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Accepted manuscript

1 **Technical Report**

2 **The present state-of-the-art of blue carbon repository in India: A meta-analysis**

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21 **Abstract**

22 The present study collated data on the Indian blue carbon repository (mangroves, seagrasses,
23 and salt marshes) from peer-reviewed literature on carbon stock assessment. This meta-
24 analysis indicated that the blue carbon ecosystems of India could have a collective carbon
25 stock of 67.35 Tg C (mangroves, seagrass, and salt marsh accounting for 67 Tg C, 0.0630 Tg
26 C, and 0.0049 Tg C, respectively). Several studies have ubiquitously measured the spatial
27 extent of mangroves (~4991 km²) and seagrasses (~517 km²) in India; however, the salt
28 marshes (290 – 1398 km²) have contradictions in estimates. The green payments against the
29 blue carbon ecosystems of India can be as high as ~ 9.6 billion US \$, whereas the social cost
30 of carbon sequestered by these ecosystems can vary between 0.47 and 5.43 billion US \$. The
31 present study also identified the key research areas that require priority to minimize the
32 uncertainties in blue carbon stock assessment to foster a robust ecosystem-based approach for
33 climate change adaptation in the country. The study identified that less than half of the total
34 mangrove habitats of India are yet to be sampled leaving a scope of substantial uncertainty in
35 nationwide blue carbon estimates. The spatial extent of India's salt marshes is another aspect
36 that needs to be delineated with a higher confidence level.

37 **Keywords:** Blue carbon; mangrove; seagrass; salt marsh; aboveground biomass;
38 belowground biomass; soil carbon pool

39 **1. Introduction**

40 Ever since the menace of anthropogenic greenhouse gas emissions has been recognized, the
41 global scientific community has been desperately looking for options to mitigate climate
42 change (Alongi, 2020). In this regard, the coastal vegetated ecosystems have been identified
43 to play a crucial role in sequestering atmospheric carbon dioxide (CO₂) that can enable us to
44 achieve climate change mitigation at the national and global scales (Taillardat et al., 2018).

45 The term blue carbon refers to the carbon stored by the vegetated coastal habitats throughout
46 the globe. By definition, blue carbon is the carbon captured by marine living organisms
47 (Nellemann et al., 2009). Mangroves, seagrasses meadows, and tidal salt marshes are the
48 conventional blue carbon ecosystems, which sequester carbon in their living aboveground
49 biomass (AGB) (leaves, stems, and branches), belowground biomass (BGB) (roots), non-
50 living biomass (like litter and deadwood), and underlying sediments (Bertram et al., 2021).
51 Blue carbon may sequester over the short term (decennial) in biomass and over longer
52 (millennial) time scales in sediments (Macreadie et al., 2017; 2019). The mangroves are
53 distributed from the tropics to subtropics, whereas seagrass is abundant from the polar to
54 tropical areas (Pendleton et al., 2012). The approximate estimates of blue carbon stock in
55 mangroves, seagrasses, and saltmarshes per unit area are 956 Mg C ha⁻¹, 142 Mg C ha⁻¹, and
56 593 Mg C ha⁻¹, respectively (Twilley et al., 1992; Donato et al., 2011; Fourqurean et al.,
57 2012; Alongi, 2014; Ouyang and Lee, 2014; Saderne et al., 2019). Carbon burial rates for the
58 mangrove and seagrass are 34.4 ± 5.9 Tg yr⁻¹ and 48–112 Tg yr⁻¹, respectively (Pendleton et
59 al., 2012 and the references therein). Ever since the global scientific community recognized
60 the ill effects of anthropogenic carbon emission and the menace of climate change, they have
61 been looking for solutions to combat this evil by reducing emissions from direct deforestation
62 and forest degradation (Andoh and Lee, 2018; Duchelle et al., 2018; Jackson and Sparks,
63 2020). Deforestation and forest degradation contribute to approximately 11 percent of carbon
64 emissions, and various United Nations (UN) mechanisms, including the widely-referred
65 Reducing emissions from deforestation and forest degradation (REDD+) program proposed
66 offering economic incentives for developing countries to reduce emissions. The mode of
67 operations includes reducing deforestation, halting forest degradation, conservation of forest
68 C stocks, enhancement of forest C stocks, and sustainable management of forests (Johnson et
69 al., 2019). In addition, under the Paris agreement, The Nationally Determined Contributions

70 (NDCs) provide an opportunity to report the extent of carbon captured as a measure of the
71 country's voluntary reduction of carbon emissions. The carbon sequestration potential of the
72 blue carbon ecosystems offers tremendous potential to offset the carbon emissions (Taillardat
73 et al., 2018; Alongi, 2020).

74 Several marine ecosystems do not meet the essential criteria for inclusion within the
75 blue carbon framework, despite participating in the global carbon cycle. For example, coral
76 reefs and oyster reefs dominated habitats contribute to climate change adaptation through
77 energy dissipation and contribution to sediments, but not through greenhouse gas mitigation,
78 as the process of calcification releases CO₂ (Lovelock and Duarte, 2019). On the other hand,
79 pelagic ecosystems, like marine fauna and phytoplankton, have been suggested to be included
80 as blue carbon ecosystems; however, their contribution to climate change mitigation through
81 long-term carbon preservation is still debatable (Lovelock and Duarte, 2019). In the present
82 date, specifically, the three coastal ecosystems, namely mangroves, seagrass, and tidal
83 marshes, are ubiquitously established as blue carbon ecosystems because of their high carbon
84 stocks, long-term carbon storage capacity, potential to manage greenhouse gas emissions, and
85 other adaptation characteristics (Kuwae and Hori, 2019; Macriedie et al. 2019; Chen and
86 Xue, 2020). Owing to their potential in offsetting the atmospheric carbon imbalance, several
87 authors have advocated that assessment of the carbon stocks in these ecosystems at a
88 nationwide scale is beneficial for both formulating adequate conservation and adaptation
89 strategies as well as facilitating their proper inclusion in the carbon-financing network
90 (Taillardat et al., 2018; Gallagher, 2017).

91 India, in this regard, has a long coastline of > 7500 km and shelters all these three
92 blue carbon ecosystems (Kathiresan, 2018; Jayanthi et al., 2018; Thangaradjou and Bhatt,
93 2018; Viswanathan et al., 2020). The mangrove ecosystem is the principal and relatively
94 well-studied blue carbon ecosystem in India. The mangroves in India cover an area of 4991

95 km² (Indian State of Forest Report, 2021). The east coast and west coast harbor 60 % and 14
96 % of these mangroves, respectively. The Andaman and Nicobar Islands shelter the remaining
97 26%. India and Bangladesh share the Sundarban mangroves, the world's largest single tract of
98 mangrove forest. Sundarban (2112 km²) followed by Bhitarkanika (130 km²) and Pichavaram
99 (10 km²) are some of the prominent mangrove forests of India (Dasgupta and Shaw 2013).

