

MIT Open Access Articles

This is a supplemental file for an item in DSpace@MIT

Item title: Sustainable Lithium Recovery from Hypersaline Salt-Lakes by Selective Electrodialysis: Transport and Thermodynamics

Link back to the item: <https://hdl.handle.net/1721.1/152311>

13 Number of Pages: 43. Number of Figures: 13. Number of Tables: 40.

Contents

List of Figures

⁷⁸ List of Tables

¹⁹³ A Computational Methods

¹⁹⁴ A.1 Nernst-Planck Model for Ion Transport in Electrodialysis

¹⁹⁵ Based on chemical thermodynamics, the fundamental relationship between the Gibbs free energy ¹⁹⁶ and the temperature, pressure and species mole fractions can be expressed as Eq. 1

$$
dG = -SdT + v dP + \sum_{i} (RT \ln \gamma_i c_i + z_i F \Psi) dN_i
$$
\n(1)

197 where S (J mol⁻¹) and T (K) denote the system entropy and temperature, v (m³ mol⁻¹) and P 198 (Pa) denote the specific molar volume and system pressure, and γ_i (-) and C_i (mol L⁻¹) denote the rational activity coefficient and concentration of species i, z_i (-), F (C mol⁻¹), Ψ (V) and N_i 200 (mol) denote the ion valency, Faraday's constant, electric potential and molar quantity of species i, ²⁰¹ respectively.

Based on classical irreversible thermodynamics, the species molar flux can be modeled with the first spatial derivative of the species chemical potential, when the operating point is sufficiently close to thermodynamic equilibrium. ³ The Nernst-Planck (NP) equation, as described by Eq. 3, is obtained in combination with Eq. 1, in the isothermal and isobaric (NPT) ensemble

$$
J_i = -L_i \frac{d\mu_i}{dz} \tag{2}
$$

$$
= -D_i^{mem} \nabla c_i^{mem} - \frac{D_i^{mem} z_i F}{RT} c_i^{mem} \nabla \Psi \tag{3}
$$

where L_i denotes the proportionality constant in the chemical potential gradient and $D_i^{mem} =$ L_iRT/c_i^{mem} (m² s⁻¹) denotes the Fickian diffusion coefficient of species i. The terms in the NP equation correspond to ion transport from diffusion and electromigration, which are driven by the concentration and electric potential gradients, respectively.

²⁰⁶ For a binary salt, the expressions for the cation and anion fluxes can be written based on Eq. 3, ²⁰⁷ as provided in Eq. 4 and 5

$$
J_c = -D_c^{mem} \nabla c_c^{mem} - \frac{D_c^{mem} z_c F}{RT} c_c^{mem} \nabla \Psi
$$
\n
$$
(4)
$$

208

$$
J_a = -D_a^{mem} \nabla c_a^{mem} - \frac{D_a^{mem} z_a F}{RT} c_a^{mem} \nabla \Psi \tag{5}
$$

²⁰⁹ where subscripts c and a denote the cation and anion thermodynamic states and transport prop-²¹⁰ erties, respectively. Under the assumption of solution electroneutrality, the cation and anion fluxes ²¹¹ are constrained by Eq. 6

$$
J_a = \frac{I_{den}}{z_a F} - \frac{z_c}{z_a} J_c \tag{6}
$$

212 where I_{den} (A m⁻²) denotes the applied current density. The stoichiometry coefficients of the dissociated ions $(\nu_c$ and $\nu_a)$ can be related to their respective ionic valencies with Eq. 7. Further, ²¹⁴ the stoichiometry and diffusion coefficients of the dissociated ions can be related to the apparent ²¹⁵ salt diffusion coefficient with the Nernst-Hartley equation, as depicted in Eq. 8.

$$
\frac{z_a}{z_c} = -\frac{\nu_c}{\nu_a} \tag{7}
$$

216

$$
D_{c-a} = \frac{(\nu_a + \nu_c) D_a D_c}{\nu_a D_c + \nu_c D_a}
$$
(8)

217 The transport number of ion i, as represented by τ_i (-), which is defined as the fraction of the current ²¹⁸ conducted by the ion across the ion exchange membrane relative to the total applied current, can ²¹⁹ be expressed as a function of the ion concentration, diffusion coefficient and valency in Eq. 9.

$$
\tau_i = \frac{z_i^2 D_i c_i^{mem}}{\sum_j z_j^2 D_j c_j^{mem}}\tag{9}
$$

 T_{220} Together with classical solution-diffusion theory, 4 Eq. 4–9 can be condensed to obtain an explicit ²²¹ relationship between the cation and anion fluxes and the applied current density and concentration ²²² gradients, as provided in Eq. 10

$$
J_i = \left\{ \frac{\tau_i I_{den}}{z_i F} + B_i \left[C_i^{d, int} - C_i^{c, int} \right] \right\}
$$
\n
$$
(10)
$$

where B_i (m s⁻¹) denote the solute permeability coefficient for diffusion, and $C_i^{d,int}$ 223 where B_i (m s⁻¹) denote the solute permeability coefficient for diffusion, and $C_i^{d,int}$ (mol L⁻¹) and $C_i^{c,int}$ and $C_i^{c,int}$ (mol L⁻¹) denote the solute concentration in the solution phase along the membrane-²²⁵ solution interface for the diluate and concentrate electrolyte streams, respectively. The model was successfully generalized for multi-ionic mixtures by modeling the solutes as individual ions instead of $_{227}$ binary salts.⁵ As discussed in a prior publication from our group,⁵ the transport number and solute permeability coefficient will become ion-specific, and are a function of the membrane properties and the feed composition, salinity and acidity. Experiments with multi-ionic solutions with the representative compositions must be used to accurately determine the respective coefficients.

²³¹ The conventional Nernst-Planck equation assumes that the ion transport by convection is small ²³² as compared to the contributions from diffusion and electromigration.⁵ This assumption typically ²³³ holds for monovalent selective ion exchange membranes that are designed for salt production, as a result of their thick composite structure that renders the water permeability coefficients low.^{1,6} ²³⁵ To capture convective coupling between the solvent and ion transport in the event of significant t_{z} transmembrane water flux, the model should be extended to include convection.⁷ In this work, we ²³⁷ experimented with multi-layered composite ion exchange membranes. Consequently, the IEMs are ²³⁸ thick and the water permeability of the membranes is usually about an order of magnitude lower t_{239} than unmodified IEMs.⁶ With these membranes, our experiments indicated that the kinetics of ²⁴⁰ water transport is approximately an order of magnitude lower the corresponding values for the ions, which aligns with recent reports.^{2,6} As a consequence, the molar ion flux by convection (which is defined as the product of the ion concentration and the molar water flux⁸) accounts for less than ²⁴³ 1 % of the total ion flux in all of our experiments. These results suggest that convection likely ²⁴⁴ played a minor role in ion transport with the IEMs in this study.

²⁴⁵ Mass transfer limitations arising from concentration polarization in the electrolyte streams, as depicted in Fig. 2B, are incorporated with Eq. $11⁹$ 246

$$
\Delta C_i = \left(\frac{\tau_i - t_{cu,i}}{D_i}\right) \left(\frac{I_{den}}{z_i F}\right) \left(\frac{2h}{\text{Sh}}\right)
$$
\n(11)

²⁴⁷ where ΔC_i (mol L⁻¹) denotes the concentration difference between the bulk flow and the membrane the interface, $t_{cu,i}$ (-) is the integral counter-ion transport number of species i, D_i (m² s⁻¹) is the Fickian 249 diffusion coefficient, h (m) is the channel height and Sh (-) is the Sherwood number. The Sherwood $_{250}$ number is calculated based on mass transfer correlations proposed by McGovern et al.⁹ for the ²⁵¹ spacers adopted in our experiments.

252 The integral counter-ion transport number, $t_{cu,i}$ (-), is calculated based on Eq. 12, in accordance

253 with the ED literature.^{5,10} Following which, the limiting current density $(I_{den,i}^{lim})$ of the each ion is ²⁵⁴ calculated to ensure that the experiments are conducted within the ohmic regime. Any current ²⁵⁵ excess of this limit leads to water dissociation along the fluid-membrane interface, or accelerates the $t_{\text{transport}}$ transport of multivalent and co-ions, 11 both of which negatively impacts the monovalent selectivity of the IEMs. The limiting current density for each ion i can be calculated with Eq. 13^5 257

$$
t_{cu,i} = \frac{z_i D_i C_i^{d,bulk}}{\sum_{j=1}^{N} z_j D_j C_j^{d,bulk}}
$$
\n(12)

258

$$
I_{den,i}^{lim} = \left(\frac{D_i z_i F}{\tau_i - t_{cu,i}}\right) \left(\frac{2h}{\text{Sh}}\right) C_i^{d,bulk} \tag{13}
$$

 259 where h (m) denotes the characteristic length of the flow channels.

