
  530 245.
- 531 [17] J. Mattusch, R. Wennrich, A. C. Schmidt, W. Reisser, *Fresenius. J. Anal. Chem.* **2000**, *366*, 200.
- 532 [18] D. K. Gupta, S. Chatterjee, Arsen. Contam. Environ. Issues Solut. 2017, 1.
- 533 [19] T. Shi, H. Liu, J. Wang, Y. Chen, T. Fei, G. Wu, *Environ. Sci. Technol.* **2014**, *48*, 6264.
- [20] V. Bandaru, C. Daughtry, E. Codling, D. Hansen, S. White-Hansen, C. Green, *Int. J. Environ. Res. Public Health* 2016, *13*, 606.
- 536 [21] X. Li, X. Liu, M. Liu, C. Wang, X. Xia, Int. J. Appl. Earth Obs. Geoinf. 2015, 36, 41.
- 537 [22] W. Zhou, J. Zhang, M. Zou, X. Liu, X. Du, Q. Wang, Y. Liu, Y. Liu, J. Li, Sci. Rep. 2019, 9, 1.
- 538 [23] K. Mao, H. Zhang, Z. Wang, H. Cao, K. Zhang, X. Li, Z. Yang, *Biosens. Bioelectron.* 2020, 148, 111785.
- 540 [24] P. Devi, A. Thakur, R. Y. Lai, S. Saini, R. Jain, P. Kumar, *TrAC Trends Anal. Chem.* 2019, 110,
  541 97.
- 542 [25] L. Cui, J. Wu, H. Ju, Biosens. Bioelectron. **2016**, *79*, 861.
- [26] K. Matsunaga, Y. Okuyama, R. Hirano, S. Okabe, M. Takahashi, H. Satoh, *Chemosphere* 2019, 224, 538.
- 545 [27] S. H. Wen, Wang, Y. H. Yuan, R. P. Liang, J. D. Qiu, Anal. Chim. Acta **2018**, 1002, 82.
- 546 [28] J. Zhang, M. P. Landry, P. W. Barone, J. H. Kim, S. Lin, Z. W. Ulissi, D. Lin, B. Mu, A. A.
- 547Boghossian, A. J. Hilmer, A. Rwei, A. C. Hinckley, S. Kruss, M. A. Shandell, N. Nair, S. Blake, F.548Şen, S. Şen, R. G. Croy, D. Li, K. Yum, J. H. Ahn, H. Jin, D. A. Heller, J. M. Essigmann, D.
- 549 Blankschtein, M. S. Strano, *Nat. Nanotechnol.* **2013**, *8*, 959.
- J. P. Giraldo, M. P. Landry, S. M. Faltermeier, T. P. McNicholas, N. M. Iverson, A. A.
  Boghossian, N. F. Reuel, A. J. Hilmer, F. Sen, J. A. Brew, M. S. Strano, *Nat. Mater.* 2014, *13*,
  400.