100 Seagrass meadows in India thrive in both shallow coastal water and offshore islands.
101 India, at present, shelters approximately 517 km² seagrass cover (Geevarghese et al., 2016).
102 The east coast of India shelters the majority of seagrasses in Palk Bay (330 km²), Gulf of
103 Mannar (69 km²), and Chilika Lagoon (85 km²), with smaller patches observed on the west
104 coast in the Gulf of Kutch, Gujarat, and in the lagoons of Lakshadweep in the Arabian Sea,
105 and the Andaman and Nicobar waters in the Bay of Bengal (Ganguly et al., 2018). Fourteen
106 species of seagrass have been identified from the Indian coast, belonging to seven genera, of
107 which Palk Bay has the highest seagrass species diversity (Patro et al., 2017). Thangaradjou
108 and Bhatt (2018) reported sixteen seagrass species with an approximate cover of 500 km² at
109 isolated locations along the coast, lagoons, backwaters, and estuaries. In India, salt marshes
110 are distributed in seven coastal districts/union territories, viz. Gujarat, Daman and Diu,
111 Maharashtra, Tamil Nadu, Puducherry, Andhra Pradesh, and Andaman & Nicobar Islands,
112 covering an approximate area of 1600 km² (Banerjee et al. 2017).

113 In the present study, we summarized state-of-the-art data on the blue carbon
114 repository of India (Fig. 1) and estimated the social cost of this blue carbon, i.e. the cost that
115 the global environment has to bear for a unit carbon emission towards the atmosphere
116 (Bertram et al., 2021). We considered three ecosystems, i.e., mangroves, seagrass, and salt
117 marshes, as blue carbon ecosystems because only the scientific community unanimously
118 accepts these three as blue carbon ecosystems (Lovelock and Duarte, 2019). From the
119 available published peer-reviewed literature, we further investigated the dominant species

120 and genera contributing to the blue carbon stock of India. We also characterized the natural
121 and anthropogenic threats that these ecosystems face at present.

122 **2. Methodology**

123 **2.1 Review and search strategy**

124 We used Google search engine and literature databases ‘Google Scholar’, ‘PubMed’, ‘Web of
125 Science’ and ‘ScienceDirect’. We searched for the publications using key terms like ‘blue
126 carbon’, ‘mangrove’, ‘aboveground biomass’, ‘belowground biomass’, ‘mangrove soils’,
127 ‘carbon stock’, ‘seagrass’, ‘saltmarsh’, ‘tidal marsh’. We searched these terms in combination
128 with the name of the country ‘India’ and its coastal states namely ‘West Bengal’, ‘Odisha’,
129 ‘Andhra Pradesh’, ‘Tamil Nadu’, ‘Kerala’, ‘Maharashtra’, ‘Goa’, and ‘Gujarat’, and the ‘Bay
130 of Bengal’ and the ‘Arabian Sea’. We also searched using the names of the coastal union
131 territories of India, namely ‘Lakshadweep’, ‘Daman and Diu’, ‘Puducherry’, and the
132 ‘Andaman and Nicobar Islands’. We downloaded the published papers utilizing institutional
133 access to ‘ScienceDirect’ and the ‘Google Scholar’ search engine. The present review and
134 subsequent data synthesis followed the Preferred Reporting Items for Systematic Reviews
135 and Meta-Analyses (PRISMA) guidelines (Moher et al., 2009).

136 **2.2 Selection criteria of papers**

137 Based on primary screening, 235 potentially eligible research articles were shortlisted.
138 Followed by this, the full text of those articles was downloaded. Research articles with
139 unavailable full-text were excluded from this review. Research articles published only in the
140 English language were considered in this study. Selection criteria of the papers were i)
141 articles that are original work and not review papers, ii) articles for which full-text is
142 available on the internet, iii) articles published anytime i.e. no timeframe was set, and iv)

143 articles that had the data of at least one of the three blue carbon ecosystems (mangrove,
144 seagrass, and saltmarsh or tidal marsh) considered in this study.

145 **2.3 Data extraction and compilation**

146 Only the mean data of blue carbon stock per unit area reported from any region within India
147 were considered in this study. Following the above-mentioned searching strategy and
148 selection criteria, we considered 35 published papers and edited book chapters where the
149 authors have quantified estimates of blue carbon from the different locations in India. We
150 converted all data of the blue carbon repository to a single unit, Mg C ha⁻¹.

151 While collating the data, we observed that many locations were sampled more than once and
152 have multiple results. At the same time, not all of these studies took into account all the
153 compartments (like aboveground carbon (AGC), belowground carbon (BGC), and soil
154 carbon) to quantify the total blue carbon stock for the entire region. Under such
155 circumstances, we calculated the mean of blue carbon stock per unit area (from all the
156 available studies on that site) in each of the compartments of that ecosystem and multiplied it
157 with the total area of that site. In this way, we computed the total blue carbon in each
158 compartment and added the resultants to derive the total blue carbon stock at each site. Many
159 of the sites in the case of all three blue carbon ecosystems were not sampled at all. To derive
160 an approximate estimate of the nationwide blue carbon stock for India, the existing data (on
161 region-specific total carbon stock) for each of the ecosystems were upscaled by assuming that
162 the already measured carbon content per unit area holds for the entire coverage of that
163 ecosystem in the country. In other words, the average blue carbon stock per unit area of a
164 particular blue carbon ecosystem (computed from the existing data) was multiplied to the
165 spatial extent of the area that underwent no sampling to date. Based on the number of studies
166 conducted and the area of each blue carbon ecosystem covered throughout India, the

167 uncertainty in nationwide blue carbon estimation was gauged. The parameter of whether all
168 the blue carbon compartments were considered or not also enabled us to develop an idea of
169 the uncertain aspects of the blue carbon estimation that requires attention in the future.