²⁶⁰ A.2 Counter-ion Transport across Ion Exchange Membranes

²⁶¹ Ion exchange membranes (IEM) are water-swollen polymers that comprises two phases: 1) a charged ²⁶² gel phase formed by the hydrocarbon polymer chains and the hydrophilic ion exchange functional ²⁶³ groups and 2) an interstitial phase formed from the pores, interstices and structural defects of the gel 264 phase.⁶ The void size is dictated by the swelling degree of the IEM, which is a function of the water $_{265}$ volume fraction of the membrane.² In conventional ED literature, the interstitial phase is assumed to be filled by an electroneutral solution when the IEMs are contacted with a saline stream.¹²

 Driven by an electrochemical potential gradient, ions partition from the solution to the inter- $\frac{13}{268}$ stitial phase of the IEM, and are transported across the IEM by diffusion and electromigration.¹³ The ion selectivity of the IEM is governed by their respective ease of ion partitioning, and their respective mobilities within the interstitial phase. The Donnan exclusion mechanism is commonly employed to rationalize the relative differences in ion partitioning across the solution-membrane interface. Under this framework, the Gibbs free energy of the ions is assumed to be continuous across the solution-membrane interface to ensure chemical stability, as described in Eq. 14

$$
\psi_i = \frac{\gamma_i^{mem}C_i^{mem}}{\gamma_i^{d,int}C_i^{d,int}} = f_w \left[\exp\left(-\frac{z_i F}{RT} \Delta \phi_{Donnan} \right) \right]
$$
\n(14)

274 where f_w (-) represent the water volume fraction within the membrane, $\Delta\phi_{Donnan} = \phi_i^{mem}$ –

 $\phi^{d,int}_i$ $\alpha_i^{d,int}$ (V) denote the Donnan potential and, γ_i^{mem} (-) and $\gamma_i^{d,int}$ 275 $\phi_i^{a,int}$ (V) denote the Donnan potential and, γ_i^{mem} (-) and $\gamma_i^{a,int}$ (-) represent the ion activity ²⁷⁶ coefficient within the membrane polymer and in the solution along the membrane-solution interface, respectively. The activity coefficients are predicted using Manning's counter-ion condensation^{12,14} 278 and Pitzer-Kim models, $15,16$ using the method described by Fan and Yip. 2 Based on these theories, ²⁷⁹ the ion concentration within the membrane phase is normalized by the volume of the interstitial 280 phase of the IEM.^{2,13}

²⁸¹ Electroneutrality conditions are applied to the bulk solution and IEM's interstitial phases, as ²⁸² described by Eq. 15 and 16

$$
\sum_{i}^{N} z_i C_i^{d.int} = 0 \tag{15}
$$

283

$$
C_{fixed}^{mem} + \sum_{i}^{N} z_i C_i^{mem} = 0
$$
\n(16)

284 where N (-) is the total number of ions and C_{fixed}^{mem} (mol L⁻¹) denotes the molar ion exchange capacity ²⁸⁵ of the IEM.¹⁷ Here, we set C_{fixed}^{mem} to be 1.68 M based on prior ion exchange capacity measurements on cation exchange membranes (CEM).^{2,18} Eq. 14, 15 and 16 are solved simultaneously using the ²⁸⁷ constrained trust region method to derive the concentrations of the partitioned ions and the resultant P_{288} Donnan potential, implemented with numerical solvers in Python.¹⁹

 The mobility of counter ions in polysulfonate CEM has received emerging interest in recent literature. Using an extended Mackie-Meares framework, Fan et al. demonstrated that a linear relationship exists between the ratio of the ion diffusion coefficient within the membrane relative to the bulk solution, and the exponential of the square of the ionic valency.¹³ The observed reduction in the apparent diffusion coefficient within the polymer matrix arises from spatial hindrance from the tortuosity of interstitial phase, and the electrostatic friction between the cations and the ion exchange groups.²⁰ The relationship between the bulk and interstitial phase diffusion coefficients is given by Eq. 17.

$$
D_i^{mem} = D_i^{d,int} \left(\frac{f_w}{2 - f_w}\right)^2 \exp(-A_{el} z_i^2)
$$
 (17)

where D_i^{mem} (m² s⁻¹) and $D_i^{d,int}$ 297 where D_i^{mem} (m² s⁻¹) and $D_i^{d,int}$ (m² s⁻¹) denote the ion diffusion coefficient within the interstitial 298 and bulk solution phases, respectively, and A_{el} (-) represents an electrostatic friction parameter that is a function of the fixed charge density (C_{fixed}^{mem}) and the apparent dielectric constant. A_{el}

³⁰⁰ is commonly employed as a regression parameter in IEM diffusion experiments, and the derived son constant ranges between 0.329 to 0.691 for commercial CEMs.¹³

 The Nernst-Planck equation is utilized in combination with the extended Mackie-Meares and Donnan equilibria models to analyze the implications of current density increments. Based on 304 our experiments, when the applied current density is 2.5 mA cm⁻² or higher, we found that ion transport by electromigration dominates; this conclusion is corroborated by prior experimental α ₃₀₆ and computational studies.^{10,21,22} When concentration-driven diffusion is small as compared to electric potential-driven electromigration, based on the Nernst-Planck equation, the ratio of ion fluxes between two distinct species tends towards the expression given by Eq. 18.

$$
\frac{J_i^{Mod.}}{J_j^{Mod.}} \to \frac{D_i^{mem}}{D_j^{mem}} \frac{z_i}{z_j} \frac{C_i^{mem}}{c_j^{mem}} \tag{18}
$$

³⁰⁹ Eq. 14 and 17 are leveraged to relate the ion concentration and diffusion within the interstitial ³¹⁰ phase to the bulk solution phase. The resultant expression is provided by Eq. 19

$$
\frac{J_i^{Mod.}}{J_j^{Mod.}} = \frac{D_i^{d,int}}{D_j^{d,int}} \frac{z_i}{z_j} \frac{C_i^{d,int}}{C_j^{d,int}} \overline{\gamma}_{i,j} \exp\left[-A_{el}\left(z_i^2 - z_j^2\right)\right] \exp\left[-\frac{\Delta\phi_{Domain}F}{RT}\left(z_i - z_j\right)\right]
$$
(19)

311 where $\bar{\gamma}_{i,j}$ (-) represents the ratio of activity coefficients between the solution and the interstitial phase, between species i and j, respectively. For calculations of Li/Mg and Na/Mg flux ratios, the $\overline{\gamma}_{Li,Mg}$ and $\overline{\gamma}_{Na,Mg}$ ranges between 0.2874 – 0.2954 and 0.2698 – 0.2948, respectively. In this expression, the Donnan potential and the interfacial concentrations are functions of the applied 315 current density, while A_{el} is a constant material parameter for a given CEM. The L²-norm error between the experimental and model ion flux ratios is minimized using the constrained trust region algorithm. The derived A_{el} value is 0.412, which is within the reported range for commercial CEMs (0.329 - 0.691).

³¹⁹ Eq. 19 indicates that the ion flux ratio between two species is proportional to the interfacial ³²⁰ concentration ratio and two exponential terms. The first exponential term (defined as the mobility ³²¹ factor) corresponds to the differences in the ion mobility arising from electrostatic friction within the CEM¹³ while the second exponential term (defined as the partitioning factor) corresponds to the differences in ion partitioning from Donnan exclusion.⁶ Multivalent cations experience stronger electrostatic friction within the CEM due to interactions with the negatively charged sulfonate moi t_{25} eties.^{2,6} Therefore, the mobility factor between monovalent and multivalent cations is greater than unity. An increment in the applied current density leads to a reduction in the ion concentrations along the membrane-solution interface, improving the effectiveness of Donnan exclusion in rejecting multivalent ions. This leads to an increase in the resultant Donnan potential, amplifying the parti- tioning factor across the interface. When the applied current density is amplified, the combination of the mobility and partitioning factors in Eq. 19 induces a larger than proportional increase in the resultant ion flux ratios between monovalent and multivalent ions.