- [30] M. H. Wong, J. P. Giraldo, S. Y. Kwak, V. B. Koman, R. Sinclair, T. T. S. Lew, G. Bisker, P. W. Liu,
   M. S. Strano, *Nat. Mater.* 2017, *16*, 264.
- T. T. S. Lew, V. B. Koman, K. S. Silmore, J. S. Seo, P. Gordiichuk, S.-Y. Kwak, M. Park, M. C.-Y.
  Ang, D. T. Khong, M. A. Lee, M. B. Chan-Park, N.-H. Chua, M. S. Strano, *Nat. Plants* 2020, *6*,
  404.
- 558 [32] S.-Y. Kwak, M. H. Wong, T. T. S. Lew, G. Bisker, M. A. Lee, A. Kaplan, J. Dong, A. T. Liu, V. B.
   559 Koman, R. Sinclair, C. Hamann, M. S. Strano, *Annu. Rev. Anal. Chem.* 2017, *10*, 113.
- 560 [33] T. T. S. Lew, V. B. Koman, P. Gordiichuk, M. Park, M. S. Strano, *Adv. Mater. Technol.* 2020, *5*,
  561 1900657.
- 562 [34] T. T. S. Lew, M. H. Wong, S. Y. Kwak, R. Sinclair, V. B. Koman, M. S. Strano, *Small* 2018, 14, 1802086.
- 564 [35] S. Y. Kwak, T. T. S. Lew, C. J. Sweeney, V. B. Koman, M. H. Wong, K. Bohmert-Tatarev, K. D.
   565 Snell, J. S. Seo, N. H. Chua, M. S. Strano, *Nat. Nanotechnol.* **2019**, *14*, 447.
- 566 [36] T. T. S. Lew, M. Park, Y. Wang, P. Gordiichuk, W.-C. Yeap, S. K. Mohd Rais, H.
   567 Kulaveerasingam, M. S. Strano, *ACS Mater. Lett.* 2020, *2*, 1057.
- 568 [37] Y. Takahashi, R. Minamikawa, K. H. Hattori, K. Kurishima, N. Kihou, K. Yuita, *Environ. Sci.*569 *Technol.* 2004, *38*, 1038.
- 570 [38] N. Li, J. Wang, W.-Y. Song, *Plant Cell Physiol.* **2016**, *57*, 4.
- 571 [39] S. Nafisi, A. Sobhanmanesh, K. Alimoghaddam, A. Ghavamzadeh, H. A. Tajmir-Riahi, DNA Cell
  572 Biol. 2005, 24, 634.
- 573 [40] L. Zhang, X. Z. Cheng, L. Kuang, A. Z. Xu, R. P. Liang, J. D. Qiu, *Biosens. Bioelectron.* 2017, *94*,
  574 701.
- 575 [41] A. G. Beyene, A. A. Alizadehmojarad, G. Dorlhiac, N. Goh, A. M. Streets, P. Král, L. Vuković, M.
  576 P. Landry, *Nano Lett.* 2018, *18*, 6995.
- 577 [42] M. Zheng, A. Jagota, M. S. Strano, A. P. Santos, P. Barone, S. G. Chou, B. A. Diner, M. S.
   578 Dresselhaus, R. S. McLean, G. B. Onoa, G. G. Samsonidze, E. D. Semke, M. Usrey, D. J. Watts,
   579 Science (80-.). 2003, 302, 1545.
- 580 [43] Y. Wu, S. Zhan, F. Wang, L. He, W. Zhi, P. Zhou, Chem. Commun. 2012, 48, 4459.
- 581 [44] M. P. Landry, L. Vuković, S. Kruss, G. Bisker, A. M. Landry, S. Islam, R. Jain, K. Schulten, M. S.
   582 Strano, *J. Phys. Chem. C* 2015, *119*, 10048.
- 583 [45] D. P. Salem, X. Gong, A. T. Liu, V. B. Koman, J. Dong, M. S. Strano, J. Am. Chem. Soc. 2017, 139,
   584 16791.
- 585 [46] D. P. Salem, M. P. Landry, G. Bisker, J. Ahn, S. Kruss, M. S. Strano, *Carbon N. Y.* **2016**, *97*, 147.
- 586 [47] G. M. Newkirk, H. Wu, I. Santana, J. P. Giraldo, J. Vis. Exp. 2018, 2018, 58373.
- 587 [48] S. Kruss, M. P. Landry, E. Vander Ende, B. M. A. Lima, N. F. Reuel, J. Zhang, J. Nelson, B. Mu, A.
   588 Hilmer, M. Strano, *J. Am. Chem. Soc.* **2014**, *136*, 713.
- [49] Y. G. Zhu, G. X. Sun, M. Lei, M. Teng, Y. X. Liu, N. C. Chen, L. H. Wang, A. M. Carey, C. Deacon,
   A. Raab, A. A. Meharg, P. N. Williams, *Environ. Sci. Technol.* **2008**, *42*, 5008.
- [50] H. F. Bakhat, Z. Zia, S. Fahad, S. Abbas, H. M. Hammad, A. N. Shahzad, F. Abbas, H. Alharby,
   M. Shahid, *Environ. Sci. Pollut. Res.* 2017, *24*, 9142.
- 593 [51] P. N. Williams, A. Villada, C. Deacon, A. Raab, J. Figuerola, A. J. Green, J. Feldmann, A. A.
  594 Meharg, *Environ. Sci. Technol.* 2007, *41*, 6854.
- 595 [52] Y. H. Su, S. P. McGrath, F. J. Zhao, *Plant Soil* **2010**, *328*, 27.
- 596 [53] G. L. Silva, V. Ediz, D. Yaron, B. A. Armitage, J. Am. Chem. Soc. 2007, 129, 5710.
- 597 [54] A. A. Meharg, New Phytol. 2003, 157, 25.
- 598 [55] M. Srivastava, L. Q. Ma, J. A. G. Santos, *Sci. Total Environ.* **2006**, *364*, 24.
- 599 [56] Z. Souri, N. Karimi, L. M. Sandalio, Front. Cell Dev. Biol. 2017, 5, DOI
- 600 10.3389/fcell.2017.00067.