170 **3. Results and Discussion**

171 **3.1 The variability of blue carbon stock in the Indian mangrove ecosystem**

172 Published literature suggests that the aboveground biomass contributes most to the
173 blue carbon stock, followed by soil organic carbon and belowground biomass (Table 1).
174 Bhitarkanika mangrove forest has the maximum range of aboveground and belowground
175 biomass carbon (55 to 720 Mg C ha⁻¹ and 23 to 220 Mg C ha⁻¹) amongst all mangrove forests
176 of India (Rasquinha and Mishra 2020) studied so far. Though Sundarban is the largest
177 mangrove forest in the world (and India), the carbon stock of Sundarban per unit area is
178 lesser than Bhitarkanika, Pichavaram mangrove, and Vypin Cochin mangrove region **but**
179 **comparable with that observed in the Gulf of Kachchh** (Table 1). Ray et al. (2011) stated that
180 the carbon stock of Sundarban is lower than a typical terrestrial tropical forest. They also
181 depicted that the soil pH is the principal (35.2%) contributing factor in mangrove biomass as
182 well as carbon stock, followed by soil salinity (26.2%) and nutrients (15.2%, total inorganic
183 nitrogen, and total extractable phosphorus). However, Banerjee et al. (2013) and Mitra et al.
184 (2011) reported that the relatively higher salinity in the central region of Sundarban caused
185 the subsequent lowering of biomass and hence, the carbon content in the mangrove floras of
186 this region. Chowdhury et al. (2019) indicated that freshwater scarcity due to reduced flow
187 from upper reaches has led to a significantly elevated degree of salinization in the mangrove
188 soils of Sundarban. The reduced riverine flow decreased the nutrient replenishment and that,
189 in turn, has taken a heavy toll on the proper physiological functioning of several mangrove
190 species. However, to combat the ongoing degradation of the Sundarban mangroves (Indian

191 part), several restoration and conservation endeavors have been undertaken in the past two to
192 three decades that include afforestation of around 17,000 ha of mangrove habitat maintaining
193 the intrinsic species diversity (Vyas and Sengupta 2012).

194 Rasquinha and Mishra (2020) emphasized the impact of harvesting mangroves on the
195 aboveground and belowground biomass carbon stock. They observed that the local people in
196 Bhitarkainika harvest mangroves for fuel demands. Integrated management plans to
197 accommodate the local cultural and economic needs have become a dire need to prevent such
198 indiscriminate harvesting. Bhomia et al. (2016) stated that harvesting pressure and changes in
199 land use (like conversion of mangrove habitats to aquaculture) adversely affect the carbon
200 stock of mangroves. They suggested proper management practices by enabling these
201 communities to avail themselves of carbon offset/conservation payments under approved
202 climate change mitigation strategies and actions.

203 Joshi and Ghose (2014) attributed tidal inundation level as a crucial regulatory factor for
204 mangrove biomass while working in the Lothian Island of Sundarban. They also stated that
205 freshwater turnover is a governing factor for mangrove growth. They further observed that
206 low salinity regimes are principally responsible for enhanced species diversity and mixed
207 vegetation community, which leads to higher blue carbon stocks. Sahu et al. (2016) observed
208 that available nitrogen showed a negative correlation with total plant biomass while working
209 on the Mahanadi mangroves.

210 The study of the species-specific biomass of different mangrove species showed
211 marked variability amongst the biomass carbon stock (Table 2). In Sundarbans, *Avicennia*
212 *marina* showed the highest above ground carbon (AGC) per unit area, followed by *A. alba*
213 and *A. officinalis* (Ray et al. 2011; Banerjee et al. 2013; Raha et al. 2013). In Bhitarkanika, *A.*
214 *officinalis* was the highest contributor to AGC, followed by *Heritiera littoralis* and *H. fomes*

215 (Anand et al. 2020; Banerjee et al. 2020). In contrast to the mangrove forests of the east
216 coast, *Rhizophora mucronata* showed the highest AGC in the Vypin-Cochin region on the
217 west coast of India (ShyleshChandran et al. 2020). However, in the Gulf of Kachchh,
218 *Avicennia marina* exhibited a wide range of AGC (Thivakaran et al. 2020).

219 The biomass of different mangroves species showed diverse dynamics with salinity. Banerjee
220 et al. (2013) reported a significant negative correlation of soil salinity with the biomass
221 carbon stock of *Sonneratia apetala*, whereas a significant positive correlation with *A. alba*
222 and *Excoecaria agallocha*. They stated that a negative correlation between *S. apetala*
223 biomass and salinity reflects the sensitivity of this species to ambient salinity, and a
224 significant positive correlation between salinity and other two species signifies high salinity
225 tolerance of those species (*A. alba* and *E. agallocha*). Banerjee et al. (2012) depicted an
226 increase in mangrove soil organic carbon and soil pH in the western part of Indian
227 Sundarbans. Mitra et al. (2012) also reported that the mangrove forest in the eastern Indian
228 Sundarbans exhibits comparatively lower organic carbon density, and the organic carbon and
229 carbon density decreased with depth. Banerjee et al. (2020), while working in the
230 Bhitarkanika mangroves, reported that with increasing carbon load in the soil, the growth of
231 *A. marina* decreases, whereas *A. officinalis* and *X. granatum* exhibited an opposite trend. In
232 this regard, Sandilyan et al. (2010) stated that the salinity rise has accentuated and affected
233 almost all the mangrove habitats of India. They indicated that the change in the salinity
234 region not only compromises the species diversity but also disrupts the biological
235 productivity of this crucial ecosystem. In the highly saline Gulf of Kachchh region,
236 *Rhizophora mucronata* and *Ceriops tagal* exhibited very low AGC; however, the AGC in
237 *Avicennia marina* stands was significantly high in several sites (Thivakaran et al. 2020),
238 indicating that *A. marina* acclimatizes well in the polyhaline circumstances and stores a
239 substantial quantity of carbon.

240 Though the blue carbon ecosystems of India store substantial quantities of carbon; these
241 ecosystems are under severe threats of varying types. Regional sea-level rise, coastal erosion,
242 anthropogenic conversion to aquaculture plots, recurrent tropical cyclones are some of the
243 potential hazards that take a heavy toll on the mangroves of India (Chaudhuri et al., 2015).
244 Unlike the mangroves, the seagrasses and salt marshes of India have received much less
245 attention from the perspective of assessing the threats. However, Ganguly et al. (2017) and
246 Banerjee et al. (2018) pointed out that fishing operations with mechanically operated boats
247 and trap nets, tourism activities, and discharge of untreated nutrient-rich discharges from
248 industries and aquaculture ponds can significantly deteriorate the functioning of the
249 seagrasses in India. Patro et al. (2017) mentioned a near absence of threat assessment in the
250 Indian salt marshes; however, they pointed out that the salt marshes are also susceptible to
251 coastal eutrophication, pollution, and overgrazing. Moreover, the perception of salt marshes
252 as wastelands prevails in many parts of India, which leads to neglect not knowing their
253 potential in sequestering carbon and offering various ecosystem services.