A.3 Impact on Ion Partitioning and Overall Selectivity

 Ion selectivity is a result of differences in 1) ion partitioning and 2) ion mobility across the mem b^{334} brane.^{6,23} In this study, a two-pronged computational approach was adopted to investigate the ion selectivity of the composite cation exchange membranes. In the first approach, as described in Section 2.3 and Section A.1 of the SI, the transport numbers of the respective ions under differ- ent operating conditions (i.e., current densities, feed salinities, composition and solution pH) were determined, and were systematically compiled in Supp. Table 40. As described in Eq. 1 and 2, the transport number is defined as the proportion of current conducted by the ion relative to the total applied current, and it characterizes the combined effects of ion partitioning and mobility for electromigration across the IEM.^{5,23} As a result, the effective ion selectivity of both the PEI surface layer and the PS-DVB substrate is captured by the regressed transport numbers in our study. The separation factors were calculated with Eq. 1 and 2, leveraging the derived transport numbers and ion permeability coefficients.

 In the second approach, as described in Section 2.3 and Section A.2 of the SI, we seek to under- stand the impact of the feed solution concentration and solution pH on efficacy of Donnan exclusion ³⁴⁷ for multivalent cation rejection, using the method developed by Fan and Yip.² This approach was selected because prior studies on hypersaline electrodialysis indicated that the weakening of Donnan exclusion was the primary mechanism for the observed reduction in counter-ion/co-ion selectivity with high concentration feed solutions. $2,24$ In accordance with the theory of Donnan exclusion, the $\frac{1}{351}$ ion concentrations within the PEI surface layer were simulated with Eq. 3,² and are presented

 in Supp. Fig 13. Donnan exclusion from the positively-charged surface layer reduces the parti $t_{\rm 553}$ tioning rate of multivalent cations, as a consequence of their higher charge densities. 25 Therefore, the monovalent ions (i.e., Li⁺, Na⁺, K⁺) have a greater partition coefficient as compared to Mg²⁺, aligning with the relative ion transport rates observed in Supp. Fig. $2 - 12$. Given that the PS-DVB substrate is negatively charged, exclusion of cations by the Donnan effect is negligible across the $\frac{1}{357}$ interface between the PEI surface layer and the PS-DVB substrate.²⁵

B Experimental Methods

B.1 Experimental Apparatus

 The bench-scale experimental setup that is employed for membrane characterization is depicted in Supp. Fig. 1. The cathode and anode are fashioned from platinum-coated titanium electrodes. A total of 20 feed spacers and 2 end spacers, each with a 0.5 mm thickness and a 45 \degree mesh orien- tation, are placed in the electrolyte channels. Sodium sulfate (0.2 M) solutions are used as the catholyte and anolyte to suppress chlorine gas production and stabilize the solution $\rm pH.^{26}$ The diluate, concentrate and rinse circuits, comprising 2L, 4L and 4L of the respective solutions, are operated in a batch configuration. By starting with the same composition and concentrations on both streams, the contribution from concentration-driven diffusion can be decoupled and minimized between the experiments.²⁷ A larger concentrate solution volume is selected to slow the rate of in- crease of concentrate stream over the course of the experiment, reducing the impact of ion transport $\frac{1}{270}$ from back-diffusion.¹ Cross-flow is maintained at 95 L h⁻¹ with centrifugal pumps (Iwaki MD55R) and valved rotameters. A potentiometric feedback loop regulates the temperature of the solutions and a DC power supply generates the electric potential gradient for ion transport (GW Instek GPR- 60600). The solution volumes are monitored with digital mass scales (Ohaus Scout Pro SP601). The solution conductivity, pH, cell potential difference and temperature are recorded at 60 Hz with a digital acquisition unit (Vernier LabQuest, Hach HQ440d). The concentrate and diluate solutions are sampled periodically and analyzed with ICP-OES. The wavelengths of the respective elements are selected to avoid signal interference, and are summarized in Supp. Table 1.

The experiments in this study were all conducted at a constant current density to accurately

Supplementary Figure 1: Schematic diagram of the bench-scale electrodialysis adopted in this investigation. The diluate, concentrate and rinse loops are composed of 2, 4 and 4 L of electrolyte solutions, are cycled through the electrodialysis stack with centrifugal pumps. Over 10 alternating cell pairs of CEMs and AEMs comprise the membrane stack, which is housed inside of a PCCell ED 200 unit. A counter-flow heat exchanger regulates the electrolye stream temperatures to a pre-set value. An external direct current power supply is used to generate the potential difference for ion transport. The illustration is adapted from our prior publication and used with permission from E lsevier.¹

 σ characterize the transport numbers of the respective ions. $5,10$ This is because the transport numbers ³⁸⁰ are defined as the proportion of the current conducted by a particular species, and constant current $_{381}$ conditions ensure that the normalization is standardized between the various experiments. 5,10 In our bench-scale setup with 10 cell pairs, the majority of the voltage drop occurs at the electrodes. ¹ 382 ³⁸³ In the event of unprecedented side redox reactions, we would likely observe effervescence in the rinse l_{384} 10 loop or solid deposition at the electrodes.¹⁰ However, we did not observe any noticeable changes ³⁸⁵ between the different current densities during our experiments.

Supplementary Table 1: Selected wavelengths for spectroscopic analysis with ICP-OES.

³⁸⁶ B.2 Solution Composition and Results

³⁸⁷ In this paper, experiments are conducted on binary cation and multicomponent salt-lake brines ³⁸⁸ from Chile and China to ascertain the relative ion transport rates. The ionic composition of the ³⁸⁹ solutions are provided in Supp. Tables 2, 3 and 4, respectively. The feed salinities vary from 10 to 390 250 g/L, the solution pH from 3 to 7, and the current density from 2.5 to 30.0 mA cm⁻².

 The ion composition of the diluate stream is characterized based on ICP-OES. The propagated uncertainties in the respective ion concentration are all under 4.5 %. To provide a comprehensive database upon which future IEMs can be compared against, the experimental results are systemically tabulated. Using the computational framework described in Section A, the ion transport number, Donnan potential and the electrostatic friction parameter are determined based on Eq. 8 in the main text, for each of the tested compositions. The ion-specific permeability coefficients were determined based on diffusion experiments with multi-ionic feed solutions as described in our prior $p_{\text{sublications}}^{1,5,26}$ The agreement between the model predictions and the empirical measurements is illustrated in Supp. Fig. 2-12.

⁴⁰⁰ Ion selectivity arises as a result of differences in 1) ion partitioning and 2) ion mobility across the $_{401}$ membrane.^{5,6} The ion selectivity sequence for a variety of selective electrodialysis membranes has been systematically compiled in review articles by Luo et al.⁶ and Ying et al.²⁸ In general, Donnan ⁴⁰³ exclusion reduces the partitioning rate of multivalent ions, as a consequence of their higher charge densities.² Therefore, the monovalent ions (L_i^+, Na^+, K^+) have a greater partition coefficient as compared to Mg^{2+} , aligning with the relative ion transport rates observed in Supp. Fig. 2 – 12. ⁴⁰⁶ In accordance with the theory of Donnan exclusion, the ion concentrations within the PEI surface l_{407} layer were simulated with Eq. 3 and are presented in Supp. Fig 13.² The model indicated that the relative concentrations of $Li⁺$ to $Mg²⁺$ within the IEM increased up to 10, as compared to 1.8 in

Supplementary Table 2: Nominal ionic composition of binary cation feed solutions based on brine from Salar de Atacama, at a solution molarity of 0.35 M.

Brine Composition	Nominal Composition (g/L)		Solution Molarity (M)	
(Abbreviation)	Li^+ Mg ²⁺ Cl ⁻ SO ₄ ²⁻ TDS			
$\overline{\text{Li}^+-\text{Mg}^2}$ +-Cl ⁻ (LM-C)	0.34 2.05 7.70 0.00 10.09			0.35
Li^{+} -Mg ²⁺ -SO ₄ ²⁻ (LM-S)	0.49 2.98 0.00 15.13 18.59			0.35
Li ⁺ -Mg ²⁺ -Cl ⁻ -SO ₄ ² (LM-CS) 0.34 2.09 7.39 0.62 10.44				0.35

Salt Lake, Location	Nominal Composition (g/L)								
						Li^+ Na ⁺ K ⁺ Mg ²⁺ Ca ²⁺ Cl ⁻ SO ₄ ²⁻ TDS			
	1.19	69.01	- 17.89	7.31	< 0.01	- 143.72	12.06	251.18	
	0.33	19.23	4.99	2.04	< 0.01	40.10	3.36	70.00	
Salar de Atacama,	0.14	8.24	2.14	0.87	< 0.01	17.17	1.44	30.00	
Chile	0.05	2.75	0.71	0.29	< 0.01	5.72	0.48	10.00	

Supplementary Table 3: Nominal ionic composition of feed solution based on brine from Salar de Atacama, Chile, at total dissolved concentrations of 10, 30, 70 and 250 g/L .

Supplementary Table 4: Nominal ionic composition of feed solution based on brine from Qaidam Lake, China, at total dissolved concentrations of 10, 30, 70 and 250 g/L .