- 601 [57] L. Q. Ma, K. M. Komar, C. Tu, W. Zhang, Y. Cai, E. D. Kennelley, *Nature* **2001**, *409*, 579.
- 602 [58] S. Jeong, H. S. Moon, K. Nam, J. Hazard. Mater. 2014, 280, 536.
- 603 [59] E. Indriolo, G. N. Na, D. Ellis, D. E. Salt, J. A. Banks, *Plant Cell* **2010**, *22*, 2045.
- 604 [60] H. Shen, Z. He, H. Yan, Z. Xing, Y. Chen, W. Xu, W. Xu, M. Ma, J. Proteomics 2014, 105, 46.
- [61] J. Wang, F. J. Zhao, A. A. Meharg, A. Raab, J. Feldmann, S. P. McGrath, *Plant Physiol.* 2002,
  130, 1552.
- 607 [62] X. Wan, M. Lei, T. Chen, J. Yang, H. Liu, Y. Chen, Environ. Sci. Pollut. Res. 2015, 22, 16631.
- 608 [63] Y. H. Su, S. P. McGrath, Y. G. Zhu, F. J. Zhao, New Phytol. **2008**, 180, 434.
- 609 [64] A. A. Meharg, L. Jardine, *New Phytol.* **2003**, *157*, 39.
- 610 [65] C. Cai, N. A. Lanman, K. A. Withers, A. M. DeLeon, Q. Wu, M. Gribskov, D. E. Salt, J. A. Banks,
   611 *Curr. Biol.* 2019, *29*, 1625.
- 612 [66] B. Muthukumar, B. L. Joyce, M. P. Elless, C. N. Stewart, *Plant Physiol.* **2013**, *163*, 648.
- 613 [67] J. F. Ma, N. Yamaji, *Trends Plant Sci.* **2006**, *11*, 392.
- [68] Z. He, H. Yan, Y. Chen, H. Shen, W. Xu, H. Zhang, L. Shi, Y. G. Zhu, M. Ma, *New Phytol.* 2016, 209, 746.
- 616 [69] Y. G. Zhu, C. N. Geng, Y. P. Tong, S. E. Smith, F. A. Smith, Ann. Bot. 2006, 98, 631.
- 617 [70] S. FJ DITUSA, E. B. Fontenot, R. W. Wallace, M. A. Silvers, T. N. Steele, A. H. Elnagar, K. M.
  618 Dearman, A. P. Smith, *New Phytol.* 2016, 209, 762.
- 619 [71] A. A. Meharg, J. Hartley-Whitaker, *New Phytol.* **2002**, *154*, 29.
- 620 [72] G. T. Clark, J. Dunlop, H. T. Phung, Aust. J. Plant Physiol. 2000, 27, 959.
- 621 [73] F. J. Zhao, J. F. Ma, A. A. Meharg, S. P. McGrath, New Phytol. 2009, 181, 777.
- 622 [74] R. D. Reeves, A. J. M. Baker, T. Jaffré, P. D. Erskine, G. Echevarria, A. van der Ent, *New Phytol.* 623 **2018**, *218*, 407.
- 624 [75] O. Cavuslar, H. Unal, *RSC Adv.* **2015**, *5*, 22380.
- 625 [76] X. Gao, R. A. Root, J. Farrell, W. Ela, J. Chorover, *Appl. Geochemistry* **2013**, *38*, 110.
- 626 [77] F. Wu, D. Deng, S. Wu, X. Lin, M. H. Wong, *Environ. Sci. Pollut. Res.* **2015**, *22*, 8911.
- [78] X. Wang, L. Q. Ma, B. Rathinasabapathi, Y. Cai, Y. Guo Liu, G. Ming Zeng, *Environ. Sci. Technol* **2011**, 45, 9719.
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631 FIGURES