254 **3.2 Carbon stock in the Indian seagrasses and salt marshes**

255 The available literature suggests that amongst the seagrass distribution in India, the biomass
256 and carbon stock data are available only from the Palk Bay and Chilka Lagoon. Unlike
257 mangroves, the aboveground biomass carbon of the seagrass was significantly lower than the
258 belowground biomass. Amongst the 14 species of seagrass reported from the Palk Bay,
259 *Cymodocea serrulata* and *Syringodium isoetifolium* are the most dominant (Govindasamy
260 and Arulpriya 2011; Ganguly et al. 2017). Ganguly et al. (2017) reported the total organic
261 carbon stored as seagrass biomass in the Palk Bay ($0.94 \text{ Mg C ha}^{-1}$). They also observed that
262 the partial pressure of CO_2 in water and photosynthetically active radiation best explained the
263 variability of net community production (NCP) of seagrass meadows in the Palk Bay.
264 Ganguly et al. (2018) reported the aboveground biomass was between 0.20 and 0.96 Mg C

265 ha⁻¹, whereas the belowground biomass carbon ranged between 0.31 and 2.93 Mg C ha⁻¹
266 while working in both Palk Bay and Chilika Lagoon. Organic carbon in the top 1 m soil in
267 the seagrass meadows ranged between 107 and 143 Mg C ha⁻¹ (Ganguly et al. 2018).
268 Ganguly et al. (2018) also reported that *Cymodocea* sp. showed the highest biomass amongst
269 the five species studied. They recorded the highest recorded NCP from the mixed bed of
270 *Thalassia hemprichi* and *Syringodium isoetifolium*. They also attributed biomass density,
271 changes in species compositions, and richness to the long-term burial and storage of organic
272 carbon.

273 We found only four studies on the salt marsh carbon stocks from India. Kaviarasan et
274 al. (2019) reported the biomass and carbon stock of four salt marsh species, *Suaeda maritima*,
275 *Sesuvium portulacastrum*, *Arthrocnemum indicum*, and *Salicornia brachiata* from Tuticorin,
276 southeast coast of India. They reported the highest AGB in *A. indicum* in both dry and wet
277 seasons (10.91 ± 0.15 g cm⁻² and 14.87 ± 0.68 g cm⁻²), whereas they reported the highest
278 BGB in *S. maritima* in both seasons, as 2.01 ± 0.35 g cm⁻² and 4.49 ± 0.35 g cm⁻²,
279 respectively. They also reported sediment organic carbon stocks, which ranged between 8.42
280 ± 0.64 to 54.46 ± 1.46 Mg C ha⁻¹. Chaudhary et al. (2018) observed that the salt marsh
281 species *Salicornia brachiata* dominates the Gujarat coastlines. They observed that the AGC
282 varied between 0.05 and 2.21 Mg C ha⁻¹, whereas the BGC ranged from 0.009 to 0.335 Mg C
283 ha⁻¹. These observations were in parity with Rathore et al. (2016). They also worked in the
284 coastal regions of Gujarat and reported a soil organic carbon stock varying between 4.5 and
285 8.2 Mg C ha⁻¹. Das et al. (2015), while working on the salt marshes in and around the
286 Sundarban region, observed much lesser AGC and BGC compared to what was reported by
287 Rathore et al. (2016), Chaudhary et al. (2018), and Kaviarasan et al. (2019). However, they
288 observed a moderate range of 9.4 to 13.4 Mg C ha⁻¹ in the salt marsh soils of the Sundarban.

289 **3.3 An up-scaled estimate of total blue carbon stock in India**

290 While several studies attempted to quantify the region-specific blue carbon stock within
291 India; very few of these derived an area-integrated total estimate of blue carbon. According to
292 the latest Indian State of Forest Report (2021), twelve states (including the union territories)
293 have mangroves in their coastlines with a total area of 4991 km². Out of these states, five
294 states and one union territory, namely West Bengal (2112 km²), followed by Gujarat (1117
295 km²), Andaman and Nicobar Islands (616 km²), Andhra Pradesh (404 km²), Maharashtra
296 (320 km²), and Odisha (251 km²) comprise 97% of the country's mangrove cover. Table 1
297 shows that Gujarat, Andhra Pradesh, and Maharashtra, despite covering 1841 km² of
298 mangroves, **have only one study on carbon stock assessment and that too only from Gujarat,**
299 **where the AGC was measured across several sites (Thivakaran et al. 2020).** Ray et al. (2011),
300 while working on the Sundarbans, derived that the total carbon stock in the aboveground and
301 belowground live biomass accounts for 21.13 Tg C, whereas the soil carbon pool (up to 30
302 cm depth) holds as much as 5.49 Tg C, thus, comprising a total of 26.62 Tg C (in 2112 km²).
303 The composite mean AGC, BGC, soil carbon pool in the mangroves of Odisha
304 (Bhiatarkanika and Mahanadi) per unit area was close to 145 Mg C ha⁻¹, 46.7 Mg C ha⁻¹, and
305 47 Mg C ha⁻¹, respectively. Multiplication of these estimates with the total mangrove cover
306 of Odisha (25100 hectares) shows that the state holds as much as 6 Tg C (in 251 km²).
307 Similarly, the soil and total biomass carbon observed in Pichavaram mangroves and Vypin-
308 Cochin mangroves indicate that Tamil Nadu and Karnataka can store almost 0.003 Tg C (in
309 45 km²) and 0.002 Tg C (in 9 km²). Besides these states, the carbon content in all the
310 mangrove compartments is not available. Thus, we upscaled the estimates observed for the
311 2417 km² area (comprising West Bengal, Odisha, Tamil Nadu, and Karnataka) for the entire
312 country's mangrove cover of 4991 km² following a simple unitary method. This exercise
313 shows that Indian mangrove forests can store almost 67 Tg C of blue carbon in the live
314 biomass and soils (up to 30 cm depth).

315 Contrary to the mangroves, the seagrass ecosystem in India covers only 517 km² as per the
316 latest estimates (Geevarghese et al. 2016), out of which 93% occurs in the Palk Bay, Gulf of
317 Mannar, and Chilika Lagoon. Published records of carbon stock exist for all these three
318 places in the recent past. Ganguly et al. (2017; 2018) observed that Palk Bay of Chilika
319 lagoon stores almost 0.047 Tg C and 0.0098 Tg C, respectively. Recently, Kaladharan et al.
320 (2020) measured the total blue carbon stock in the Gulf of Mannar and Palk Bay. Their
321 observations for Palk Bay are in good agreement (0.043 Tg C) with Ganguly et al. (2017).
322 Kaladharan et al. (2020) reported that the Gulf of Mannar stores almost 0.0018 Tg C. Thus,
323 93% of the seagrass stand in India together holds ~0.0586 Tg C. Extrapolating this estimate
324 for the rest of the unsampled 7 %, the figure comes to around 0.0630 Tg C.