Salt Lake, Location						Nominal Composition (g/L)		
						Li^+ Na ⁺ K ⁺ Mg ²⁺ Ca ²⁺ Cl ⁻ SO ₄ ²		TDS
Qaidam Lake, China	0.31 0.09 0.04 0.01	56.30 15.79 6.77 2.26	1.23 0.53 0.18	4.40 20.20 5.67 2.43 0.81	< 0.01 < 0.01 < 0.01 < 0.01	134.20 37.65 16.14 5.38	34.10 9.57 4.10 1.37	249.51 70.00 30.00 10.00

⁴⁰⁹ the feed solution.

⁴¹⁰ Due to the smaller differences in charge density between the monovalent ions, Donnan exclusion 411 does not significantly influence the relative partitioning rate of Li^+ compared to Na⁺ and K⁺. ⁴¹² Despite that, as illustrated in Supp. Fig. 2 - 12, our experiments reveal faster transport rates f_{413} for Na⁺ and K⁺. The differences in relative transport rates of the monovalent cations have been rationalized in the literature by their respective mobilities within the ion exchange membrane. ^{6,13} The order of the diffusion coefficients follows the descending sequence: $H_3O^+ > K^+ > Na^+ > Li^+,$ 416 which agrees with our experiments, as illustrated in Supp. Fig. $2 - 12$.

Supplementary Figure 2: Comparisons between the experimental measurements and model predictions for 10 g/L multicomponent brines from Salar de Atacama, Chile, under a constant current density of 2.5 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 3: Comparisons between the experimental measurements and model predictions for 30 g/L multicomponent brines from Salar de Atacama, Chile, under a constant at pH 7, for current density of (A) 2.5 and (B) 7.5 mA cm⁻², respectively.

Supplementary Figure 4: Comparisons between the experimental measurements and model predictions for 70 g/L multicomponent brines from Salar de Atacama, Chile, under a constant current density of 2.5 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 5: Comparisons between the experimental measurements and model predictions for 70 g/L multicomponent brines from Salar de Atacama, Chile, under a constant current density of 7.5 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 6: Comparisons between the experimental measurements and model predictions for 70 g/L multicomponent brines from Salar de Atacama, Chile, under a constant current density of 15.0 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 7: Comparisons between the experimental measurements and model predictions for 10 g/L multicomponent brines from Qaidam Lake, China, under a constant current density of 2.5 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 8: Comparisons between the experimental measurements and model predictions for 30 g/L multicomponent brines from Qaidam Lake, China, under a constant at pH 7, for current density of (A) 2.5 and (B) 7.5 mA cm⁻², respectively.

Supplementary Figure 9: Comparisons between the experimental measurements and model predictions for 70 g/L multicomponent brines from Qaidam Lake, China, under a constant current density of 2.5 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 10: Comparisons between the experimental measurements and model predictions for 70 g/L multicomponent brines from Qaidam Lake, China, under a constant current density of 7.5 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 11: Comparisons between the experimental measurements and model predictions for 70 g/L multicomponent brines from Qaidam Lake, China, under a constant current density of 15.0 mA cm⁻² at pH (A) 7, (B) 5 and (C) 3, respectively.

Supplementary Figure 12: Comparisons between the experimental measurements and model predictions for 250 g/L multicomponent brines from (A) Salar de Atacama, Chile and (B) Qaidam Lake, China, under a constant current density of 2.5 mA cm⁻² at pH 7; corresponding results for 250 g/L multicomponent brine from Salar de Atacama, Chile, under a constant current density of 30.0 mA cm⁻² at (C) pH 7 and (D) pH 3, respectively.

⁴¹⁷ C Supplemental Analysis

⁴¹⁸ C.1 Charge Density Impact on Donnan Exclusion

Supplementary Figure 13: Concentration of Li⁺, Mg²⁺ and Cl[−] ions within the PEI layer of the CEM for a constant volumetric charge density, as a function of the external solution concentration along the fluid-membrane interface. Relative concentrations of Li^+ to Mg^{2+} decay with increasing solution concentration, arising from weakening Donnan exclusion effects.² The volumetric charge densities of the PEI layer are (A) 0.5 M, (B) 1.68 M, and (C) 5.0 M, respectively, to simulate the effects of the lowered volumetric charge density with acid pre-treatment for salt-lake brine applications.

 Our zeta potential measurements presented in Fig. 3C indicate that the interfacial potential of the diffuse layer decreases by approximately 41.7 %, likely as a result of a lowered volumetric ⁴²¹ charge density under low solution pH. This coincides with an increase in Mg²⁺ leakage by 18 %, and a decrease in the monovalent cation fluxes by 14 % or greater, suggesting that the weakening of Donnan exclusion is principal for the selectivity decline.

 Using the terminology as defined in Figure 3D and E in the main manuscript, Eq. 14–16 in the SI are solved simultaneously to analyze the impact of the volumetric charge density on the monovalent selectivity of the CEM. Here, a CEM with volumetric charge densities of 0.50, 1.68 and 5.00 M are ⁴²⁷ used to simulate the impacts of the solution pH on the charged moiety density of the PEI layer, 2,29 as illustrated in Supp. Figures 13A, B and C, respectively.

⁴²⁹ With a high volumetric charge density of 5.0 M, as observed in Supp. Figure 13C, a large 430 Li⁺/Mg²⁺ molar ratio of the partitioned ions exceeding 30 is obtained at 10 g L⁻¹ TDS (0.35 M). While the partitioning selectivity decreases with increasing feed solution concentrations, a Li^+/Mg^{2+} ⁴³² molar ratio that is greater than parity is obtained even with salt-lake brine concentrations. When the

S24

 volumetric charge density is reduced from 1.68 to 0.5 M to mimic our zeta potential observations, the Li^+/Mg^{2+} molar ratio of the partitioned ions decreases drastically across the entire concentration range, and falls under parity once the solution concentration exceeds the 0.5 M. When the solution concentration exceeds the volumetric charge density of the CEM, more co-ions (i.e., Cl^-) will partition into the interstitial phase, and additional counter-ion partitioning (i.e., Li^+ , Mg^{2+}) is necessary to maintain the electroneutrality condition. As a consequence, our model indicate that the resultant Donnan potential will attenuate, and more multivalent cations will be able to partition successfully into the CEM, reducing the effectiveness of Donnan exclusion of Mg^{2+} ions. Therefore, our numerical findings corroborate our empirical inference of the weakening of Donnan exclusion under low solution pH.

C.2 Composition Impacts on Monovalent Selectivity

 In our prior NF publication, we observed that apparent Li/Mg separation factors may be overesti- mated by up to 40 % when the anionic composition of the feed solution is simplified to just one anion 446 (either Cl[−] or SO₄²⁻).³⁰ This arises from transport coupling between the cations and anions, induced by electroneutrality constraints. For example, with Li^+ -Mg²⁺ -SO₄²⁻ feed solutions, the apparent Li⁺ and Mg²⁺ rejection was significantly overestimated as a result of the poor permeability of SO²[−] ⁴⁴⁹ ions. When the Cl[−]/SO²⁻ molar ratio was accurately replicated with the Li⁺-Mg²⁺-Cl⁻-SO²⁻ feed solutions, the absolute errors for Li/Mg separation factors fell under 15 %. Similar to our previous NF study, our ED experiments with binary cation feed solutions overpredict the Li/Mg separation factors by 50 to 250 %. Unlike NF, however, the errors in the Li/Mg separation factors remained large even when the Cl[−]/SO^{2−} molar ratio was accurately replicated with the Li⁺-Mg²⁺-Cl[−]-SO^{2−} feed solutions. This result indicates that the other monovalent cations have an influence on the relative ionic flux between Li^+ and Mg^{2+} .

 Similar to NF, the diluate and concentrate streams are still subjected to same electroneutrality constraint. In electrodialysis, however, the bulk anion and cation ions are transported separately through the AEM and CEM, and the two ion exchange membranes behave as separate conductors that are connected in series.⁵ For a given current density, ED experiments with Li^{+} -Mg²⁺-SO₄⁻ (LMS) feed solutions will register a much greater SO_4^{2-} ion flux than that with Li⁺-Mg²⁺-Cl–SO₄⁻

⁴⁶¹ (LMCS) solutions, albeit with a significantly amplified potential difference across the membrane stack due to the low conductance of SO_4^{2-} ions across the AEM. If the applied current density is equal in both experiments, the ion flux of SO_4^{2-} does not appear to present a significant barrier ⁴⁶⁴ to the transport of the monovalent cations between the LMS and LMCS experiments. This is likely because the AEM is not perfectly monovalent selective and will allow SO_4^{2-} leakage when the ⁴⁶⁶ potential difference across the AEM is sufficiently high.