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633 Figure 1. Screening and characterization of SWNT-based arsenite sensors. a) Comparison of sensor 634 responses  $((I_1-I_0)/I_0)$  of  $(GT)_N$ -SWNT, where N = 5 to 30, and other DNA-SWNT constructs against arsenite. Data represent mean  $\pm$  s.d. from n = 3 independent experiments. b) DNA-SWNT responses 635 towards arsenate. Data represent mean ± s.d. from n = 3 independent experiments. c) NIR 636 fluorescence spectra of (GT)<sub>5</sub>-SWNT, (GT)<sub>25</sub>-SWNT and C<sub>10</sub>-SWNT before and after exposure to 100 637  $\mu$ M arsenite. d) Calibration curve of (GT)<sub>5</sub>-SWNT against different concentrations of arsenite. Fitting 638 with kinetic adsorption model is shown in black. e) Selectivity of (GT)<sub>5</sub>-SWNT against other heavy 639 metal cations commonly present in the soil or groundwater. Data represent mean  $\pm$  s.d. from n = 3 640 641 independent experiments.



643 Figure 2. Standoff detection of arsenite introduced via root uptake in spinach and rice plants. a)

Brightfield image of spinach leaf infiltrated with (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT under 785-nm excitation. 644 Scale bar, 0.5 mm. b) Screenshots of false-colored images showing the change in fluorescence 645 intensity of embedded (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT after exposure to arsenite. Time label denotes the 646 647 time points after the introduction of arsenite. c) Normalized intensity profile of (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT embedded in spinach plants upon exposure to 10 µM arsenite root medium. Shaded region 648 represents s.e.m. from n = 5 independent biological samples. Black arrow denotes the time point of 649 arsenite introduction. d) Time profile of ratiometric sensor upon spinach plant exposure to arsenite 650 651 and deionized water in the root medium. Shaded region represents s.e.m. from n = 5 independent 652 biological samples. Black arrow denotes the time point of arsenite introduction. e) Brightfield image of rice blade infiltrated with (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT under 785-nm excitation. Scale bar, 0.5 mm. 653 f) Time profile of ratiometric sensor upon rice plant exposure to arsenite and deionized water in the 654 root medium. Shaded region represents s.e.m. from n = 5 independent biological samples. Black 655 arrow denotes the time point of arsenite introduction. 656

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Figure 3. Sensor characterization and application of TP-(GT)<sub>5</sub>-SWNT constructs in confocal imaging 658 659 of arsenite accumulation. a) Absorbance spectra of (GT)<sub>5</sub>-SWNT and TP-(GT)<sub>5</sub>-SWNT confirm the successful conjugation of TP dye as shown by the presence of the 515 nm peak. b) NIR spectra of 660 (GT)<sub>5</sub>-SWNT and TP-(GT)<sub>5</sub>-SWNT before and after 100 μM arsenite exposure. c) Changes in the 661 662 fluorescence of TP-(GT)<sub>5</sub>-SWNT and TP-C<sub>10</sub>-SWNT upon arsenite exposure as measured by the emission intensity at 540 nm. Data represent mean  $\pm$  s.d. from n = 3 independent experiments. d) 663 Confocal images capturing the spatial and temporal profile of TP-(GT)<sub>5</sub>-SWNT in the mesophyll cell of 664 spinach leaf lamina exposed to arsenite. Changes in the fluorescence intensity ( $\Delta I/I_0$ ) of TP-(GT)<sub>5</sub>-665 SWNT were false-colored for clarity. Scale bar, 10 µm. e) Fluorescence time profile of 3 randomly 666 selected regions of interest (ROI) containing TP-(GT)<sub>5</sub>-SWNT corresponding to circled areas in (d) 667 668 after arsenite exposure.