325 Unlike the seagrass ecosystems, there are contradicting observations about the total
326 extent of salt marsh cover in the country. A recent study by Viswanathan et al. (2020)
327 challenged the earlier estimates of salt marsh cover by Garg et al. 1998 (reported 1698 km²
328 during the years 1992-93) and SAC (2011) (reported 1611 km² during the years 2007-08) and
329 revised that total salt marsh cover in India sums up to only 290 km². They argued that the
330 earlier studies considered all the halophytic grasslands in the coastal zones, which led to such
331 overestimates. According to their estimate, out of the seven states where salt marshes thrive,
332 Gujarat encompasses almost 158 km² followed by Tamil Nadu (58 km²) and West Bengal (30
333 km²), and these three states together comprise 85% of the country's total salt marsh cover.
334 The carbon stock assessment has been already carried out in these three states. Table 3 shows
335 that based on the estimates of Rathore et al. (2016) and Chaudhary et al. (2018), Gujarat can
336 store 0.0012 Tg C in the salt marsh ecosystems. Similarly, the estimates of Kaviarasan et al.
337 (2019) and Das et al. (2015) indicate that the salt marshes of Tamil Nadu and West Bengal
338 can store 0.0026 Tg C and 0.00036 Tg C, respectively. Thus, 85% of the country's salt marsh
339 area stores almost 0.00416 Tg C. Extrapolating this estimate for the rest of the 15%

340 unsampled area, Indian in totality stores around 0.0049 Tg C. However, considering the
341 spatial extent of 1698 km² (as reported by Garg et al. 1998), the carbon stocks can be as high
342 as 0.0286 Tg C (Fig. 2).

343 **3.4 Monetary worth of India's blue carbon stock**

344 The present meta-analysis of the existing data shows that the carbon content in Indian
345 mangroves (67 Tg C) differs by many orders of magnitudes than that observed in the
346 seagrasses (0.0630 Tg C) and salt marshes (0.0049 Tg C). Murray et al. (2011) mentioned
347 that mangroves (18,000 US \$ ha⁻¹) have a much greater blue carbon value than salt marshes
348 and seagrasses (8,000 US \$ ha⁻¹) in terms of the benefits that they provide to humankind.
349 These estimates show that the mangroves of India are worth 8995 million US \$, and the
350 seagrasses and salt marshes are worth 414 million US \$ and 232 million US \$, respectively.
351 Thus, the cost of the total blue carbon stock of India in terms of areal cover amounts to 9.6
352 billion US \$. However, Pendleton et al. (2012) emphasized that the global environment has to
353 bear a mean cost of 41 US \$ (as per the rate of US \$ in the year 2007) per ton of CO₂
354 emission due to the damage posed by an additional load of one ton of new carbon. The global
355 scientific community refers to such cost as the social cost of carbon, and the estimates can
356 vary from 7 US \$ to 81 US \$ per ton (Pendleton et al. 2012). According to these estimates
357 (i.e., by multiplying the range of social cost of carbon mentioned by Pendleton et al. (2012)
358 with the estimated carbon stock derived in this study), the social cost of carbon trapped in the
359 Indian mangroves is worth 469 million US \$ to 5427 million US \$ (mean: 2747 million US
360 \$). In the same line, the seagrasses and the salt marshes of India store carbon worth 0.441
361 million US \$ to 5.103 million US \$ (mean 2.583 million US \$) and 0.034 million US \$ to
362 0.397 million US \$ (mean: 0.201 million US \$), respectively. According to this approach, the
363 cost of total blue carbon in India varies between 0.47 billion US \$ to 5.43 billion US \$ (mean
364 2.75 billion US \$). Pendleton et al. (2012) asserted that due to deforestation and degradation

365 of global blue carbon, environmental damage worth 6 to 42 billion US \$ occurs every year.
366 The present estimates show that the higher margin of India's blue carbon worth is almost
367 equivalent to the minimum amount of carbon lost every year globally.

368 **3.6 Uncertainties in blue carbon estimation**

369 Compared to the stretch and diversity of the blue carbon habitats and species composition,
370 many of the states in India remain under-sampled or completely data-deficient. The present
371 study indicated that mangroves in India could store much higher quantities of carbon than the
372 other two ecosystems; however, as many as eight states (including the union territories) have
373 no proper estimate of blue carbon stock in their mangrove stands. Thus, the present
374 assessment derived on extrapolation could be both under- or over-estimates, as the present
375 study indicated that 51.5%, 7%, and 15% of the country's mangrove, seagrass, and saltmarsh
376 cover, respectively, remain to be sampled from the perspective of characterizing the blue
377 carbon content. Moreover, many of the studies conducted localized sampling. A more holistic
378 approach with an increased number of sampling points can enhance the confidence level of
379 the data and minimize the uncertainties.

380 Very few of the studies considered in this paper carried out seasonal or inter-annual
381 sampling. One-time sampling can potentially contribute to temporal bias. In underwater
382 vegetations like seagrasses and salt marshes, the live biomass exhibits significant intra-annual
383 variability (Chaudhary et al., 2018). Thus, future studies should focus on year-long sampling
384 strategies and, if possible, for multiple years to better understand the standing carbon stock
385 dynamics. More ground-based studies and quadrat analyses should resolve the controversies
386 and contradictions in areal coverage, as observed in salt marshes. Besides only the spatial
387 extent, characterizing the health of vegetations like mangroves plays a crucial role in
388 assessing their biomass (Pandey et al., 2019). Lately, Bertram et al. (2021) assessed the

389 global blue carbon stock in a country-wise manner. They mentioned that the spatial extent
390 and the factors governing the carbon stocks in the blue carbon habitats are the two most
391 potent aspects that add to the uncertainty of the entire estimation protocol.

392 Among the total carbon stock in the three blue carbon ecosystems estimated in this study,
393 mangroves have the highest uncertainty as only 48.5 % of the country's mangrove cover
394 remains sampled in this regard. Sampling in seagrasses and salt marshes covered almost 93%
395 and 85% of the total area in India, respectively. Thus, the uncertainties in estimates of these
396 two ecosystems should be much less than the mangroves. However, if future studies indicate
397 that the salt marsh cover in India is beyond the assessment of Viswanathan et al. (2020) (290
398 km²), then the degree of uncertainty would substantially increase, as there is a complete
399 absence of carbon stock data from the salt marsh cover of many states. Mathur et al. (2020)
400 prepared a policy brief on the Indian blue carbon habitats and their carbon sequestration
401 potential. They considered a single mean value for the carbon stock per unit area in
402 mangroves of India (368 Mg C ha⁻¹) and estimated a total CO₂ sequestration potential of
403 702.42 Tg CO₂ for the entire country. However, considering the spatial variability reported in
404 the existing scientific literature base, we estimated that the total carbon content in the Indian
405 mangroves (67 Tg C \approx 245.7 Tg CO₂ equivalent) is much less than that approximated by
406 Mathur et al. (2020).