 \sim On the other hand, if the AEM rejects SO_4^{2-} ions perfectly (hypothetical ideal monovalent selective AEM), then the ionic flux across the AEM would be negligible in experiments with $Li⁺$ - Mg^{2+} -SO²⁻ (LMS) feed solutions, and the ohmic resistance across the AEM would tend towards ⁴⁷⁰ infinity. In this scenario, we would expect to see the same behavior in the Li/Mg separation factors as with NF, where the low transport rate of SO_4^{2-} limits the net ionic flux of Li⁺ and Mg²⁺. However, our AEMs are not perfectly selective, and significant SO_4^{2-} ion fluxes were measured with ⁴⁷³ our LMS experiments.

⁴⁷⁴ C.3 Lithium Concentration in Salar de Atacama

 In this section, we assess the implications of our research findings on the process duration and land area requirements for direct salt-lake lithium extraction. Currently, over 50 % of the global lithium supply originates from continental and geothermal brines, leveraging evaporitic technology f_{478} for solution concentration.³¹ Presently, the lithium-enriched hypersaline brines are pumped into evaporation ponds and concentrated by direct solar irradiation, over an average production cycle of ⁴⁸⁰ two years.³² Carbonates and phosphates of sodium and potassium are injected into the evaporation ponds when the solution concentrations of magnesium and lithium ions are close to their respective ⁴⁸² saturation indices.³³ The precipitated salts are subsequently collected and washed with fresh wa t_{483} ter.³¹ The production dependence on evaporation ponds renders the lithium supply to be extremely price inelastic and slow to respond to the market demand. Further, only large flat areas in low humidity climates that are not prone to monsoon or seasonal changes are suitable to function as evaporation ponds. As a result, there are very few suitable flat lands that coincide geographically $\frac{487}{100}$ with lithium-rich aqueous deposits that can be exploited for direct lithium extraction.³³ With the appropriate selective IEM, an industrial-scale electrodialysis system can concentrate lithium in a

⁴⁸⁹ continuous process, potentially overcoming the reliance on evaporation ponds and access untapped ⁴⁹⁰ lithium deposits for direct extraction.

⁴⁹¹ The Salar de Atacama salt-lake in Chile produces 121,190 tones of lithium carbonate per year with 3000 km² of salt flat area.³¹ Each production cycle takes approximately 2 years, and over 493 2.7×10^6 m³ of fresh water is consumed for solid salt isolation. On a molar basis, the Chilean salt-lake produces 8.99×10^6 moles of Li on average per day, and the land area impact per mole of 495 Li produced is 3.34×10^{-4} km² mol⁻¹. For electrodialysis, the land area requirement per mole of ⁴⁹⁶ Li produced can be calculated with Eq. 20

$$
\overline{A}_{Li} = \frac{\text{LAI } P_{ED}}{t_{Prod} \ J_{Li} \ A_{Mem}} \tag{20}
$$

497 where \overline{A}_{Li} (km² mol⁻¹) denotes the normalized land area requirement, LAI (km² GW⁻¹) denotes 498 the land area impact per unit power, P_{ED} (GW) denotes the power consumed by the industrial-scale electrodialysis stack, t_{Prod} (h) denotes the daily production time, J_{Li} (mol m⁻² h⁻¹) denotes the soo molar flux of Li, and A_{Mem} (m²) denotes the total membrane area of the IEMs.

 μ ₅₀₁ Here, we leverage our experimental results on the 250 g L⁻¹ Chilean brines to project the energy and land area requirements for lithium extraction. An industrial-scale electrodialysis module comprising 100,000 cell pairs that is typically used for salt production is employed to estimate ED 's performance.⁹ To simulate DLE application, the diluate and concentrate stream compositions $\frac{1}{205}$ correspond to the salt-lake brine and a 0.1 M NaCl solution.³⁴ The land area impact of a photovoltaic solar farm to meet ED's electrical work requirement is calculated based on the survey by Palmer- Wilson et al., 35 which incorporates the spatial demands arising from power generation, storage and transmission in a solar farm.

⁵⁰⁹ Assuming a 10 h daily production cycle, using existing monovalent selective CEMs driven by μ ₅₁₀ a current density of 30.0 mA cm⁻², our model indicates that over 7.70 ×10⁶ moles of Li can be ⁵¹¹ extracted per day with a commercial-scale ED unit operation; the quantity of Li extracted from a ⁵¹² single ED operation is over 85 % of the current capacity of Salar de Atacama. In a single stage, 513 over 85.8 % of the dissolved Li can be extracted, and the product Li^{+}/Mg^{+} ratio decreases by a ⁵¹⁴ factor of 6.58. Correspondingly, based on the land area impact survey, a solar farm operating for 10 h daily with a total footprint between 11.35 to 12.84 km² can generate sufficient electrical work to

power the continuous ED process.³⁵ The normalized land requirement (\overline{A}_{Li}) for ED is estimated to $_{517}$ be between 1.21×10^{-6} to 1.67×10^{-6} km² mol⁻¹, which is over two orders of magnitude lower than ⁵¹⁸ the corresponding value obtained for the evaporation ponds at Salar de Atacama. The favorable ⁵¹⁹ results arise because the electrical work required for ion transport in ED is significantly lower than the latent heat consumed to vaporize water in a evaporation pond. 36

⁵²¹ D Experimental Data

Supplementary Table 5: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 10 g/L, solution pH of 7, under a current density of 2.5 mA cm−² .

min)	$Li+$ $\overline{}$	$\overline{\text{Mg}}^{2+}$ -	$\rm Na^+$ $\overline{}$	$\rm K^+ (-$	Cl^{-} $\overline{}$	$\overline{}$ $\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.929	0.974	0.917	0.815	0.898	1.038
30	0.847	0.977	0.787	0.660	0.818	1.071
45	0.815	0.971	0.651	0.515	0.726	1.098
60	0.745	0.943	0.599	0.392	0.596	1.106
120	0.434	0.833	0.213	0.096	0.252	1.147

Supplementary Table 6: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 10 g/L, solution pH of 5, under a current density of 2.5 mA cm⁻².

min)	$Li+$ $\overline{}$	$\overline{\text{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$K^+(-$	Cl^- \sim	$\overline{}$
θ	1.000	1.000	1.000	1.000	1.000	1.000
15	0.925	0.984	0.876	0.771	0.826	0.981
30	0.877	0.969	0.765	0.594	0.701	0.988
45	0.791	0.950	0.648	0.448	0.540	0.980
60	0.710	0.908	0.523	0.324	0.422	0.941
120	0.292	0.811	0.131	0.072	0.115	0.906

Supplementary Table 7: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 10 g/L, solution pH of 3, under a current density of 2.5 mA cm−² .

Supplementary Table 8: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 30 g/L, solution pH of 7, under a current density of 2.5 mA cm⁻².

min	$_{\rm Li^{+}}$ $\overline{}$	$\overline{\text{Mg}^2}$ + $\overline{}$	$Na+$ $\overline{}$	$K^+(-)$	$Cl^{-}(-)$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.953	0.980	0.921	0.846	0.912	1.245
30	0.833	0.958	0.788	0.671	0.786	1.203
45	0.778	0.936	0.712	0.546	0.711	1.285
60	0.717	0.914	0.634	0.446	0.632	1.298
90	0.553	0.892	0.456	0.272	0.454	1.221
120	0.413	0.870	0.310	0.137	0.321	1.107

Supplementary Table 9: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 30 g/L, solution pH of 7, under a current density of 7.5 mA cm−² .

min	$\rm Li^+$ \overline{a}	$\overline{\text{Mg}^{2+}}$ -	$Na+$ $\overline{}$	$K^+(-)$	Cl^- $\overline{}$	$\overline{}$
0	1.000	1.000	1.000	1.000	1.000	1.000
15	0.889	0.979	0.816	0.730	0.862	1.000
30	0.764	0.960	0.649	0.516	0.749	1.004
45	0.747	0.968	0.501	0.326	0.637	1.007
60	0.555	0.952	0.334	0.191	0.518	1.012
90	0.171	0.927	0.067	0.055	0.268	1.009
120	0.024	0.583	0.001	0.018	0.051	0.982

Supplementary Table 10: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L, solution pH of 7, under a current density of 2.5 mA cm−² .

min)	$Li+$ $\qquad \qquad -$	$\overline{\text{Mg}^{2+}}$ -	$Na+$ $\overline{}$	$K^+(-)$	Cl^- $\overline{}$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
30	0.963	0.973	0.954	0.944	0.950	0.985
60	0.907	0.922	0.876	0.825	0.884	0.965
90	0.878	0.901	0.832	0.768	0.833	0.955
120	0.796	0.851	0.764	0.672	0.747	0.937
$150\,$	0.767	0.829	0.736	0.624	0.716	0.932

Supplementary Table 11: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L, solution pH of 5, under a current density of 2.5 mA cm⁻².