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670 Figure 4. Arsenite detection with nanobionic sensor based on arsenic hyperaccumulator Pteris cretica. a) Brightfield image of Pteris cretica frond infiltrated with (GT)<sub>5</sub>-SWNT and C<sub>10</sub>-SWNT under 671 785-nm excitation. Scale bar, 0.5 mm. b) Time lapse images showing the intensity changes of 672 673 embedded nanosensors upon arsenite exposure. Time denotes the time points after arsenite 674 introduction via root uptake. c) Fluorescence intensity changes of SWNT nanosensors embedded in spinach, rice and Pteris cretica plants exposed to 10 µM arsenite root medium. Data represent mean 675 676  $\pm$  s.e.m. from n = 5 Independent biological samples. d) Arsenite concentration in *Pteris cretica* frond treated with 10  $\mu$ M, 5  $\mu$ M, 1  $\mu$ M, 0.1  $\mu$ M arsenite solution and deionized water in the root medium. 677 The concentrations in nmol g<sup>-1</sup> DW are translated from sensor intensity responses. Data represent 678 mean ± s.e.m. from n = 5 independent biological samples. e) Contour plot of plant nanobionic 679 680 sensor's limit of detection as a function of uptake solution volume and root fresh weight after 7 681 days. Cross indicates the minimum detection limit of 4.7 nM (0.6 ppb). f) Contour plot of plant 682 nanobionic sensor's limit of detection as a function of uptake solution volume and root fresh weight after 14 days. Cross indicates the minimum detection limit of 1.6 nM (0.2 ppb). 683

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686 Figure 5. Application of nanosensors to investigate the arsenite uptake pathway in *Pteris cretica*.

a) False-colored images showing the embedded nanosensor response upon the exposure of *Pteris* 

688 *cretica* roots to medium containing only arsenite, arsenite and silica, or arsenite and phosphate.

689 Green and <u>orange arrows</u> correspond to  $(GT)_5$ -SWNT and  $C_{10}$ -SWNT respectively. Scale bars, 0.5 mm.

b) Time profile of normalized nanosensor intensity upon exposure to the different root media.

691 Shaded region represents s.e.m. from n = 5 independent biological samples.

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- 692 **Table 1.** Kinetic parameters estimated from the time profile of arsenite uptake in *Pteris cretica*
- 693 reported by nanosensors. Values obtained from model fitting are in agreement with reported values
- 694 from previously published reports.

Kinetic parameters	Fitting values	Reported values <sup>a</sup>	References
К <sub>м</sub> (µМ)	5.84 ± 1.63	6 - 25	[77,78]
I <sub>max</sub> (nmol g <sup>1</sup> root FW h <sup>1</sup> )	3.35 ± 1.02	8 - 10	[61]
k <sub>d</sub> (h <sup>2</sup> )	(1.23 ± 0.42) x 10 <sup>-3</sup>	N.A.	

- <sup>a</sup>Reported values were based on kinetic data on *Pteris vittata* plants.
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ToC Text:

Exploiting the natural ability of plants to pre-concentrate and extract arsenite from the belowground
environment, a living plant nanobionic sensor is engineered for non-destructive arsenite monitoring.
By embedding near-infrared fluorescent nanosensors within the mesophyll, living plants are

- 701 converted into environmental sensors for real-time arsenite detection. This demonstration opens
- new frontiers for plant-based sensors in environmental monitoring and food safety research.

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