407 **4. Conclusion**

408 Collating all the data acquired so far on blue carbon stock in India, this study infers that India
409 could collectively store around 67.35 Tg C within the mangrove, seagrass, and salt marsh
410 habitats. This measurement is almost 2.85 times lower than the other holistic estimates from
411 India. The present study also shows that almost 99.5 % of this carbon stock is within the
412 mangrove ecosystems. These observation advocates that mangrove restoration and

413 conservation should receive priority. Conserving or enhancing the extent of mangroves can
414 enable India to play a leading role in combatting global climate change due to anthropogenic
415 carbon emissions. Effective restoration and afforestation endeavors should be taken up
416 throughout the nation to enhance the blue carbon stock in all the conventional ecosystems.
417 According to the latest estimates, mangroves encompass 4991 km² in India, and seagrass
418 cover extends for around 517 km². However, there lies considerable uncertainty regarding the
419 spatial extent of salt marshes in India. Many of the mangrove forests (or patches) in India are
420 yet to receive any attention for measuring their carbon stock. However, adequate
421 measurements exist for most of the prominent seagrass and salt marsh patches of this country.
422 Biased sampling in some of the mangroves and the complete absence of data in the others
423 could result in substantial underestimation or overestimation. Despite so many factors, the
424 authors of this paper believe that the estimates derived in this study are close to ground
425 reality. Future studies should orient accordingly to minimize the uncertainties and derive an
426 estimate of blue carbon stock with more confidence, as payments and returns against the
427 amount of carbon sequestered would play a crucial role in governing the blue economy of the
428 near future.

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661 **Figure captions**

662 **Fig. 1** The study area map showing the prime locations of the blue carbon ecosystems along
 663 the Indian coastline. The mangrove, seagrass, and salt marsh sites, shown in the figure are
 664 based on available pieces of literature considered in the present study.

665 **Fig. 2** A snapshot of data on blue carbon repository in India

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673 **Table 1** The collation of data on the carbon stock in the aboveground biomass (AGB),
 674 belowground biomass (BGB), total biomass (TB), and the topsoil (up to 30 cm depth)
 675 observed in the various mangrove forests of India. As per availability, the data is reported as
 676 either mean \pm standard deviation (from the mean) or range (minimum to maximum).

Place	AGB carbon (Mg C ha ⁻¹)	BGB carbon (Mg C ha ⁻¹)	TB carbon (Mg C ha ⁻¹)	Soil carbon (Mg C ha ⁻¹)	References
Sundarbans (WB)	39.9 \pm 14.1	9.6 \pm 3.4		17.4 \pm 2.3	Ray et al. (2011)
Sundarbans (WB)	8.9 to 50.9			2.9 to 25.9	Joshi and Ghose (2014)

Sundarbans (WB)				28.5 ± 2.0	Banerjee et al. (2012)
Sundarbans (WB)				20.4 ± 5.6	Mitra et al. (2012)
Bhitarkanika (OD)				16.4 ± 2.6	Banerjee et al. (2018)
Bhitarkanika (OD)	278.9				Pandey et al. (2019)
Bhitarkanika (OD)	179.7 ± 67.0 to 196.1 ± 182.1	64.0 ± 55.0 to 68.3 ± 17.1		15.63 ± 7.22 to 16.49 ± 6.59	Rasquinha and Mishra (2020)
Bhitarkanika (OD)			7 ± 4 to 100 ± 11	92 ± 20 to 177 ± 14	Bhomia et al. (2016)
Bhitarkanika (OD)			131.06 ± 11.08		Anand et al. (2020)
Bhitarkanika (OD)	143.6 ± 38.0			5.5 ± 1.7	Banerjee et al. (2020)
Bhubaneswar and Bhitarkanika (OD)				54.3 ± 3.0	Pattnayak et al. (2019)
Mahanadi (OD)	93.2 ± 21.6			5.9 ± 1.5	Banerjee et al. (2020)
Mahanadi (OD)	62.5 ± 6.4 to 62.8 ± 11.3	26.7 ± 2.6 to 27.9 ± 4.9	89.1 ± 8.9 to 90.6 ± 16.2	54.3 ± 3.0 to 60.9 ± 5.6	Sahu et al. (2016)
Chandaka Wildlife Sanctuary (OD)				47.51 ± 2.16	Pattnayak et al. (2019)
Mandovi-Zuari (GA)				0.6 to 198.7	Krishnan and Bharathi (2009)
Mandovi-Zuari (GA)				6.9 to 38.6	Shynu et al. (2015)
Karankadu, Palk Bay (TN)	29.8 ± 12.7				Prasanna et al. (2017)
Pichavaram (TN)				34.6 ± 15.0	Ranjan et al. (2011)
Pichavaram (TN)				19.1 to 48.7	Gnanamoorthy et al. (2019)
Northern Kerala (KL)				17.3 to 259.8	Resmi et al. (2016)
Cochin Metropolis (KL)				94.8 to 253.2	Sebastian and Chacko (2006)
Vypin-Cochin Region (KL)	54.3 ± 36			18.3 ± 3.9	ShyleshChandran et al. (2020)
Kerala (KL)			58.6 ± 0.5	81.3 ± 10.2	Harishma et al. (2020)

South Andaman (AN)				34.8 ± 6.3 to 51.8 ± 8.6	Dinesh et al. (2004)
Gulf of Kachchh (GJ)	36.4				Thivakaran et al. (2020)
WB – West Bengal; OD – Odisha; GA – Goa; TN – Tamil Nadu; KL – Kerala; AN – Andaman and Nicobar Islands; GJ – Gujarat					

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696 **Table 2** The collation of data on the mangrove species-specific carbon stock in the
697 aboveground biomass (AGB) or total biomass (TB) observed in the various mangrove forests
698 of India. As per availability, the data is reported as either mean ± standard deviation (from the
699 mean) or range (minimum to maximum). Standalone single values are mean without any
700 standard deviation.