Supplementary Table 12: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L , solution pH of 3, under a current density of 2.5 mA cm⁻².

min)	$Li+$ $\overline{ }$	Mg^{2+} $\overline{}$	$Na+$ $\overline{}$	$K^+(-)$	$Cl^{-}(-)$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
30	0.951	0.948	0.923	0.914	0.913	0.977
60	0.917	0.907	0.860	0.834	0.862	0.939
90	0.876	0.879	0.813	0.768	0.830	0.916
120	0.850	0.866	0.794	0.716	0.784	0.914
150	0.832	0.858	0.747	0.690	0.768	0.917

Supplementary Table 13: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L , solution pH of 7, under a current density of 7.5 mA cm⁻².

min)	$Li+$ $\overline{ }$	$\overline{\text{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$K^+(-)$	Cl=(-)	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.928	0.989	0.931	0.902	0.935	0.992
30	0.898	0.982	0.876	0.823	0.894	1.000
60	0.866	0.984	0.811	0.753	0.838	1.005
90	0.816	0.963	0.751	0.678	0.777	0.980
120	0.615	${0.950}$	0.564	0.515	0.585	0.983

Supplementary Table 14: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L, solution pH of 5, under a current density of 7.5 mA cm−² .

min	$\rm Li^+$ $\qquad \qquad -$	$\overline{\text{Mg}^{2+}}$ -	$Na+$ $\overline{}$	$K^+(-)$	Cl^- $\overline{}$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.955	0.995	0.958	0.919	0.948	0.998
30	0.907	0.994	0.881	0.821	0.897	0.980
60	0.873	0.949	0.832	0.753	0.831	0.982
90	0.793	0.934	0.742	0.673	0.710	0.894
120	0.742	0.875	0.675	0.602	0.641	0.912

Supplementary Table 15: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L, solution pH of 3, under a current density of 7.5 mA cm⁻².

Supplementary Table 16: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L , solution pH of 7, under a current density of 15.0 mA cm⁻².

min)	$Li+$ $\overline{ }$	$\overline{\text{Mg}}^{2+}$ $\overline{}$	$Na+$ $\overline{}$	$K^+(-$	Cl ⁻ (-	<u>. —</u> $\overline{ }$
	$1.000\,$	1.000	1.000	1.000	1.000	1.000
15	0.917	0.982	0.862	0.805	0.871	0.985
30	0.855	0.969	0.752	0.647	0.788	1.000
45	0.788	0.973	0.621	0.505	0.676	1.010
60	0.686	0.929	0.502	0.356	0.554	0.960
120	0.271	0.903	0.129	0.031	0.171	0.967

Supplementary Table 17: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L , solution pH of 5, under a current density of 15.0 mA cm⁻².

min)	$Li+$ $\overline{}$	$\overline{\text{Mg}^{2+}}$ -	Na^{+} $\overline{}$	$K^+(-)$	Cl ⁻ (-	$\overline{}$
0	1.000	1.000	1.000	1.000	1.000	1.000
15	0.929	0.990	0.898	0.837	0.841	0.997
30	0.833	0.987	0.746	0.641	0.742	0.959
45	0.763	0.898	0.650	0.506	0.614	0.965
60	0.602	0.868	0.474	0.346	0.378	0.788
120	0.500	0.751	0.342	0.203	0.245	0.823

Supplementary Table 18: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 70 g/L , solution pH of 3, under a current density of 15.0 mA cm−² .

min	$\rm Li^+$ $\qquad \qquad -$	$\overline{\text{Mg}^{2+}}$ -	$Na+$ \overline{a}	$K^+(-)$	Cl^- . –	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.911	0.983	0.881	0.802	0.926	0.970
30	0.819	0.949	0.750	0.626	0.804	0.946
60	0.621	0.854	0.560	0.430	0.567	0.901
90	0.557	0.820	0.469	0.330	0.463	0.813
120	0.474	0.775	0.374	0.239	0.360	0.779

Supplementary Table 19: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 250 g/L, solution pH of 7, under a current density of 2.5 mA cm⁻².

Supplementary Table 20: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 250 g/L, solution pH of 7, under a current density of 30.0 mA cm⁻².

min)	$\rm Li^+$ $\overline{ }$ $\overline{ }$	$\overline{\text{Mg}}^{2+}$ -	$Na+$ $\overline{}$	$K^+(-$	Cl^- ($\overline{}$	$\overline{}$
	$1.000\,$	1.000	1.000	1.000	1.000	1.000
30	0.787	0.989	0.753	0.724	0.996	0.946
60	0.691	0.967	0.656	0.599	0.800	0.990
90	0.537	0.952	0.463	0.477	0.655	0.981
120	0.254	0.935	0.108	0.196	0.245	0.904
150	0.142	${0.919}$	0.020	0.011	0.135	0.957

Supplementary Table 21: Transient behavior of the normalized ion concentrations for Chilean brine at a total dissolved solid concentration of 250 g/L , solution pH of 3, under a current density of 30.0 mA cm⁻².

min)	$\rm Li^+$ $\overline{ }$	$\rm Mg^{2+}$ -	$Na+$ $\overline{}$	$K^+(-)$	Cl ⁻ (-)	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
30	0.856	0.999	0.825	0.794	0.999	1.176
60	0.729	0.949	0.710	0.659	0.797	1.250
90	0.572	0.899	0.506	0.519	0.650	1.179
120	0.275	0.849	0.120	0.211	0.246	0.974
150	0.152	0.799	0.022	0.012	0.134	1.011

Supplementary Table 22: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 10 g/L, solution pH of 7, under a current density of 2.5 mA cm−² .

min)	$\rm Li^+$ $\overline{}$	$\overline{\text{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$\rm K^+$ $\overline{}$	Cl^- $\overline{}$	$\overline{}$
θ	1.000	1.000	1.000	1.000	1.000	1.000
15	0.933	0.984	0.856	0.851	0.890	0.999
30	0.850	0.962	0.702	0.518	0.759	0.996
45	0.757	0.941	0.526	0.304	0.672	0.986
60	0.658	0.933	0.406	0.199	0.548	0.999
90	0.393	0.886	0.175	0.100	0.337	0.998
120	0.077	0.766	0.004	0.023	0.144	0.983

Supplementary Table 23: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 10 g/L, solution pH of 5, under a current density of 2.5 mA cm⁻².

Supplementary Table 24: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 10 g/L, solution pH of 3, under a current density of 2.5 mA cm⁻².

min)	$Li+$ $($ $-$	$\overline{\text{Mg}^2}$ + $\overline{}$	$Na+$ $\overline{}$	$K^+(-)$	$\rm Cl^-$ $ ($	$\overline{}$
θ	1.000	1.000	1.000	1.000	1.000	1.000
15	0.901	0.950	0.865	0.840	0.895	1.017
30	0.863	0.940	0.744	0.713	0.804	1.017
45	0.765	0.919	0.618	0.477	0.680	1.016
60	0.663	0.901	0.502	0.384	0.615	1.007
90	0.446	0.902	0.267	0.123	0.389	1.004
120	0.119	0.833	0.059	0.018	0.198	1.006

Supplementary Table 25: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 30 g/L, solution pH of 7, under a current density of 2.5 mA cm−² .

min)	$Li+$ $\overline{}$	$\overline{\mathrm{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$\rm K^+$ ($\qquad \qquad -$	Cl^- $\overline{}$	$\overline{}$
0	1.000	1.000	1.000	1.000	1.000	1.000
15	0.906	0.961	0.883	0.826	0.896	0.962
30	0.819	0.939	0.771	0.677	0.800	1.133
45	0.740	0.875	0.674	0.536	0.701	1.143
60	0.636	0.831	0.554	0.384	0.588	1.132
90	0.464	0.787	0.367	0.215	0.411	1.062
120	0.298	0.765	0.206	0.087	0.255	1.107

Supplementary Table 26: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 30 g/L , solution pH of 7, under a current density of 7.5 mA cm−² .

min)	$Li+$ $\overline{}$	$\overline{\text{Mg}}^{2+}$ $\overline{}$	$Na+$ $\overline{}$	$\rm K^+ ($ -`	Cl^- $\overline{}$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.934	0.993	0.860	0.796	0.882	1.008
30	0.848	0.961	0.701	0.590	0.755	1.006
45	0.768	0.964	0.580	0.447	0.640	1.025
60	0.659	0.936	0.427	0.275	0.498	1.012
120	0.064	0.815	0.040	0.055	0.056	0.959

Supplementary Table 27: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L, solution pH of 7, under a current density of 2.5 mA cm⁻².