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Mangrove Species	Biomass carbon (Mg C ha ⁻¹)	Place	Reference
<i>Acanthus ilicifolius</i>	0.002 to 0.02 (AGB)	Ernakulam (KL)	ShyleshChandran et al. (2020)
<i>Aegiceras corniculatum</i>	35.2 (AGB)	Sundarban (WB)	Joshi and Ghose (2014)
<i>Aegiceras corniculatum</i>	165.03 ± 10.87 (AGB)	Bhitarkanika (OD)	Anand et al. (2020)
<i>Aegiceras corniculatum</i>	108.77 ± 13.67 (AGB)	Bhitarkanika (OD)	Banerjee et al. (2020)
<i>Agialitis rotundifolia</i>	54.77 to 64.03 (TB)	Sundarban (WB)	Ray et al. (2011)
<i>Avicennia alba</i>	0.5 to 27.28 (TB)	Sundarban (WB)	Ray et al. (2011)

<i>Avicennia alba</i>	6.05 ± 1.19 to 11.02 ± 1.42 (AGB)	Sundarban (WB)	Mitra et al. (2011)
<i>Avicennia alba</i>	41.65 to 64.57 (TB)	Sundarban (WB)	Banerjee et al. (2013)
<i>Avicennia alba</i>	32.7 (AGB)	Sundarban (WB)	Joshi and Ghose (2014)
<i>Avicennia alba</i>	24.73 to 29.09 (AGB)	Sundarban (WB)	Raha et al. (2013)
<i>Avicennia alba</i>	0.5 to 27.28 (TB)	Sundarban (WB)	Ray et al. (2011)
<i>Avicennia alba</i>	20.22 (AGB)	Sundarban (WB)	Saha et al. (2019)
<i>Avicennia marina</i>	8.80 to 38.63 (TB)	Sundarban (WB)	Ray et al. (2011)
<i>Avicennia marina</i>	22.84 to 26.92 (AGB)	Sundarban (WB)	Raha et al. (2013)
<i>Avicennia marina</i>	21.14 (AGB)	Sundarban (WB)	Saha et al. (2019)
<i>Avicennia marina</i>	7.63 ± 1.08 to 35.65 ± 2.63 (AGB)	Bhitarkanika-Mahanadi (OD)	Banerjee et al. (2020)
<i>Avicennia marina</i>	67.47 ± 20.09 (TB)	Vellar-Coleroon (TN)	Kathiresan et al. (2013)
<i>Avicennia marina</i>	5.05 to 108.40 (AGB)	Gulf of Kachchh (GJ)	Thivakaran et al. (2020)
<i>Avicennia officinalis</i>	0.56 to 20.68 (TB)	Sundarban (WB)	Ray et al. (2011)
<i>Avicennia officinalis</i>	25.80 to 29.75 (AGB)	Sundarban (WB)	Raha et al. (2013)
<i>Avicennia officinalis</i>	6.70 (AGB)	Sundarban (WB)	Saha et al. (2019)
<i>Avicennia officinalis</i>	1.73 ± 0.01 to 280.83 ± 21.29 (AGB)	Bhitarkanika-Mahanadi (OD)	Banerjee et al. (2020)
<i>Avicennia officinalis</i>	1.38 to 36.16 (AGB)	Ernakulam (KL)	ShyleshChandran et al. (2020)
<i>Bruguiera cylindrica</i>	0.19 (AGB)	Ernakulam (KL)	ShyleshChandran et al. (2020)
<i>Ceriops sp.</i>	2.01 to 10.40 (TB)	Sundarban (WB)	Ray et al. (2011)
<i>Ceriops tagal</i>	0.18 to 2.26 (AGB)	Gulf of Kachchh (GJ)	Thivakaran et al. (2020)
<i>Excoecaria agallocha</i>	8.83 to 10.07 (TB)	Sundarban (WB)	Ray et al. (2011)
<i>Excoecaria agallocha</i>	14.93 ± 0.63 to 24.98 ± 1.02 (AGB)	Sundarban (WB)	Mitra et al. (2011)
<i>Excoecaria agallocha</i>	13.89 to 24.24 (TB)	Sundarban (WB)	Banerjee et al. (2013)
<i>Excoecaria agallocha</i>	10.20 to 11.92 (AGB)	Sundarban (WB)	Raha et al. (2013)
<i>Excoecaria agallocha</i>	5.83 (AGB)	Sundarban (WB)	Saha et al. (2019)
<i>Excoecaria agallocha</i>	6.05 ± 1.19 to 11.02 ± 1.42 (AGB)	Bhitarkanika (OR)	Anand et al. (2020)
<i>Excoecaria agallocha</i>	1.64 ± 0.41 to 44.95 ± 2.53 (AGB)	Bhitarkanika-Mahanadi (OD)	Banerjee et al. (2020)

<i>Heritiera fomes</i>	135.20 ± 6.02 (AGB)	Bhitarkanika (OD)	Anand et al. (2020)
<i>Heritiera littoralis</i>	159.39 ± 11.21 (AGB)	Bhitarkanika (OD)	Anand et al. (2020)
<i>Rhizophora mucronata</i>	3.44 ± 1.45 to 114.05 ± 10.29 (AGB)	Bhitarkanika-Mahanadi (OD)	Banerjee et al. (2020)
<i>Rhizophora mucronata</i>	38.05 ± 9.53 (TB)	Vellar-Coleroon (TN)	Kathiresan et al. (2013)
<i>Rhizophora mucronata</i>	2.43 to 105.82 (AGB)	Ernakulam (KL)	ShyleshChandran et al. (2020)
<i>Rhizophora mucronata</i>	0.14 to 0.31 (AGB)	Gulf of Kachchh (GJ)	Thivakaran et al. (2020)
<i>Sonneratia apetala</i>	15.39 ± 1.73 to 84.79 ± 1.22 (AGB)	Sundarban (WB)	Mitra et al. (2011)
<i>Sonneratia apetala</i>	6.77 to 39.10 (TB)	Sundarban (WB)	Banerjee et al. (2013)
<i>Sonneratia apetala</i>	19.79 (AGB)	Sundarban (WB)	Saha et al. (2019)
<i>Sonneratia apetala</i>	21.46 to 30.26 (AGB)	Sundarban (WB)	Raha et al. (2013)
<i>Sonneratia caseolaris</i>	8.001 to 25.88 (AGB)	Ernakulam (KL)	ShyleshChandran et al. (2020)
<i>Xylocarpus granatum</i>	148.43 ± 17.90 (AGB)	Bhitarkanika (OD)	Anand et al. (2020)
<i>Xylocarpus granatum</i>	0.31 ± 0.10 to 3.25 ± 0.31 (AGB)	Bhitarkanika-Mahanadi (OD)	Banerjee et al. (2020)
<i>Xylocarpus mekongensis</i>	76.20 ± 12.36 (AGB)	Bhitarkanika (OD)	Banerjee et al. (2020)

WB – West Bengal; OD – Odisha; TN – Tamil Nadu; KL – Kerala; GJ – Gujarat

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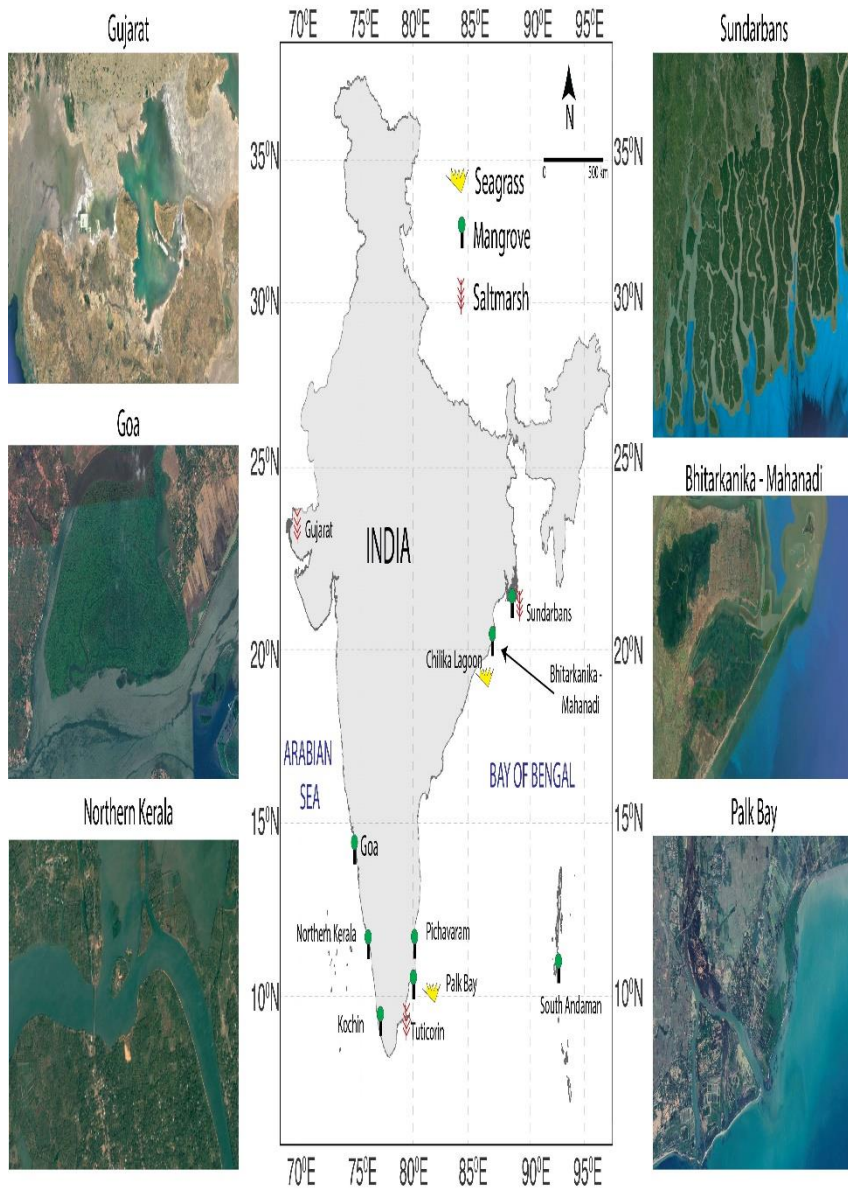
707 **Table 3** The collation of data on the carbon stock in the aboveground biomass (AGB),
 708 belowground biomass (BGB), total biomass (TB), and the topsoil (up to 30 cm depth)
 709 observed in the various seagrass and saltmarsh patches of India. As per availability, the data
 710 is reported as either mean ± standard deviation (from the mean) or range (minimum to
 711 maximum).

Place	AGB carbon (Mg C ha ⁻¹)	BGB carbon (Mg C ha ⁻¹)	Soil carbon (Mg C ha ⁻¹)	References
Palk Bay (TN) (Seagrass)	0.91 ± 0.06 to 0.97 ± 0.10	2.51 ± 0.31 to 2.91 ± 0.26	139.6 (up to 1 m)	Ganguly et al. (2017)

Palk Bay (TN) and Chilika Lagoon (OD) (Seagrass)	0.20 to 0.96	0.30 to 2.90	129 (up to 1 m)	Ganguly et al. (2018)
Tuticorin (TN) (Saltmarshes)	6.32 to 14.87 ± 0.68	1.86 ± 0.21 to 4.49 ± 0.35	8.42 ± .640 to 54.46 ± 1.46 (up to 30 cm)	Kaviarasan et al. (2019)
Gujarat coastline (GJ) (Saltmarshes)	0.05 to 2.1	0.009 to 0.335		Chaudhary et al. (2018)
Gujarat coast (GJ) (Saltmarshes)	0.77 to 1.93		4.5 to 8.2 (up to 30 cm)	Rathore et al. (2016)
Sundarban (WB) (Saltmarshes)	0.64 to 0.71	0.02 to 0.03	9.4 to 13.4 (up to 30 cm)	Das et al. (2015)
WB – West Bengal; OD – Odisha; TN – Tamil Nadu; GJ - Gujarat				

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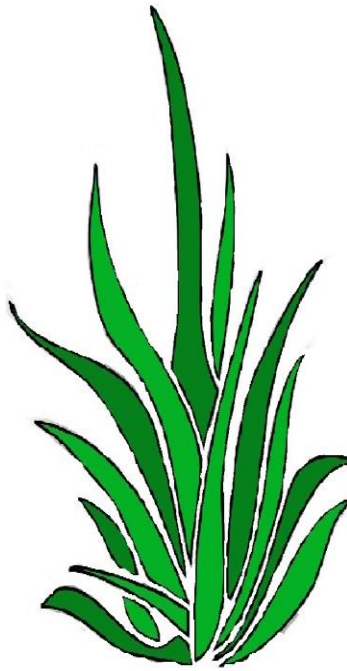


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Spatial extent (km ²)	4795	517	290 - 1698
Species count	46	16	15
Number of studies on blue carbon assessment	24	3	4
Percentage of total areal cover sampled	48.5	93	85
Ecosystem carbon content (Tg C)	67	0.063	0.0049 - 0.0286



Mangroves



Seagrasses



Saltmarshes

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