Supplementary Table 28: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L, solution pH of 5, under a current density of 2.5 mA cm−² .

min)	$\rm Li^+$ $\overline{}$	$\overline{\text{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$\rm K^+$ Γ (–).	Cl^- $\overline{}$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.964	0.980	0.962	0.916	0.975	0.957
30	0.932	0.954	0.907	0.855	0.897	0.935
45	0.883	0.890	0.837	0.764	0.806	0.871
90	0.834	0.873	0.759	0.657	0.721	0.860
120	0.857	0.904	0.757	0.621	0.734	0.915
150	0.709	0.785	0.621	0.499	0.566	0.785

Supplementary Table 29: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L, solution pH of 3, under a current density of 2.5 mA cm⁻².

min)	$Li+$ $\overline{}$	Mg^{2+} $\overline{}$	$Na+$ $\overline{}$	$\rm K^+$ $\overline{}$	Cl^- $\overline{}$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
30	0.970	0.969	0.934	0.767	0.934	0.987
60	0.881	0.926	0.860	0.774	0.877	0.949
90	0.809	0.856	0.768	0.627	0.789	0.927
120	0.756	0.807	0.698	0.574	0.730	0.901
180	0.669	0.749	0.594	0.428	0.631	0.850

Supplementary Table 30: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L , solution pH of 7, under a current density of 7.5 mA cm⁻².

min)	$\rm Li^+$ $\overline{}$	$\overline{\text{Mg}}^{2+}$ (–	Na^{+} $\overline{}$	$K^+(-)$	$Cl^{-}(-)$	$\overline{}$
	1.000	$1.000\,$	1.000	1.000	1.000	1.000
15	0.918	0.966	0.916	0.831	0.975	0.992
30	0.752	0.918	0.735	0.549	0.802	0.971
60	0.687	0.913	0.605	0.433	0.692	0.976
90	0.627	0.911	0.486	0.295	0.624	0.937
120	0.389	0.851	0.254	0.128	0.414	0.929

Supplementary Table 31: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L , solution pH of 5, under a current density of 7.5 mA cm−² .

Supplementary Table 32: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L, solution pH of 3, under a current density of 7.5 mA cm⁻².

min)	$Li+$ $\qquad \qquad -$	$\overline{\text{Mg}^{2+}}$ -	$Na+$ $\overline{}$	$\rm K^+$ $\overline{}$	Cl^- $(-$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.945	0.990	0.929	0.868	0.938	0.931
30	0.903	0.946	0.866	0.747	0.878	0.911
60	0.842	0.904	0.803	0.697	0.821	0.876
90	0.767	0.838	0.684	0.571	0.738	0.848
120	0.672	0.755	0.594	0.456	0.633	0.818

Supplementary Table 33: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L , solution pH of 7, under a current density of 15.0 mA cm⁻².

min)	$Li+$ $\qquad \qquad -$	$\overline{\text{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$\rm K^+$ $\overline{ }$	Cl^- $\overline{}$	$\overline{ }$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.971	1.020	0.931	0.800	0.980	1.084
30	0.793	0.917	0.713	0.537	0.811	1.013
60	0.696	0.914	0.563	0.437	0.681	0.993
90	0.654	0.932	0.448	0.289	0.618	1.011
120	0.344	0.872	0.192	0.104	0.387	0.960
150	0.083	0.812	0.029	0.020	0.223	0.970

Supplementary Table 34: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L, solution pH of 5, under a current density of 15.0 mA cm⁻².

min)	$\rm Li^+$ $\qquad \qquad -$	$\overline{\text{Mg}^{2+}}$ $\overline{}$	$Na+$ $\overline{}$	$K^+(-)$	Cl^- $\overline{}$	<u>. —</u> $\overline{ }$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.911	0.979	0.881	0.802	0.926	0.970
30	0.819	0.944	0.750	0.626	0.804	0.946
60	0.621	0.850	0.560	0.430	0.567	0.901
90	0.557	0.816	0.469	0.330	0.463	0.813
120	0.474	0.772	0.374	0.239	0.360	0.779
150	0.389	0.722	0.278	0.171	0.271	0.726

Supplementary Table 35: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 70 g/L, solution pH of 3, under a current density of 15.0 mA cm⁻².

min)	$Li+$ $\overline{}$	Mg^{2+} $\overline{}$	$Na+$ $\overline{}$	$\rm K^+$ $\overline{}$	Cl^- $\overline{}$	$\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
15	0.912	0.994	0.898	0.855	0.914	0.783
30	0.849	0.927	0.811	0.731	0.831	1.160
60	0.809	0.886	0.756	0.634	0.775	0.849
90	0.732	0.823	0.615	0.525	0.696	0.879
120	0.604	0.713	0.509	0.360	0.553	1.208
150	0.464	0.599	0.358	0.193	0.432	1.154

Supplementary Table 36: Transient behavior of the normalized ion concentrations for Chinese brine at a total dissolved solid concentration of 250 g/L , solution pH of 7, under a current density of 2.5 mA cm⁻².

min)	Li ⁺ \overline{a}	Mg^{2+} $\overline{}$	Na^{+} $\overline{}$	$K^+(-$	$\rm Cl^-$ l ($\overline{}$
	1.000	1.000	1.000	1.000	1.000	1.000
30	0.991	0.993	0.985	0.964	0.900	0.947
60	0.981	0.987	0.881	0.928	1.107	1.134
90	0.972	0.980	0.960	0.892	0.995	1.193
120	0.963	0.973	0.901	0.856	1.048	1.160
150	0.953	0.967	0.878	0.806	1.068	1.136

Supplementary Table 37: Transient behavior of the normalized ion concentrations for dual cation $(Li^+ - Mg^{2+} - Cl^-)$, abbreviated as LMC) brine at a total dissolved solid concentration of 0.35 M, solution pH of 7, under a current density of 2.5 mA cm^{-2} .

min) t.	$Li+$ $\overline{ }$	$\overline{\text{Mg}}^{2+}$ $\overline{}$	$Cl^{-}(-)$	$\overline{}$
	1.000	1.000	1.000	0.00
15	0.755	0.985	0.900	0.00
30	0.513	0.947	0.819	0.00
45	0.283	0.931	0.751	0.00
60	0.106	0.899	0.674	0.00
90	0.015	0.754	0.542	0.00
120	0.009	0.598	0.420	0.00

Supplementary Table 38: Transient behavior of the normalized ion concentrations for dual cation

υ Ι		ш -1	1018 -	◡ェ $\overline{}$	\cup ^{\cup} Δ \mathbf{I}
	0	1.000	1.000	1.000	1.000
	15	0.744	0.983	0.905	1.009
	30	0.451	0.976	0.812	0.987
	45	0.257	0.949	0.752	0.992
	60	0.082	0.909	0.685	0.991
	90	0.019	0.770	0.536	0.997
	120	0.014	0.627	0.418	0.986

Supplementary Table 39: Transient behavior of the normalized ion concentrations for dual cation (Li⁺ - Mg²⁺ - Cl[−] - SO₄⁻, abbreviated as LMCS) brine at a total dissolved solid concentration of 0.35 M, solution pH of 3, under a current density of 2.5 mA cm^{-2} .

Supplementary Table 40: Compiled ion transport numbers for the Chilean and Chinese salt-lake brine experiments. Supplementary Table 40: Compiled ion transport numbers for the Chilean and Chinese salt-lake brine experiments.

References

- (1) Ahdab, Y. D.; Rehman, D.; Lienhard, J. H. Brackish water desalination for greenhouses: Im- proving groundwater quality for irrigation using monovalent selective electrodialysis reversal. Journal of Membrane Science 2020, 610, 118072.
- (2) Fan, H.; Yip, N. Y. Elucidating conductivity-permselectivity tradeoffs in electrodialysis and reverse electrodialysis by structure-property analysis of ion-exchange membranes. Journal of Membrane Science 2019, 573, 668–681.
- (3) Yaroshchuk, A.; Bruening, M. L.; Zholkovskiy, E. Modelling nanofiltration of electrolyte solu-tions. Advances in Colloid and Interface Science 2019, 268, 39–63.
- (4) Foo, Z. H.; Rehman, D.; Coombs, O. Z.; Deshmukh, A.; Lienhard, J. H. Multicomponent Fick-⁵³² ian solution-diffusion model for osmotic transport through membranes. *Journal of Membrane* Science 2021, 640, 119819.
- (5) Rehman, D.; Ahdab, Y. D.; Lienhard, J. H. Monovalent selective electrodialysis: Modelling multi-ionic transport across selective membranes. Water Research 2021, 199, 117171.
- (6) Luo, T.; Abdu, S.; Wessling, M. Selectivity of ion exchange membranes: A review. Journal of Membrane Science 2018, 555, 429–454.
- (7) Kingsbury, R. S.; Coronell, O. Modeling and validation of concentration dependence of ion exchange membrane permselectivity: Significance of convection and Manning's counter-ion condensation theory. Journal of Membrane Science 2021, 620, 118411.
- (8) Kingsbury, R. S.; Zhu, S.; Flotron, S.; Coronell, O. Microstructure Determines Water and Salt Permeation in Commercial Ion-Exchange Membranes. ACS Applied Materials & Interfaces 2018, 10, 39745–39756.
- (9) McGovern, R. K.; Weiner, A. M.; Sun, L.; Chambers, C. G.; Zubair, S. M.; Lienhard, J. H. ⁵⁴⁵ On the cost of electrodialysis for the desalination of high salinity feeds. Applied Energy 2014, 136, 649–661.

(10) Strathmann, H. Ion-exchange membrane separation processes; Elsevier, 2004.

- (11) Wenten, I. G.; Khoiruddin, K.; Alkhadra, M. A.; Tian, H.; Bazant, M. Z. Novel ionic separa-₅₄₉ tion mechanisms in electrically driven membrane processes. Advances in Colloid and Interface Science 2020, 284, 102269.
- (12) Kamcev, J.; Galizia, M.; Benedetti, F. M.; Jang, E.-S.; Paul, D. R.; Freeman, B. D.; Man- ning, G. S. Partitioning of mobile ions between ion exchange polymers and aqueous salt solu- tions: importance of counter-ion condensation †. Physical Chemistry Chemical Physics 2016, 18, 6021.
- (13) Fan, H.; Huang, Y.; Billinge, I. H.; Bannon, S. M.; Geise, G. M.; Yip, N. Y. Counterion Mobility in Ion-Exchange Membranes: Spatial Effect and Valency-Dependent Electrostatic 557 Interaction. ACS ES&T Engineering 2022, 2, 1274-1286.
- (14) Kamcev, J.; Paul, D. R.; Freeman, B. D. Ion activity coefficients in ion exchange polymers: Applicability of Manning's counterion condensation theory. Macromolecules 2015, 48, 8011– 8024.
- (15) Kim, H. T.; Frederick, W. J. Evaluation of Pitzer Ion Interaction Parameters of Aqueous 562 Electrolytes at 25℃. 1. Single Salt Parameters. Journal of Chemical and Engineering Data 1988, 33, 177–184.
- (16) Kim, H. T.; Frederick, W. J. Evaluation of pitzer ion interaction parameters of aqueous mixed electrolyte solutions at 25 °C. 2. Ternary mixing parameters. Journal of Chemical and Engi-neering Data 1988, 33, 278–283.
- (17) Kingsbury, R. S.; Wang, J.; Coronell, O. Comparison of water and salt transport properties of ion exchange, reverse osmosis, and nanofiltration membranes for desalination and energy applications. Journal of Membrane Science 2020, 604, 117998.
- (18) Ying, J.; Lin, Y.; Zhang, Y.; Jin, Y.; Li, X.; She, Q.; Matsuyama, H.; Yu, J. Mechanistic insights into the degradation of monovalent selective ion exchange membrane towards long-₅₇₂ term application of real salt lake brines. *Journal of Membrane Science* 2022, 652, 120446.
- (19) Virtanen, P. et al. SciPy 1.0: fundamental algorithms for scientific computing in Python. Nature Methods 2020, 17, 261–272.
- (20) Liu, H.; She, Q. Influence of membrane structure-dependent water transport on conductivity- permselectivity trade-off and salt/water selectivity in electrodialysis: Implications for osmotic electrodialysis using porous ion exchange membranes. Journal of Membrane Science 2022, 650, 120398.
- (21) Martí-Calatayud, M. C.; Buzzi, D. C.; García-Gabaldón, M.; Bernardes, A. M.; Tenório, J. A.; Pérez-Herranz, V. Ion transport through homogeneous and heterogeneous ion-exchange mem- branes in single salt and multicomponent electrolyte solutions. Journal of Membrane Science $\frac{582}{2014}$, $\frac{466}{45}$ -57.
- (22) Długołecki, P.; Anet, B.; Metz, S. J.; Nijmeijer, K.; Wessling, M. Transport limitations in ion exchange membranes at low salt concentrations. Journal of Membrane Science 2010, 346, 163–171.
- (23) Alkhadra, M. A.; Conforti, K. M.; Gao, T.; Tian, H.; Bazant, M. Z. Continuous Separation of Radionuclides from Contaminated Water by Shock Electrodialysis. Environmental Science & $Technology$ 2019, 54 , 527–536.
- (24) Abu-Rjal, R.; Chinaryan, V.; Bazant, M. Z.; Rubinstein, I.; Zaltzman, B. Effect of concentra-tion polarization on permselectivity. Physical Review E 2014, 89, 012302.
- (25) Aydogan Gokturk, P.; Sujanani, R.; Qian, J.; Wang, Y.; Katz, L. E.; Freeman, B. D.; Crum- $\lim_{z \to z_2}$ lin, E. J. The Donnan potential revealed. *Nature Communications* 2022, 13, 1–7.
- (26) Ahdab, Y. D.; Rehman, D.; Schücking, G.; Barbosa, M.; Lienhard, J. H. Treating Irriga- tion Water Using High-Performance Membranes for Monovalent Selective Electrodialysis. ACS $ES\&T$ *Water* **2021**, 1, 117-124.
- (27) Tedesco, M.; Hamelers, H. V.; Biesheuvel, P. M. Nernst-Planck transport theory for (reverse) electrodialysis: I. Effect of co-ion transport through the membranes. Journal of Membrane Science 2016, 510, 370–381.
- (28) Ying, J.; Lin, Y.; Zhang, Y.; Yu, J. Developmental Progress of Electrodialysis Technologies 600 and Membrane Materials for Extraction of Lithium from Salt Lake Brines. ACS ES&T Water $\frac{601}{2023}$, 3, 1720–1739.
- (29) Ding, D.; Yaroshchuk, A.; Bruening, M. L. Electrodialysis through nafion membranes coated with polyelectrolyte multilayers yields >99% pure monovalent ions at high recoveries. Journal of Membrane Science 2022, 647, 120294.
- (30) Foo, Z. H.; Rehman, D.; Bouma, A. T.; Monsalvo, S.; Lienhard, J. H. Lithium Concentration from Salt-Lake Brine by Donnan-Enhanced Nanofiltration. Environmental Science & Technol- $\log y$ 2023, 57, 6320–6330.
- (31) Vera, M. L.; Torres, W. R.; Galli, C. I.; Chagnes, A.; Flexer, V. Environmental impact of direct $\frac{609}{1000}$ lithium extraction from brines. Nature Reviews Earth & Environment 2023, 4, 149–165.
- (32) Xu, P.; Hong, J.; Qian, X.; Xu, Z.; Xia, H.; Tao, X.; Xu, Z.; Ni, Q. Q. Materials for lithium recovery from salt lake brine. Journal of Materials Science 2021, 56, 16–63.
- (33) Zhang, Y.; Hu, Y.; Wang, L.; Sun, W. Systematic review of lithium extraction from salt-lake brines via precipitation approaches. Minerals Engineering 2019, 139 .
- (34) Alkhadra, M. A.; Su, X.; Suss, M. E.; Tian, H.; Guyes, E. N.; Shocron, A. N.; Conforti, K. M.; Pedro De Souza, J.; Kim, N.; Tedesco, M.; Khoiruddin, K.; Wenten, G.; Santiago, J. G.; Hatton, T. A.; Bazant, M. Z. Electrochemical Methods for Water Purification, Ion Separations, and Energy Conversion. Chemical Reviews 2022, 122, 13547–13635.
- (35) Palmer-Wilson, K.; Donald, J.; Robertson, B.; Lyseng, B.; Keller, V.; Fowler, M.; Wade, C.; Scholtysik, S.; Wild, P.; Rowe, A. Impact of land requirements on electricity system decarbon- μ ⁵²⁰ isation pathways. Energy Policy 2019, 129, 193–205.
- (36) Mistry, K. H.; Lienhard, J. H. Generalized least energy of separation for desalination and other chemical separation processes. Entropy 2013, 15, 2046–2080.